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1 **Dietary shifts drive the slowdown of declining methylmercury related health risk**
2 **in China**

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16 **Abstract**

17 Chinese population suffers severe health risk from dietary methylmercury (MeHg)
18 exposure. However, the temporal change of such risk and socioeconomic driving
19 factors remain unknown. This study investigates this issue by compiling time-series
20 inventory of China's MeHg-related health risk at the provincial scale and revealing
21 critical socioeconomic influencing factors through structural decomposition analysis.
22 Results show that the per-fetus IQ decrements from dietary MeHg exposure have
23 declined by 60% nationally during 2004-2019. Such decline results from the joint
24 effects of dietary shifts (contributing 44%) and the decrease of MeHg concentrations in
25 foods consumed (contributing 56%). However, the declining trend has slowed down
26 since 2014 and even leveled off after 2016, which is mainly affected by dietary pattern
27 changes. Especially, the increased intake level and proportion of fishes in
28 underdeveloped provinces of China have dominated the slowdown of declining trend
29 after 2016. Moreover, the affluence and education levels have significantly negative
30 associations with per-fetus IQ decrements. Rich and well-educated people have higher
31 ability of risk perception, which indicates the importance of rational consumption
32 patterns. Our findings can help develop socioeconomic regulatory policies on reducing
33 per-fetus IQ decrements from dietary MeHg exposure in China.

34 **Keywords:** Mercury pollution; Human health; Methylmercury exposure; Food
35 system; Socioeconomic factors; Dietary patterns.

36 **Introduction**

37 Mercury (Hg) is a global pollutant that can be transported globally and impose
38 adverse impacts on human beings (Chen et al., 2019; Giang and Selin, 2016; Li et al.,
39 2020a). One of its most toxic forms, methylmercury (MeHg), can bioaccumulate in
40 food webs and pose serious health risk to human beings (Driscoll et al., 2013; Sundseth
41 et al., 2017). Exposure to MeHg has been associated with neurodevelopmental delays
42 in children and cardiovascular impairment in adults (Grandjean et al., 2012; Hu et al.,
43 2017; Roman et al., 2011). In addition, it has adverse effects on renal, reproductive, and
44 immune systems (Bose-O'Reilly et al., 2016; Henriques et al., 2019). The World Health
45 Organization (WHO) has treated Hg as one of the top ten chemicals of major public
46 health concern (World Health Organization, 2017). To protect people from Hg-related
47 health risk, 128 nations (including China) have signed the Minamata Convention on
48 Mercury. The Article 16 of the convention emphasizes “Health aspects” and Article 19
49 points out the requirements for assessing the health risk of Hg (United Nations
50 Environment Programme, 2013).

51 China is particularly important in global Hg cycle. On one hand, China is the largest
52 Hg emitter in the world, contributing about 25% (565 t) of global atmospheric Hg
53 emissions in 2015 (United Nations Environment Programme, 2019). On the other hand,
54 Chinese population has suffered from serious MeHg-related health risk. For example,
55 MeHg intake in China resulted in 0.14 points of per-fetus IQ decrements and 7360
56 deaths from fatal heart attacks in 2010 (Chen et al., 2019). Moreover, China has been
57 experiencing in recent years and will continue to experience tremendous

58 socioeconomic transitions which leads to significant changes in people's dietary
59 patterns (He et al., 2019). Understanding the historical temporal changes of health risk
60 from dietary MeHg exposure and identifying socioeconomic driving factors can
61 provide more explicit hotspots for socioeconomic policies to reduce health risk from
62 dietary MeHg exposure in China. However, these points have not been well
63 characterized.

