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1 **Hydrological, physicochemical and metabolic signatures in groundwater and**
2 **snowmelt streams in the Japanese Alps**

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9 **Highlights**

- 10 • Morphological, hydrological and physicochemical conditions as well as metabolic
11 activity are assessed across streams
- 12 • Comparisons are made between snowmelt and groundwater fed systems
- 13 • Snowmelt fed streams experience no flow and storm events
- 14 • Groundwater streams had higher environmental stability and higher metabolic
15 activity
- 16 • Patterns were similar to other regions of the globe, but there was higher inter-
17 stream variation

18 **Abstract**

19 The unique hydrology and physicochemistry of alpine streams provide an important
20 influence on the structure and function of inhabiting biological communities. A
21 substantial body of research exists on alpine streams across many regions of the
22 globe (e.g. Europe, North and South America and Greenland). To date, however, there
23 have been few studies investigating the environmental conditions present in alpine
24 streams across the Japanese archipelago. The lack of information on alpine streams
25 in Japan is problematic as unique regional climates, e.g. some of the highest levels of
26 snowfall globally, are likely to have repercussions for morphological, hydrological,
27 physicochemical and metabolic signatures, causing them to differ from those observed
28 in other regions. In this study we compare the morphology, hydrology,
29 physicochemistry and metabolic activity of snowmelt and groundwater fed streams in
30 the Kamikochi region of the Japanese Alps. Stream discharge, water chemistry (major
31 ions, silica, dissolved oxygen), water temperature and channel stability were
32 measured over a period of 16 months in 2017–2018. Metabolic activity was
33 determined using Resazurin-Resorufin (Raz-Rru) Smart Tracer and variation in the
34 Raz transformation rate was assessed to understand the effects of hydrology and
35 physicochemistry on ecosystem functioning. Snowmelt streams were characterised by
36 higher variability of water temperature, water chemistry and stream discharge, both,
37 within and between sites. Indeed, two of the snowmelt streams experienced no flow
38 conditions for several periods and also floods. In comparison, water chemistry, water
39 temperature and stream discharge in groundwater fed streams were more temporally
40 stable. Metabolic activity was higher in one groundwater fed stream, attributed to
41 significant growth of macrophytes. These findings indicate that the patterns of
42 morphology, hydrology, physicochemistry and metabolic activity across streams in the

43 Japanese Alps largely resemble those identified elsewhere, although there were
44 higher levels of inter-stream variation. The diversity and inter-site variation of
45 hydrological and physicochemical conditions are likely responsible for the unique flora
46 and fauna in the streams. This study therefore indicates the potential importance of
47 habitat templates for the aquatic biodiversity hotspot in this region.

48 **Keywords:** alpine streams, water temperature, habitat templates, metabolic activity

49 **1. Introduction**

50 Alpine stream systems are under threat from climate change (Brown *et al.* 2007a;
51 Khamis *et al.* 2014), as well as other anthropogenic stressors such as nutrient
52 enrichment (Hotaling *et al.* 2017). Particular conservation concern is linked to alpine
53 streams as they support high biodiversity and a large number of endemic and rare
54 species that are absent from other aquatic ecosystems across the globe (Robinson *et*
55 *al.* 2010). The combination of high biodiversity and endemism alongside rapid
56 environmental change (Prowse *et al.* 2006) poses a disproportionate risk of species
57 extinctions and biodiversity loss (Dirnböck *et al.* 2011).

58 The unique flora and fauna inhabiting alpine systems is in part generated by the high
59 heterogeneity of environmental conditions both within and between alpine stream
60 systems (Ward 1994; Brown *et al.* 2007a; Finn & Poff 2011; Jacobsen *et al.* 2012;
61 Hotaling *et al.* 2017). In particular, streams fed by different water sources support
62 markedly different biological communities resulting in high beta- and gamma-diversity
63 across these streams (Milner *et al.* 2010). Each stream type is characterised by the
64 dominant contributing water source and has a unique combination (or signature) of
65 flow regime and water temperature, turbidity and chemistry (often characterised by
66 redox indicators, major ions and dissolved nutrients) (Brown *et al.* 2003, 2009; Hannah
67 *et al.* 2007). Environmental conditions in space and through time are more constant in
68 groundwater fed streams whereas intermediate variability in conditions occur in
69 streams which respond to seasonal snowmelt dynamics, but the highest levels of
70 variability are typically found in glacial meltwater fed streams (Brown *et al.* 2007b;
71 Milner *et al.* 2010; Khamis *et al.* 2016). Intermediate habitat templates can be
72 generated by mixtures of water sources (Brown *et al.* 2009; Milner *et al.* 2010). Only
73 specialised taxa are able to persist in the harsh conditions of unstable channels fed

74 entirely by snowmelt (Snook & Milner 2001; Cauvy-Fraunié *et al.* 2014) in comparison
75 to the more species rich assemblages in groundwater dominated streams (Brown *et*
76 *al.* 2006b, 2007b; Khamis *et al.* 2016).

77 Our understanding of alpine stream habitats is relatively well developed across many
78 regions of the globe, including; North America (Finn *et al.* 2006; Windsor *et al.* 2017),
79 South America (Jacobsen *et al.* 2014; Cauvy-Fraunié *et al.* 2015), Africa (Musonge *et*
80 *al.* 2020), Greenland (Docherty *et al.* 2018), Svalbard (Blaen *et al.* 2014) and Europe
81 (Finn *et al.* 2013; Khamis *et al.* 2016). Research in the Japanese archipelago is
82 however more limited, with existing studies focusing on the population dynamics and
83 genetics of individual invertebrate, fish and amphibian species, many of them endemic
84 (Fukumoto *et al.* 2015; Tojo *et al.* 2017b, a; Saito *et al.* 2018). As a result, the habitat
85 template, the suite of environmental conditions influencing biological communities
86 (i.e., morphology, hydrology and physicochemistry), is not as well understood for
87 Japanese alpine streams. Previous studies indicate streams in the Japanese Alps are
88 fed by several water sources common in other alpine systems, including snowmelt,
89 groundwater and to a very limited extent glacial runoff (Yoshimura *et al.* 2005; Milner
90 *et al.* 2020). It is unclear, however, how source water contributions affect the
91 physicochemical and hydrological conditions in these streams. Furthermore, it is not
92 known whether the patterns in habitat templates within and between Japanese alpine
93 streams mirror those observed across other alpine regions of the globe.

94 Many of the alpine regions studied previously have similar hydro-climatological
95 conditions and are positioned within similar climate envelopes. Climatic conditions in
96 the Japanese Alps, however, are relatively unique. As an example catchments in this
97 region receive some of the highest levels of snowfall in the world (Ueda 2014) and
98 potentially these streams may display a unique set of morphological, hydrological,

99 physicochemical and metabolic signatures, supporting novel habitat templates that are
100 different from other regions of the globe.

101 Here we assessed the morphological, hydrological, physicochemical and metabolic
102 signatures across six streams fed by two different water sources (snowmelt and
103 groundwater) across the Kamikochi region of the Japanese Alps. Water temperature,
104 discharge and water chemistry were used to characterise the systems, following the
105 important variables identified by the classification system produced by Brown *et al.*
106 (2009). Metabolic activity was measured in streams to understand how environmental
107 conditions may impact ecological functioning. As the environmental conditions and
108 ecological functioning in these systems are relatively poorly described we proposed
109 three exploratory hypotheses based on studies in other regions of the globe:

110 **H₁.** Water source will significantly influence morphological, hydrological and
111 physicochemical signatures between the study streams

112 **H₂.** Metabolic activity will vary across the stream sites and be significantly higher in
113 the groundwater systems

114 **H₃.** Hydrological, physicochemical and metabolic signatures in groundwater streams
115 will be less variable when compared to snowmelt streams

116 **2. Material and methods**

117 **2.1. Study sites**

118 The study was completed in the Kamikochi valley within the Chūbu-Sangaku National
119 Park, Nagano Prefecture, Japan (1,743 km²), a region known as the Hida Mountains
120 or northern Japanese Alps. The National Park is characterised by high mountains with
121 Mt. Tateyama (2,455 m) and Mt. Tsurugi-dake (2,926 m) at the northern reach and Mt.
122 Hotaka-dake (3,190 m) as well as Mt. Norikura (3,026 m) at the southern edge.