64 Existing studies have investigated the food sources of the health risk of MeHg. Fish
65 consumption is assumed to be the main dietary source of MeHg exposure for most
66 populations worldwide (Liu et al., 2018b; Sunderland et al., 2018). However, for certain
67 regions in China, especially in southern inland areas and Hg mining areas, rice
68 consumption is the most important pathway of dietary MeHg exposure (Li et al., 2012;
69 Zhang et al., 2010). In addition to quantifying health risks of dietary MeHg exposure,
70 previous studies have also considered the interregional food trade among Chinese
71 regions and estimated the health risks of MeHg intake from the viewpoint of whole
72 food system (Liu et al., 2020). The interregional food trade has significant impacts on
73 health risks of dietary MeHg intake in China. Taking the interregional trade into
74 account, studies have quantified the health risk of dietary MeHg exposure based on
75 particular time points in China (Chen et al., 2019; Li et al., 2020a; Li et al., 2020b).
76 However, the temporal change of health risk from dietary MeHg exposure in China and
77 the socioeconomic driving factors remain unknown.

78 The main objective of this study is to quantify the temporal changes of health risk
79 from dietary MeHg exposure at the provincial level in China during 2004-2019 and

80 identify critical socioeconomic factors influencing changes in health risk, including
81 midpoint factors (i.e., dietary factors including dietary intake level, dietary structure,
82 and food trade structure) and endpoint factors (i.e., underlying factors including
83 affluence, education levels, and traffic accessibility). Previous studies find that major
84 adverse effects of MeHg exposure are neurological, and the neurological effects are
85 closely related to dietary intake (Gao et al., 2014). Thus, this study takes the per-fetus
86 IQ decrements as the indicator for the health risk of dietary MeHg exposure. We first
87 constructed the inventory of China's provincial MeHg-related health risk during 2004-
88 2019, including compiling MeHg concentrations of foods, simulating the interregional
89 food trade, evaluating the MeHg intake, and estimating per-fetus IQ decrements from
90 dietary MeHg exposure. We then identified the midpoint and endpoint factors
91 influencing changes in the health risk by using structural decomposition analysis (SDA)
92 method and the panel regression model. Our findings can help develop socioeconomic
93 regulatory policies on reducing per-fetus IQ decrements from dietary MeHg exposure
94 in China. Moreover, these findings are enlightening to similar nations around the world
95 and thus can help promote the progress of the Minamata Convention on Mercury
96 globally.

97 **Methodology**

98 The research framework of this study consists of three components: compiling time-
99 series inventory of China's per-fetus IQ decrements from dietary MeHg exposure at the
100 provincial scale, analyzing the sources and evolution of China's per-fetus IQ
101 decrements from dietary MeHg exposure, and identifying critical socioeconomic

102 factors influencing the health risk changes ([Fig. S1](#)). This study considers 30 provinces
103 in mainland China and the coastal seas. Tibet, Hongkong, Macau, and Taiwan are not
104 considered, due to data unavailability. The details on the methods can be found in the
105 [Supplementary Information](#).

106 **Compiling provincial inventories of per-fetus IQ decrements from dietary MeHg**
107 **exposure.** This study constructs China's provincial inventories of dietary MeHg
108 exposure during 2004-2019. This part includes four steps: compiling MeHg
109 concentrations of foods, simulating the interregional trade of foods, evaluating the
110 estimated daily intake (EDI) of MeHg, and estimating per-fetus IQ decrements from
111 dietary MeHg exposure.

112 First, this study compiles MeHg concentrations of foods in China. The categories of
113 foods in this study include marine fishes, freshwater fishes, and rice. We collected both
114 total Hg (THg) and MeHg concentrations of these foods from existing studies. The list
115 of literature for THg and MeHg concentrations of foods is provided in Supplementary
116 Data 8. For the estimation of MeHg exposure, the data near contaminated sites are
117 excluded to reduce the errors. For species with only THg concentrations, we converted
118 the data into MeHg by linear relationships which were estimated from studies with both
119 THg and MeHg observations ([Fig. S2](#)). We average the MeHg concentrations every
120 four years during 2004-2019 to investigate the general trend of MeHg concentration
121 changes, which can reduce the impact of MeHg concentration changes of particular
122 time points on the general trend. The concentrations of marine fishes, freshwater fishes,
123 and rice are shown in [Fig. S3](#).