123 Kamikochi valley is approximately 18 km in length, with the average valley floor
124 elevation of 1,500 m. Up to about 12,000 years ago the River Azusa primarily flowed
125 into the Jinzu-gawa River system on the Hida-Takayama southern side of Northern
126 Japan Alps but then, potentially due to a volcanic collapse of Mt. Yake-dake (2,455 m),
127 the old Azusa River became naturally dammed and started to flow north towards
128 Matsumoto (Harayama 2015). This river catchment supports a significant biodiversity
129 hotspot in the Japanese archipelago (Tojo *et al.*, 2017a, b).

130 Six streams were selected in the study area (Figure 1), all of which are tributaries of
131 the Asuza River. Three were fed predominantly by groundwater (Shimizugawa,
132 Minamisawa and Bentenzawa) and three fed predominantly by snowmelt (Dakesawa,
133 Shirasawa and Tokusawa). The dominant water source of the streams was identified
134 through local knowledge and initial field surveys by identifying groundwater springs
135 and snowpacks. The sources of snowmelt fed stream were >2000 m elevation and
136 considerably longer (3-5 km) in comparison to groundwater fed stream sources which
137 were at ~1500 m.a.s.l elevation and relatively short (< 1 km) (see Figure 1). Mean
138 widths varied from 4 to 10 m with mean depths <30 cm. At each site a stream reach
139 of 10 m was designated as the sample site and reaches were above any bridges to
140 minimise potential anthropogenic influence on streams. Sites were typically within 200
141 m of the mainstem Azusa River.

142 **2.2. Meteorological data**

143 Meteorological data were obtained from the Shinshu University field station in the
144 Myojin area of Kamikochi (1,527 m.a.s.l; Figure 1). Data were collected at 10-minute
145 intervals using a HD9817T1 temperature sensor (Delta OHM, Padua, Italy), a 34-HT-
146 BP precipitation sensor (Ota Keiki, Tokyo, Japan) and manual snow depth
147 measurements.

148 **2.3. Morphological characterisation of the streams**

149 The morphology of the stream reaches was evaluated at the start of the sampling
150 period (July 2017). Mean width and depth of the stream channels at the sample
151 reaches were calculated from replicate measurements (n = 5) along a single transect
152 in the centre of the sampled reach. Although width and depth varied seasonally,
153 measurements taken during this initial period represent baseflow conditions
154 associated with each water source, as demonstrated in the discharge data. The relative
155 channel stability was also assessed using the bottom component of the Pfankuch
156 channel stability index (Pfankuch 1978). This method uses visual assessment of six
157 attributes (substrate brightness, angularity, consolidation, proportion of stable
158 materials, scouring and amount of clinging vegetation) and assigns them a value on a
159 1 to 4 scale. The scores are weighted according to their importance, and the sum of
160 the scores is the stability index. Lower values indicate higher channel stability.

161 **2.4. Water temperature**

162 Water temperature was recorded at 30-minute intervals over the sampling period (1st
163 August 2017 to 1st December 2018) using TinyTag dataloggers (TGP-4017; Gemini
164 Data Loggers Ltd., Chichester, UK) deployed 0.1 m above the streambed at the six
165 study sites. The number of recordings varied between sites, however, on average
166 >18000 measurements were made at each site over the sampling period. The
167 minimum number of measurements took place at Shirasawa, with only 5420 records
168 of temperature data due to the replacement logger being lost (after ~4 months). At
169 Bentenzawa groundwater stream the separate temperature logger was exposed to air
170 in December 2017 to April 2018 as the channel width constricted. For the following
171 analyses we have removed this section from the water temperature record. The
172 duration of all measurements was accounted for in further analyses.

173 **2.5. Stream discharge measurement**

174 Discharge ($\text{m}^3 \text{s}^{-1}$) was calculated using stage-discharge curves (Herschy 1993,
175 2009). We collected water depth at 30-min resolution from 1st August 2017 to 1st
176 December 2018 using several types of logger according to their availability, including:
177 Levellogger Edge 3001 (Solinst, Georgetown, Canada), dipperLog (Heron Instruments
178 Inc., Dundas, Canada) and Level TROLL 500 (In-Situ Inc., Fort Collins, USA). Water
179 depth measurements were not available for Shirasawa due to the loss of a sensor in
180 a large storm event in spring 2018. To generate stage-discharge curves we used a
181 conservative tracer approach. Briefly, we used standard cross-section approaches
182 (Herschy 2009) with a Kenek EU20 (Toyko, Japan) flow-meter in conjunction with a
183 conservative NaCl tracer dilution method (Tazioli 2011). The NaCl trace was
184 measured using a hand-held multi-parameter meter (YSI 6050000, USA). These
185 calibration measurements were completed under different flow conditions (0–
186 $1.98 \text{ m}^3 \text{ s}^{-1}$) across different seasons ($n = 5$, July–December 2017), to enable
187 discharge calculations at different times of year. Using the rating curves, we converted
188 water depth measurements into a continuous record of stream discharge across
189 sample sites.

190 **2.6. Water chemistry**

191 Triplicate water samples (125 ml) were collected at the six stream sites during the
192 2017 sampling season. Total dissolved solids (TDS) and dissolved oxygen (DO)
193 concentrations were measured at sites (mg L^{-1}) during the sample period (see Milner
194 et al. 2020 for sampling details). These samples were frozen at $-20 \text{ }^\circ\text{C}$ within 10 hours
195 of collection. Water samples were filtered (Whatman GF/F $0.7 \mu\text{m}$, GE Healthcare,
196 UK) and then analysed for the following ions; Ca^{2+} , K^+ , Mg^{2+} , Na^+ , NH_4^+ , SiO_2 , Cl^- , NO_2^-
197 , NO_3^- , SO_4^{2+} using a Hitachi U-2000 analyser and standard analytical methods for

198 these ions (APHA 2005). The concentration (mg L⁻¹) of each ion was calculated from
199 a calibration curve generated using a calibration standards.

200 **2.7. Metabolic activity**

201 The metabolic activity of five of the streams (Shimizugawa, Bentenzawa, Dakesawa,
202 Shirasawa and Tokusawa) was estimated using Resazurin-Resorufin (Raz-Rru) Smart
203 Tracing (RRST) (Haggerty *et al.* 2008, 2009; Argerich *et al.* 2011; González-Pinzón *et*
204 *al.* 2012). Unfortunately, a determination was not possible for Minamisawa, as it was
205 not possible to obtain sufficient distance between two sampling sites due to the size
206 of the stream. Across each of the five streams we conducted a single instantaneous
207 Resazurin injection (15–17th April 2018). At two sites, approximately 100 to 150 m
208 apart, 12–16 grab samples were collected to capture the tracer breakthrough curves.
209 The samples were cooled until analysis within 48 hours. Prior to fluorometric analysis
210 with a GGUN FL30 fluorescent spectrometer (Albillia Sarl, Switzerland) in bench-top
211 mode, all samples were allowed to reach room temperature and were buffered to a pH
212 of ~8.5. The buffer was made by mixing equal volumes of 1M NaH₂PO₄·H₂O and 1M
213 NaOH (Haggerty *et al.*, 2008) and a sample to buffer ratio of 100:1 was used (Blaen
214 *et al.* 2017). The transformation rate (λ_{Raz} , [s⁻¹]) of Resazurin (Raz) to Resorufin (Rru),
215 used as a proxy for metabolic activity, was then calculated according to the methods
216 presented in Argerich *et al.* (2011) and Blaen *et al.* (2017):

$$217 \quad \lambda_{Raz} = \frac{1}{\tau} \ln \left(\frac{m_{up}^{Raz}}{m_{up}^{Raz} + m_{up}^{Rru} - m_{down}^{Rru} \times Q_{ratio}} \right)$$

$$218 \quad Q_{ratio} = \frac{m_{up}^{Raz} + m_{up}^{Rru}}{m_{down}^{Raz} + m_{down}^{Rru}}$$

219 where m refers to the zeroth temporal moment (the integral of the concentration with
220 respect to time) of Raz and Rru at the upstream (up) and downstream (down) site. The

221 mean travel time between the two sites (τ) was calculated as the difference between
222 the first temporal moments at the downstream and upstream sites (Schmadel *et al.*
223 2016). The ratio of the discharges at the upstream and downstream site was included
224 to compensate for the dilution effect. More detail on the biogeochemistry of the Raz-
225 Rru reaction, and further specifics of this reactive tracer approach are provided in
226 González-Pinzón *et al.* (2012) and Knapp *et al.* (2018).

227 **2.8. Statistical analysis**

228 All data exploration and analyses were conducted in R (R Core Team 2021) and data
229 used in this study are available at DOI:10.5281/zenodo.4696775. Data were initially
230 investigated for normality, heteroscedasticity and outliers (Zuur *et al.* 2010).

231 Temperature duration curves were constructed for water temperature in each of the
232 streams (cf. Hannah *et al.* 2009; Khamis *et al.* 2015). This method provides a graphical
233 representation of the percentage of time that a specific water temperature is equalled
234 or exceeded. To generate these duration curves, we used the 'HydroTSM' package
235 (Zambrano-Bigiarini 2020), developed for calculating and visualising flow duration
236 curves. The steeper the curve the more variable the water temperature, with flat lines
237 indicating high thermal stability in the stream. We also calculated the coefficient of
238 variation (CV = standard deviation/mean) to understand temperature variability. Water
239 temperature was also used as an indicator of stream conditions, including free-flow,
240 surface freezing, flow cessation and snow cover (Brown *et al.* 2006a).

241 Stream discharge data were used to identify high and low flow events. Storm events
242 were classified as periods when the total stream discharge exceeded the mean annual
243 baseflow by 20% for >24 hours using the 'hydromad' package (Andrews *et al.* 2011).
244 We determined baseflow using one variable recursive digital filtering (Eckhardt 2005)

245 with a constant of 0.96 using the 'FlowScreen' package (Dierauer & Whitfield 2019).
246 For each identified storm event (N = 18), all of which were detected in a single
247 snowmelt fed stream (Tokusawa), we calculated maximum discharge ($\text{m}^3 \text{s}^{-1}$), mean
248 water temperature ($^{\circ}\text{C}$) and storm duration (hours). These metrics of the storm events
249 (data were first standardised to one standard deviation and mean centred) were
250 analysed using Principal Components Analysis (PCA) (Abdi & Williams 2010). We use
251 the position of storm events on the first two principal components of the PCA to classify
252 storms into different types (i.e., long duration summer storms). We calculated CV for
253 continuous data to understand variability in stream discharge over the sampling
254 period. Finally, flow cessation was indicated by complete flow cessation, i.e., zero
255 water depth, and supported by water temperature data (see method above).

256 Water chemistry (dissolved oxygen, total dissolved solids and major ion
257 concentrations) and metabolic activity data were investigated using a series of
258 Generalised Linear Models (GLMs) (Nelder & Baker 2006) and Generalised Linear
259 Mixed Models (GLMMs) (Bolker *et al.* 2009), using site as a random effect to control
260 for autocorrelation as a result of the sampling regime. Model structure, families and
261 link functions depended on the variable of interest and are reported in the results (also
262 see Table S1). All model assumptions were validated following Zuur *et al.* (2007) and
263 Thomas *et al.* (2015), by assessing the residual normality using QQ plots,
264 homogeneity of variance determined by plotting the residuals against fitted values and
265 influential observations using Cook's leverage distances.

266 Major ion data were also converted to milliequivalents per litre (meq L^{-1}) using the
267 'smwrBase' package (Lorenz 2015). Data for individual ions were then converted to
268 percentage of total major ions (%) using the 'hydrogeo' package (English 2017) for
269 Piper diagrams (Piper 1944).

270 **3. Results**

271 **3.1. Stream morphology**

272 Streams across the Kamikochi region were highly variable in their physical and
273 chemical characteristics, in relation to both water source contributions and other site-
274 specific factors (Table 1). Physicochemical characteristics were variable across
275 streams but were related to the dominant water source (groundwater or snowmelt).
276 Channel width was not significantly different between groundwater and snowmelt fed
277 streams (Square root transformed Gaussian GLM: $R^2 = 0.95$, $t_{1,4} = 105.5$, $p = 0.08$).
278 There were, however, significant differences in water depth (Gaussian GLM: $R^2 = 0.80$,
279 $t_{1,4} = 20.9$, $p = 0.01$) and the bottom component of the Pfankuch index (Gaussian
280 GLM: $R^2 = 0.65$, $t_{1,4} = 3.21$, $p = 0.033$) between snowmelt and groundwater streams.
281 In general, groundwater streams were characterised by higher channel stability (lower
282 Pfankuch index values) and depth in comparison to snowmelt fed streams (Table 1).

283 **3.2. Water temperature**

284 Variation in water temperature was high across four of the streams, with large diurnal
285 and seasonal fluctuations (Figure 2). Two groundwater fed streams (Shimizugawa and
286 Minamisawa), however, had low temporal variation in water temperature variation over
287 the 16-month study period (5.38 ± 0.43 SD, and 5.29 ± 0.46 SD, respectively). The
288 other groundwater (Bentenzawa - 3.55 ± 4.59 SD) and snowmelt streams (Dakesawa,
289 Shirasawa and Tokusawa) exhibited markedly higher levels of temperature variation
290 (6.06 ± 3.38 SD, 8.16 ± 5.47 SD and 5.24 ± 5.98 SD, respectively), and water
291 temperatures ranged from below zero up to nearly 20 °C (Figure 2). Several of the
292 loggers recorded extremely low temperatures, i.e., significantly below freezing (Figure
293 2a).

294 Duration curves for water temperature in groundwater streams had extremely shallow
295 slopes, indicating low temporal variability (Figure 2b). Conversely, snowmelt fed
296 streams had characteristic duration curves with steep slopes, indicating greater
297 variation and lower/higher stream temperature. Individual snowmelt streams,
298 however, maintained relatively unique duration curves and exhibited different levels of
299 temporal variability in water temperature. In particular, Dakesawa had a shallower
300 curve with a slope between that of the groundwater streams and other snowmelt fed
301 streams (Figure 2b).