124 Second, this study simulates the interprovincial food trade by using the multi-
125 regional input-output (MRIO) tables (Lenzen et al., 2012; Liang et al., 2014). It has
126 been widely used to simulate the interprovincial trade of foods in previous studies
127 (Deng et al., 2020; Liu et al., 2018b). Monetary data on the final demand (including
128 urban household consumption, rural household consumption, and government
129 consumption) of a given province in the MRIO tables are used to describe the sources
130 of a specific food product consumed in the province. Different sectors in the MRIO
131 tables are used to simulate the interprovincial trade of different food products. We
132 introduce a ratio (i.e., the output value of a specific fish to gross agricultural output) to
133 extract the final demand for the specific fish from the total final demand of the *Farming,*
134 *Forestry, Animal Husbandry and Fishery* sector in the MRIO tables.

135 Third, based on the MeHg concentrations of foods, interprovincial food trade, food
136 consumption, and the intake rate of each category of foods, this study calculates the
137 EDI of MeHg by equation 1.

$$138 \quad EDI_j = \sum_{ik} \frac{SC_{ijk} \times I_{ij} \times C_{ik}}{W} \quad (1)$$

139 The notation EDI_j represents the EDI of MeHg by the population in province j ;
140 SC_{ijk} indicates the source contribution of food i in province j driven by the supply
141 of province k ; I_{ij} represents the per capita intake rate ($\text{g d}^{-1} \text{capita}^{-1}$) of food i
142 consumed by the females in province j ; C_{ik} is the MeHg concentration (ng g^{-1}) of
143 food i harvested in province k ; and W represents the average body weights of
144 Chinese females of childbearing age (15-50 years old). The data for female

145 consumption of foods are obtained from the China Health and Nutrition Survey (CHNS)
146 database.

147 Finally, this study estimates per-fetus IQ decrements due to MeHg intake in various
148 provinces of China. MeHg can be transmitted to the fetus through prenatal exposure
149 and damage the brain tissues of the fetus, resulting in neurodevelopmental disorders
150 and IQ decrements for the fetus (Driscoll et al., 2013). The IQ decrements caused by
151 MeHg exposure are calculated by the dose-response relationship, based on previous
152 epidemiological studies (Axelrad et al., 2007; Giang and Selin, 2016; Rice et al., 2010).
153 The assessment of the IQ effects is shown in equation 2.

$$154 \quad \Delta IQ = \gamma \lambda \beta (\Delta EDI \times W) \quad (2)$$

155 The notation ΔIQ represents the change in IQ points and ΔEDI indicates the
156 change in EDI of MeHg. The notation W represents the average body weight of
157 Chinese female adults. The blood-intake coefficient β ($\mu\text{g Hg/L blood per } \mu\text{g Hg/day}$),
158 hair-blood coefficient λ ($\mu\text{g Hg/g hair per } \mu\text{g Hg/L blood}$), and IQ-hair mercury
159 coefficient γ (IQ points per $\mu\text{g Hg/g hair}$) represent the conversion factors of MeHg
160 from intake to blood, blood to hair, and hair to IQ, respectively.

161 **Quantifying contributions of midpoint factors.** In this study, we use the SDA method
162 to investigate the relative contributions of different factors to the changes of provincial
163 per-fetus IQ decrements during 2004-2019 (Dietzenbacher and Los, 1998; Hoekstra
164 and Bergh, 2002). We divide the ΔEDI of each province into the contributions of four
165 independent variables including dietary intake level change, dietary structure change,

166 food trade structure change, and MeHg concentration change. Therefore, for food
 167 category i and the province j , the ΔEDI in each province can be written as equation
 168 3.

$$169 \quad \Delta EDI = LS(O \odot C)e \quad (3)$$

170 The notation L represents the total food intake level per capita in this province; the
 171 $1 \times i$ row vector S represents the dietary structure, indicating the proportion of food i
 172 intake to the total food intake in this province; the \odot represents the element-wise matrix
 173 multiplication; the $i \times j$ matrix O indicates the food trade structure, indicating the
 174 proportion of food i from province j to the total consumption of the population in
 175 this province; the $i \times j$ matrix C indicates MeHg concentrations of foods, meaning the
 176 MeHg concentrations of food i in province j ; and the notation e represents a $j \times 1$
 177 unit vector.