302 Extreme low water temperature recorded on a number of sensors also hinted at
303 periods of dewatering in several streams. Of particular note were two prolonged winter
304 and spring no flow periods in both Shirasawa and Tokusawa snowmelt fed streams.
305 Tokusawa exhibited low temperatures (mean = -3.34 °C, min = -8.74 °C, max =
306 0.55 °C) over approximately 3 months, with high variation (CV = 1.82). In comparison,
307 Shirasawa had lower variation (CV = 0.32), however, only a short period of
308 temperatures below 0 °C was recorded before the logger finished recording
309 (December 2017).

310 **3.3. Stream discharge**

311 Variation in stream discharge was significant within and across streams observed in
312 the Kamikochi region (Figure 3). Snowmelt streams exhibited higher variability in
313 stream discharge (CV = 0.005–0.02) in comparison to groundwater fed streams (CV
314 = 0.001–0.006). There was, however, substantial variation in the stream discharge
315 between the individual streams and overlap between the two groups of streams
316 (groundwater and snowmelt), indicating site-specific discharge characteristics. Again,
317 Dakesawa appeared less variable than the other snowmelt streams, and Bentenzawa
318 was more variable than the other groundwater streams.

319 The snowmelt fed stream, Tokusawa, exhibited the highest levels of discharge
320 variation with a total of 18 storm events identified (Figure 3). The mean stream
321 discharge during storm events was $0.12 \pm 0.001 \text{ m}^3 \text{ s}^{-1}$ and on average the duration
322 of these events lasted for 132 hours (26–427 hours). These storm events occurred
323 relatively consistently throughout both 2017 and 2018, with no association to specific
324 months or seasons (Figure 4a). Many of the storms occurred shortly after peaks in
325 daily precipitation (Figure 4b). Several storm events, however, occurred when there
326 were low levels of precipitation, especially during spring 2017, but there were
327 substantial snowpack depths and higher spring air temperature (Figure 4b).
328 Furthermore, some discharge peaks occur after precipitation events on high snowpack
329 depths. Storms with different characteristics occurred at different times of the year,
330 and storm events clustered into two groups on the principal components (Figure 4c).
331 These groups were: (1) warm temperature, long duration storm flows that occurred in
332 summer and autumn; and (2) cold temperature, short duration storm flow events that
333 occurred in winter.

334 **3.4. Water chemistry**

335 Anion and cation concentrations were highly variable within and across streams (Table
336 2). Several ions, including NO_2^- , NO_3^- and SiO_2 , were below limits of detection at sites
337 at different sampling intervals, and NH_4^+ was only detected in 3 of the 83 samples
338 during the sample period (Table S2). Marked variation was found in the commonly
339 detected major ions, with similar patterns across the sites (Figure 5).

340 Concentrations of all major ions at the sites were markedly different, although several
341 sites were more similar to one another, in comparison to others (Figure 5). Once
342 variation between sites was controlled for using random effects in GLMMs (Table 2),
343 many of the major ion concentrations were similar between groundwater and snowmelt

344 streams. High inter-site variation, indicated by high standard deviations for random
345 effect terms and the large difference between marginal and conditional R^2 values,
346 masked differences between streams with different water sources (Table 2). Nitrate
347 (NO_3^-) concentrations were significantly different between snowmelt and groundwater
348 fed streams, with groundwater streams showing higher values (Figure 5).

349 **3.5. Metabolic activity**

350 Transformation rates of Raz to Rru were relatively low across all streams (0.58–1.22
351 $\times 10^{-3} \text{ s}^{-1}$) (Table 4). The groundwater stream Shimizugawa had a higher Raz to Rru
352 transformation ($1.22 \times 10^{-3} \text{ s}^{-1}$) than the other groundwater and the two snowmelt
353 streams. Indeed, the other groundwater stream, Bentenzawa, showed the lowest
354 transformation rate ($0.58 \times 10^{-3} \text{ s}^{-1}$). Variation between the snowmelt streams was
355 lower ($0.07 \times 10^{-3} \text{ s}^{-1}$) in comparison to that between the groundwater streams ($0.62 \times$
356 10^{-3} s^{-1}). The transformation rate for the snowmelt stream Dakesawa was not
357 determined as the peak of the breakthrough-curve at the upstream site was not
358 captured.

359 **4. Discussion**

360 Studies of alpine streams across the globe have shown relatively consistent patterns
361 of habitat templates (i.e., morphology, hydrology, physicochemistry) and biota in
362 relation to water source contributions (groundwater, snowmelt and glacial meltwater),
363 even across different biogeographical zones (Hannah *et al.* 2007; Brown *et al.* 2009).
364 Here, we show that variation in morphology, hydrology, physicochemistry and
365 metabolic signatures in streams across the Kamikochi region of Japan predominantly
366 conform to these patterns. Nevertheless, the high levels of inter-site variation, even
367 between streams with similar water source dynamics, particularly in those dominated

368 by snowmelt, appears a unique feature of the systems investigated in this study.
369 However, the relatively small number of streams studied here (n = 6) restricts widely
370 generalisable conclusions past this region of the Japanese Alps. However, if the highly
371 variable conditions observed in these streams are consistent across other alpine
372 stream systems in the Japanese archipelago, then these ecosystems could allow for
373 interesting additional insights in both hydrology and ecology in alpine stream systems.
374 Environmental conditions in the two groups of streams, groundwater and snowmelt,
375 were generally different from one another thereby supporting H₁ “*Water source will*
376 *significantly influence morphological, hydrological and physicochemical signatures*
377 *between the study streams*”. These findings are similar to those from other regions of
378 the globe (e.g., Europe, North America and the Arctic), where groundwater fed
379 streams typically show low temporal variation in discharge, water temperature and
380 water chemistry compared to other systems (Brown *et al.* 2003, 2006a, 2007b; Milner
381 *et al.* 2010). Studies from other regions commonly indicate that streams with the same
382 dominant water source are more similar to one another, i.e., groundwater streams are
383 generally all more stable with lower variation and snowmelt streams are similarly more
384 variable (Brown *et al.* 2007b, 2009; Hannah *et al.* 2007; Milner *et al.* 2010; Khamis *et*
385 *al.* 2016). The groundwater streams in the Kamikochi region were generally more
386 similar to one another than to snowmelt fed streams. Snowmelt fed streams exhibited
387 large amounts of variation between each stream. In particular, there were significant
388 differences in diurnal and seasonal variation. A large number of high flow events were
389 evident at the snowmelt stream, Tokusawa, in contrast with all other streams.
390 Discharge data, however, were not available for another snowmelt fed stream,
391 Shirasawa, due to sensor loss in 2018 and it is likely high flow events also occurred
392 here based on visual observations. Other findings, however, indicate Shirasawa

393 exhibited similar albeit less extreme hydrological variation, with temperature loggers
394 detecting flow cessation (low recorded temperatures due to air exposure or freezing;
395 Brown et al. 2006a) at a similar time to Tokusawa. Furthermore, the reason for the
396 sensor loss at the site was a large high flow event, again occurring at a similar time to
397 the series of high flow events detected in Tokusawa between April and November
398 2018.

399 Higher rates of metabolic activity are attributed to differences in water sources
400 contributions, specifically the high environmental stability in groundwater dominated
401 streams. As only one groundwater stream had a higher rate of metabolic activity than
402 the snowmelt streams, there was limited support for hypothesis H₂ "*Metabolic activity*
403 *varies significantly across sites and will be higher in the groundwater systems*".
404 Although in a study with a greater sample size different patterns may be observed,
405 several potential explanations exist for the patterns observed across the streams in
406 this study. In general, the findings in this study indicate that environmental stability
407 (i.e., stable flow regime, stable sediments and less variable water temperature), as
408 observed in the groundwater stream Shimizugawa, may be favourable for high rates
409 of metabolic activity. Across other groundwater streams the environmental conditions,
410 particularly consistent flow and temperature, enables the persistence of organic matter
411 and production of benthic algae, as well as a high rate of utilisation and decomposition
412 of these resources (Tank *et al.* 2010). Furthermore, the stability allows for the
413 colonisation of the streams by macroalgae and macrophytes, which have a strong
414 association with increased stream respiration and metabolism (Alnoee *et al.* 2016).
415 Kurz et al. (2017) modelled Raz transformation rates in vegetated flumes and found
416 that transformation rate in the advection-dispersion dominated mobile zone, where
417 vegetation was a dominant influence, was an order of magnitude higher than in the

418 transient-storage (immobile) zone. Thus, it is apparent that the abundances of
419 macrophytes, associated epiphytes and the influence of these plants on benthic
420 heterotrophic processes (Logue *et al.* 2004), influence the levels of metabolic activity
421 in streams. Certainly, Shimizugawa, the groundwater stream with Raz transformation
422 rates orders of magnitude higher than the other streams, had very high macrophyte
423 abundance (see Figure S1). Our understanding of the direct and indirect drivers of
424 stream metabolism in alpine streams with different water sources, however, is not well
425 developed and further research is required to fully determine the drivers of this
426 important ecosystem process.

427 Groundwater and snowmelt streams were generally dissimilar and groundwater
428 streams were less variable supporting the third hypothesis H₃ "*Hydrological,*
429 *physicochemical and metabolic signatures in groundwater streams will be less*
430 *variable when compared to snowmelt streams*". Thus, although the climatological
431 conditions present in the Japanese Alps are relatively unique, the dominant patterns
432 across streams (i.e., groundwater streams are more temporally stable) do not
433 markedly differ from other regions across the globe. However, significant variation in
434 water chemistry, water temperature and metabolic activity between the three
435 groundwater streams was found. These differences are likely due to a number of
436 reasons either solely or combined (i) ground water had a variable residence time in
437 subsurface flow paths; (ii) the dominant subsurface flow paths are relatively shallow
438 and the transit time of water through them is highly variable; or (iii) the underlying
439 geology and weathering processes across the three groundwater streams are
440 different. These reasons are supported by the hydrographs for the groundwater
441 streams, which although less variable than the snowmelt streams, still exhibited

442 delayed flow increases in both summer and winter in response to snow melt discharge
443 and rainfall (see Milner *et al.* 2020).

444 Across the stream systems investigated in this study, high inter-site variation was not
445 explained by the dominant water source contribution, particularly for snowmelt
446 streams. Studies from other alpine streams, suggest that even systems dominated by
447 one water source can exhibit differences in hydrology, water chemistry and ecosystem
448 functions, as a result of contributions of other water sources (e.g., interflow and surface
449 runoff; Brown *et al.* 2009). It is likely, therefore, that some of the variation in hydrology,
450 physicochemistry and metabolic activity between streams with the same dominant
451 water sources (snowmelt and groundwater) may be due to other water source
452 contributions. For example, Dakesawa, considered a snowmelt dominated stream,
453 showed intermediate degrees of variation in water temperature and discharge
454 between that of three groundwater streams and the two other snowmelt fed streams.

455 The findings of this study provide further evidence of the variability in the habitat
456 templates in alpine stream systems according to water source. The suite of
457 environmental conditions present in any given stream has important implications for
458 the structure and function of these ecosystems (Brown *et al.* 2003, 2009; Hannah *et al.*
459 *et al.* 2007). Here, we showed that stable morphological, hydrological and
460 physicochemical conditions in groundwater fed streams were typically associated with
461 higher rates of metabolic activity. This is likely due to a lower number of high flow
462 events and lower concentrations of suspended sediments (lower turbidity) in the
463 Kamikochi groundwater streams (see Milner *et al.* 2020), allowing for greater retention
464 of organic detritus and higher algal production, respectively. The lower temporal
465 variation in hydrological and physicochemical conditions in groundwater stream
466 systems potentially enables higher levels of primary productivity, as has been

467 documented in other stream systems (LaPerriere *et al.* 1989; Tockner *et al.* 1997;
468 Klein & Tockner 2000; Rott *et al.* 2006). Higher primary productivity is also expected
469 to translate to greater levels of secondary productivity, and is a potential explanation
470 for observed differences in macroinvertebrates between stream systems with different
471 water sources both in the Kamikochi streams (Milner *et al.* 2020), and other alpine
472 stream systems (Brown *et al.* 2004, 2006b). In a corollary study we showed that a
473 greater macroinvertebrate abundance in groundwater streams may have resulted from
474 higher chlorophyll *a* concentrations, an indicator of primary productivity (Milner *et al.*
475 2020).

476 **Conclusion**

477 In summary, snowmelt and groundwater streams in this unique bioregion of the
478 Japanese Alps showed highly variable environmental conditions between them,
479 providing a range of habitat templates for biotic communities. Metabolic activity, an
480 important indicator of ecosystem functioning, varied between streams with different
481 water sources due to environmental stability and its effects on biological communities
482 (e.g., submerged macrophytes). Patterns in environmental conditions across the
483 streams in the Japanese Alps were quite similar to other alpine regions of the globe,
484 but it was clear that the unique climatological conditions in the Japanese Alps did
485 create several noticeable differences, including greater inter-stream variability. This
486 high variability of the environmental conditions within and between stream systems,
487 combined with the unique flora and fauna of this biogeographic region therefore make
488 it an invaluable reserve of biodiversity across the globe.

489 Further research is now required to advance the understanding of how environmental
490 change may affect these systems into the future. The high variability in flow conditions
491 present in the snowmelt streams indicates that further meteorological change, for

492 example reduced snowpacks (Sato *et al.* 2013) and increased frequency of typhoons
493 alongside heavy rainfall events (Tsunematsu *et al.* 2013), may lead to greater
494 frequencies of extreme high and low flow events. For example, reduced snowpacks
495 may lead to increased intermittency during low flow periods of July and August in
496 snowmelt fed streams. Groundwater fed streams appear more resilient to these
497 alterations, with far lower hydrological variation, yet the impact of extreme events in
498 snowmelt streams may be significant enough to generate changes in aquatic
499 biodiversity. Hence management methods to mitigate potential losses in biodiversity
500 are essential.

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7. Figures

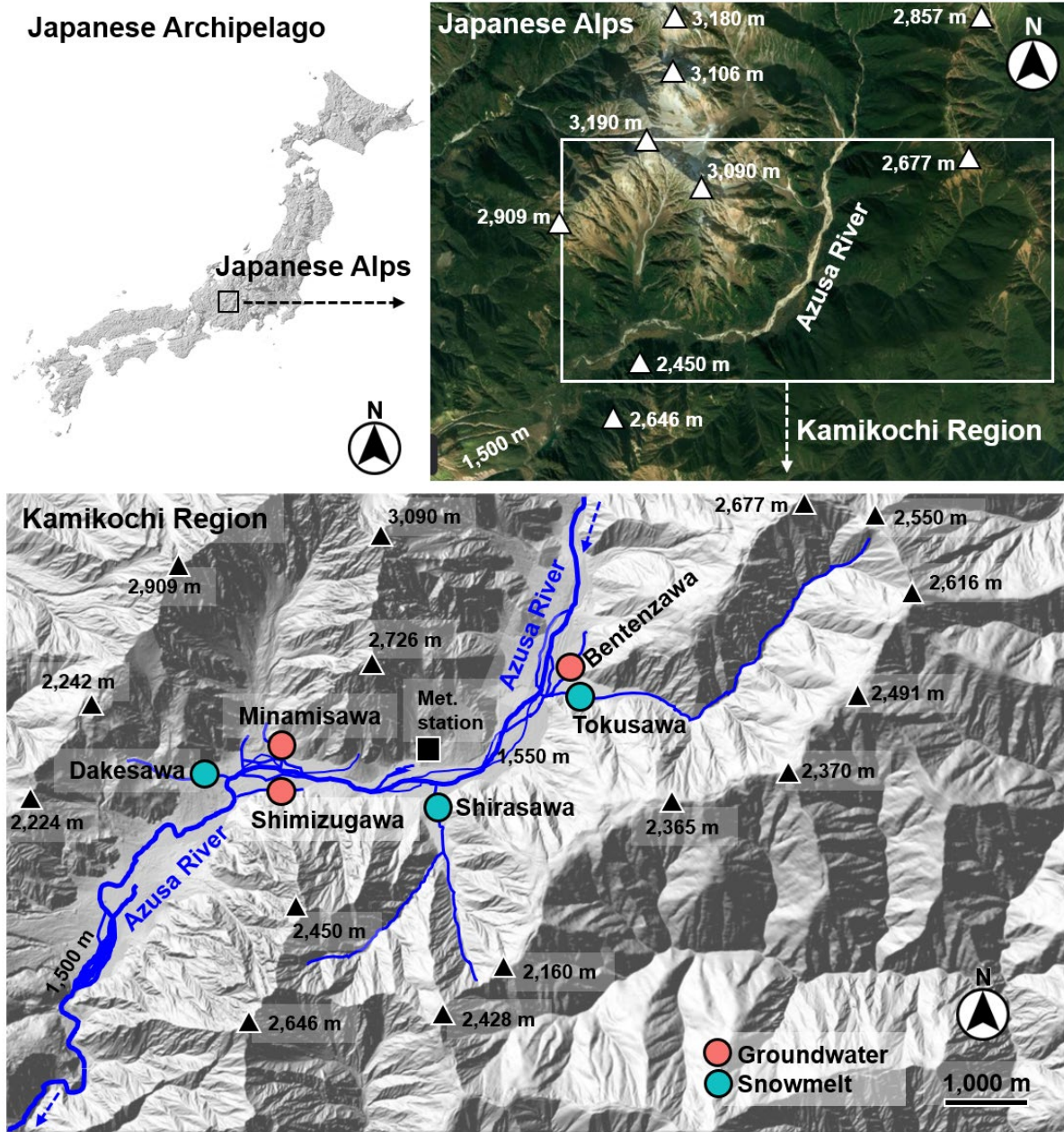


Figure 1. Sample sites at the six streams within the Kamikochi region of the Japanese Alps.

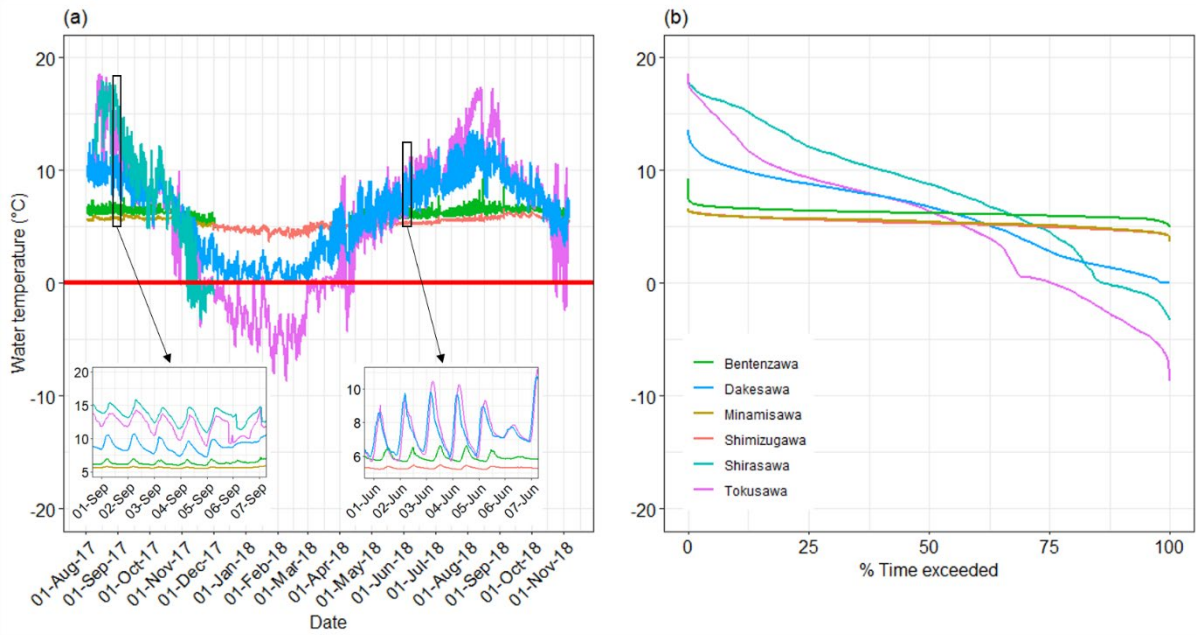


Figure 2. Water temperature (°C) across the groundwater and snowmelt fed streams. (a) Water temperature time series data. (b) Water temperature duration curves (Minamisawa and Shimizugawa overlap). Following Brown et al. (2006a), temperatures below 0 °C in (a) are indicative of flow cessation either due to freezing or drying processes. Inset graphs show diurnal variation over two separate weeks in September 2017 and June 2018.

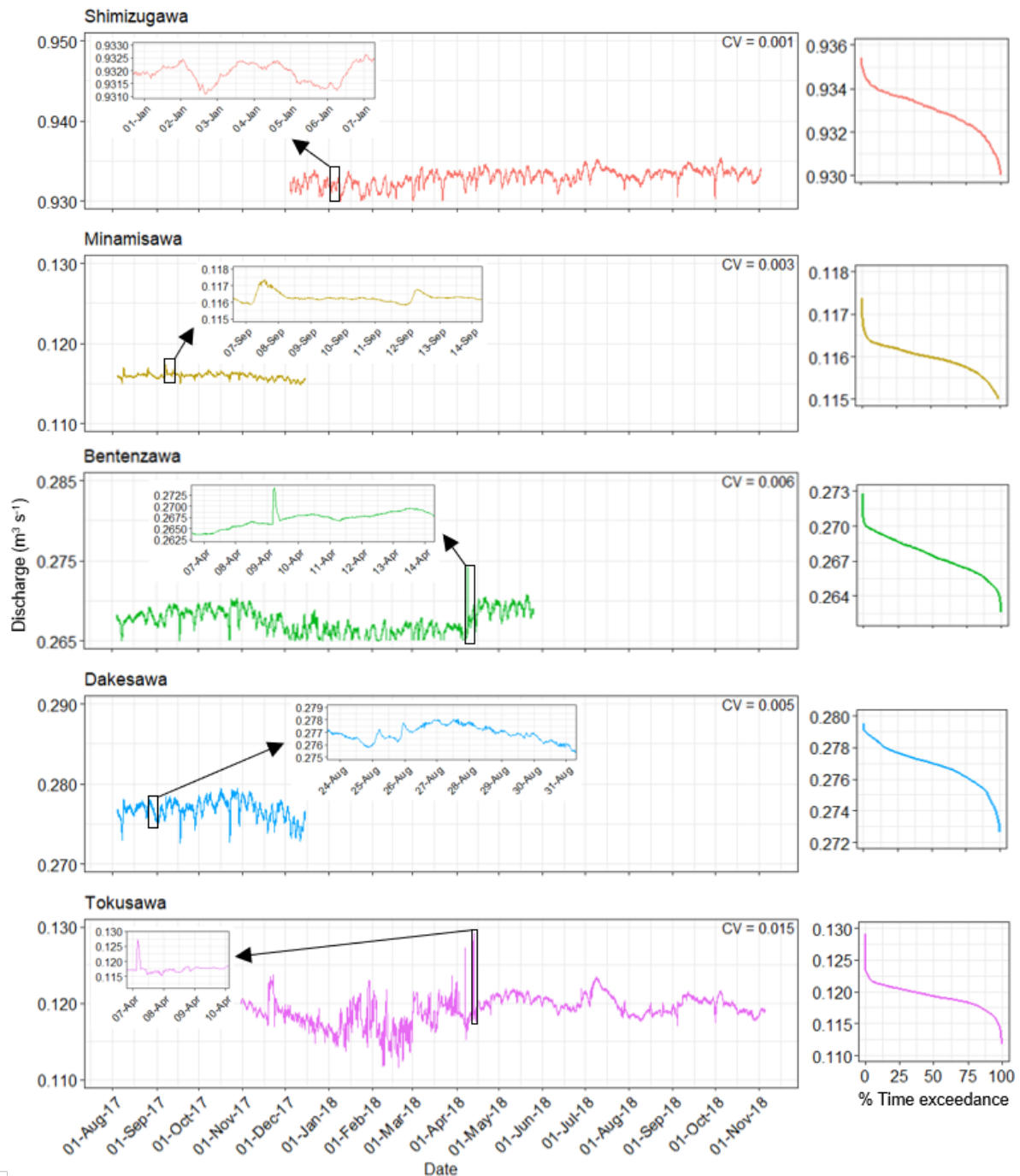


Figure 3. Stream discharge time series and variation estimates. CV = Coefficient of variation. Inset figures demonstrate discharge variation at diurnal timescales for periods selected from the wider time series in each stream. Figures to the right of the time series are flow duration curves for the time series data (x axis indicates % time exceedance for the given discharge value on the y axis).

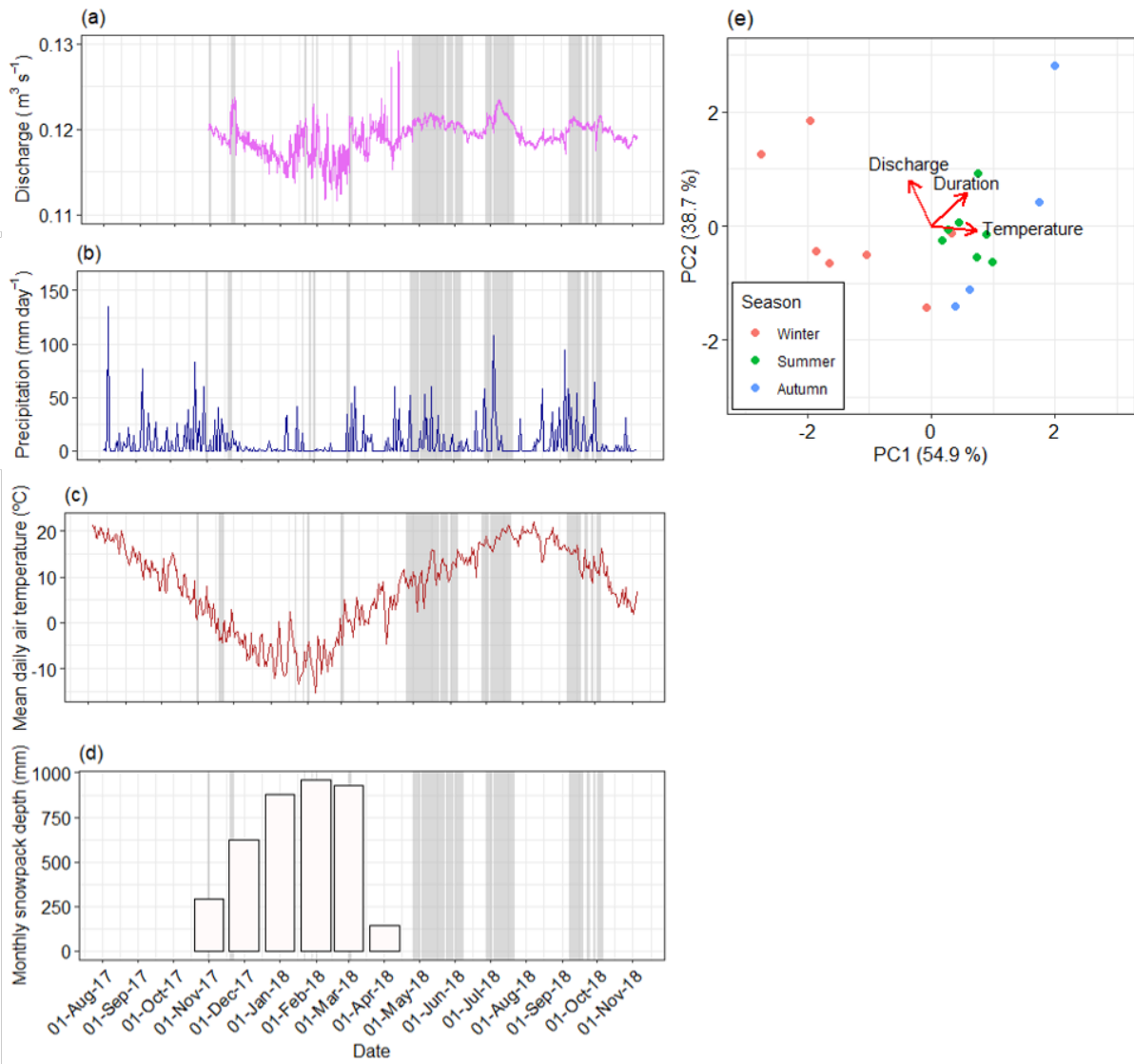


Figure 4. Storm events and characteristics over the sampling period in a snowmelt fed stream (Tokusawa). (a) Stream discharge time series. (b) Daily precipitation, (c) mean daily air temperature and (d) monthly measured snowpack depth from the Kamikochi Meteorological Station. (e) Storm event characteristics from principal components analysis (n = 18).

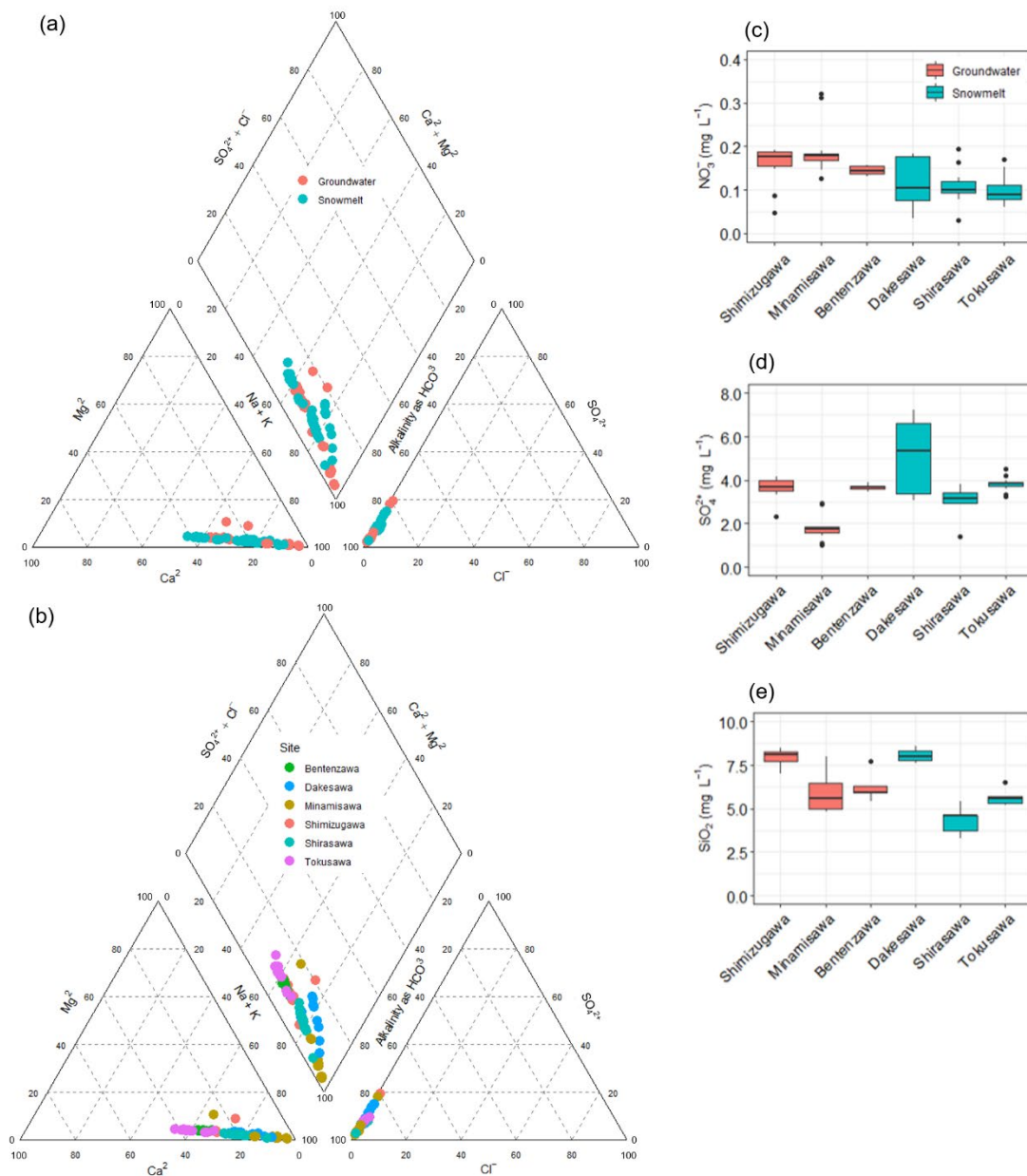


Figure 5. Variation in regularly detected major ion concentrations across the six streams in the Kamikochi region of the Japanese Alps. For piper diagrams (a-b) mg L^{-1} data were converted in percentage meq L^{-1} using the ‘smwrBase’ and ‘hydrogeo’ packages in R. Major ions not included in (a-b) are presented in boxplots in (c-e). Colours in (c-e) indicate the dominant water source contribution within each stream (red = groundwater, blue = snowmelt).

8. Tables

Table 1. Coordinates and physical characteristics of sample streams.

| Site | Decimal degrees | | Water source | Mean width (m) | Mean depth (cm) | Pfankuch Index |
|-------------|-----------------|------------|--------------|----------------|-----------------|----------------|
| | Latitude | Longitude | | | | |
| Shimizugawa | 36.250482 | 137.639672 | GW | 9.4 | 27.0 | 21 |
| Minamisawa | 36.256573 | 137.643104 | GW | 7.5 | 26.4 | 30 |
| Bentenzawa | 36.265480 | 137.690354 | GW | 4.6 | 21.8 | 20 |
| Tokusawa | 36.261673 | 137.693700 | SM | 4.2 | 17.0 | 34 |
| Dakesawa | 36.253908 | 137.634887 | SM | 4.1 | 14.7 | 38 |
| Shirasawa | 36.248523 | 137.670093 | SM | 3.6 | 11.2 | 47 |

Table 2. Variation in major ion concentrations. Statistical results are derived from a series of GLMMs. Df = Degrees of freedom (numerator and denominator).

| Ion | Mean (min–max) (mg L ⁻¹) | Model results | | | | | | |
|---------------------------------------|---|------------------|------------------|--------|-------------------------------|--------------------|----------|---------|
| | | R ² m | R ² c | Df | Random effect (variance ± SD) | Variable | t value* | p value |
| Na ⁺ | 1.34 (0.53–2.02) | 0.28 | 0.73 | 12, 67 | 0.08 ± 0.28 | Water source | 0.13 | 0.74 |
| | | | | | | Sample time (date) | 7.3 | <0.001 |
| K ⁺ | 0.22 (0.01–0.07) | 0.22 | 0.54 | 12, 67 | 0.01 ± 0.08 | Water source | 0.06 | 0.82 |
| | | | | | | Sample time (date) | 3.47 | 0.001 |
| Mg ²⁺ ** | 0.35 (0.07–1.29) | 0.03 | 0.90 | 12, 64 | 0.03 ± 0.16 | Water source | 0.02 | 0.9 |
| | | | | | | Sample time (date) | 2.71 | 0.006 |
| Ca ²⁺ | 4.47 (0.57–8.3) | 0.04 | 0.92 | 12, 66 | 5.8 ± 2.41 | Water source | 0.02 | 0.91 |
| | | | | | | Sample time (date) | 3.5 | 0.001 |
| Cl ⁻ | 0.3 (0.13–0.78) | 0.56 | 0.71 | 12, 68 | 0.002 ± 0.05 | Water source | 0.01 | 0.93 |
| | | | | | | Sample time (date) | 13.43 | <0.001 |
| NO ₂ ⁻ | 0.001 (0–0.003) | 0.38 | 0.38 | 12, 44 | 0 ± 0 *** | Water source | 2.04 | 0.047 |
| | | | | | | Sample time (date) | 2.72 | 0.009 |
| NO ₃ ⁻ ** | 0.14 (0.03–0.32) | 0.75 | 0.77 | 12, 67 | 0.0001 ± 0.007 | Water source | 6.73 | 0.002 |
| | | | | | | Sample time (date) | 12.71 | <0.001 |
| SO ₄ ²⁺ **** | 3.54 (1.03–9.36) | 0.15 | 0.60 | 12, 67 | 1.54 ± 1.24 | Water source | 0.43 | 0.55 |
| | | | | | | Sample time (date) | 1.89 | 0.056 |
| SiO ₂ **** | 6.28 (3.3–8.6) | 0.17 | 0.66 | 5, 19 | 1.19 ± 1.09 | Water source | 1.77 | 0.27 |
| | | | | | | Sample time (date) | 0.88 | 0.51 |

* Calculated from single term deletions using Satterthwaite's method

** Outlying values (Cook's distance > 0.5) were removed from the model (n = 1–2)

*** A singular model, i.e., parameters are on the boundary of the feasible parameter space, as NO₂⁻ was not detected across all sites

**** Non-significant models

Table 3. Raz transformation rates and mean water temperature at injection for four streams in the Kamikochi region of the Japanese Alps.

| Site | Water source | Raz transformation rate (10^{-3} s^{-1}) | Temperature ($^{\circ}\text{C}$) during injection \pm SD |
|-------------|---------------------|--|---|
| Shimizugawa | GW | 1.22 | 5.39 ± 0.08 |
| Bentenzawa | GW | 0.58 | 5.04 ± 0.09 |
| Shirasawa | SM | 0.88 | 3.70 ± 0.08 |
| Tokusawa | SM | 0.81 | 3.25 ± 0.05 |