178 Subsequently, the per-fetus IQ decrements can be described by equation 4 and the
 179 structural decomposition form is shown in equation 5.

$$180 \quad \Delta IQ = \gamma\lambda\beta \cdot LS(O \odot C)e \quad (4)$$

$$181 \quad \Delta IQ = \gamma\lambda\beta \cdot \Delta LS(O \odot C)e + \gamma\lambda\beta \cdot L\Delta S(O \odot C)e + \\ \gamma\lambda\beta \cdot LS(\Delta O \odot C)e + \gamma\lambda\beta \cdot LS(O \odot \Delta C)e \quad (5)$$

182 The notation ΔIQ represents the change in per-fetus IQ decrements. Items on the
 183 right side of equation 5 indicate the per-fetus IQ decrement change ΔIQ caused by
 184 dietary intake level change ΔL , dietary structure change ΔS , food trade structure
 185 change ΔO and MeHg concentration change ΔC . The SDA method has the problem
 186 of non-uniqueness (Rørmose and Olsen, 2005). There will be $n!$ kinds of decomposition

187 forms if the number of decomposed factors is n . We take the average of all first-order
188 decomposition forms to address this problem (Dietzenbacher and Los, 1998; Guan et
189 al., 2008; Liang et al., 2013).

190 **Endpoint factors influencing per-fetus IQ decrements from dietary MeHg**
191 **exposure.** Socioeconomic factors outside the food system can be uncovered by the
192 panel regression model (Cheng et al., 2021). The endpoint factors considered in this
193 study include affluence level (represented by resident incomes), education level, and
194 traffic accessibility (represented by road areas). Based on the results of the F-test and
195 Hausman test, this study chooses an individual fixed effect model (Stuart et al., 2010)
196 (equation 6).

$$197 \quad \quad \quad \ln IQ_{it} = \beta X_{it} + \mu_i + \varepsilon_{it} \quad (6)$$

198 The notation i indicates province i and notation t indicates year t ; $\ln IQ$
199 represents the natural logarithm of per-fetus IQ decrements; β represents the
200 regression coefficient of each independent variable; X represents each independent
201 variables; μ_i represents the province-specific effect for province i ; and ε_{it}
202 represents the stochastic disturbance term.

203 **Uncertainties.** The uncertainties of results in this study mainly come from the
204 compilation of MeHg intake inventory and the evaluation of per-fetus IQ decrements.
205 A Monte Carlo simulation with 10,000 samplings was conducted on variables in
206 different stages. We set P10 and P90 values of the statistical distributions as lower and
207 upper ranges of results, following previous studies (Chen et al., 2019). In 2010, the

208 population-weighted mean per-fetus IQ decrements in this study were 0.052 (0.009-
209 0.105) points. They are lower than that of Chen et al. (0.140 points) in 2010 (Chen et
210 al., 2019). Compared to the previous study, this study has conducted more detailed
211 calibrations for MeHg concentrations of foods by removing the sample data on MeHg
212 concentrations of foods from heavily contaminated areas (because these foods are
213 generally unavailable in the market). In addition, only rice, marine fishes, and
214 freshwater fishes were considered in this study, rather than all foods in the previous
215 study (Chen et al., 2019). This is because that fishes and rice are the most important
216 pathway of dietary MeHg exposure, while MeHg in other foods is less detected (Li et
217 al., 2012; Zhang et al., 2010). All the uncertainty results are presented in Supplementary
218 Data 6.

219 **The verification of model estimates.** Previous studies have found that blood
220 mercury concentrations can serve as an indicator of short-term exposure to organic
221 mercury (Hong et al., 2016; Mahaffey et al., 2004). Therefore, we estimated the
222 mercury levels in human blood based on the dose-response relationship between MeHg
223 exposure and Hg biomarkers (equation 7) and then compared them with the measured
224 data from existing studies of the corresponding years.

$$225 \quad THg_{blood} = \beta \times \Delta EDI \quad (7)$$

226 The notation THg_{blood} represents the total Hg concentrations of blood; and β is the
227 blood-intake conversion coefficient between MeHg intake and MeHg in blood (μg
228 Hg/L blood per μg Hg/ day).

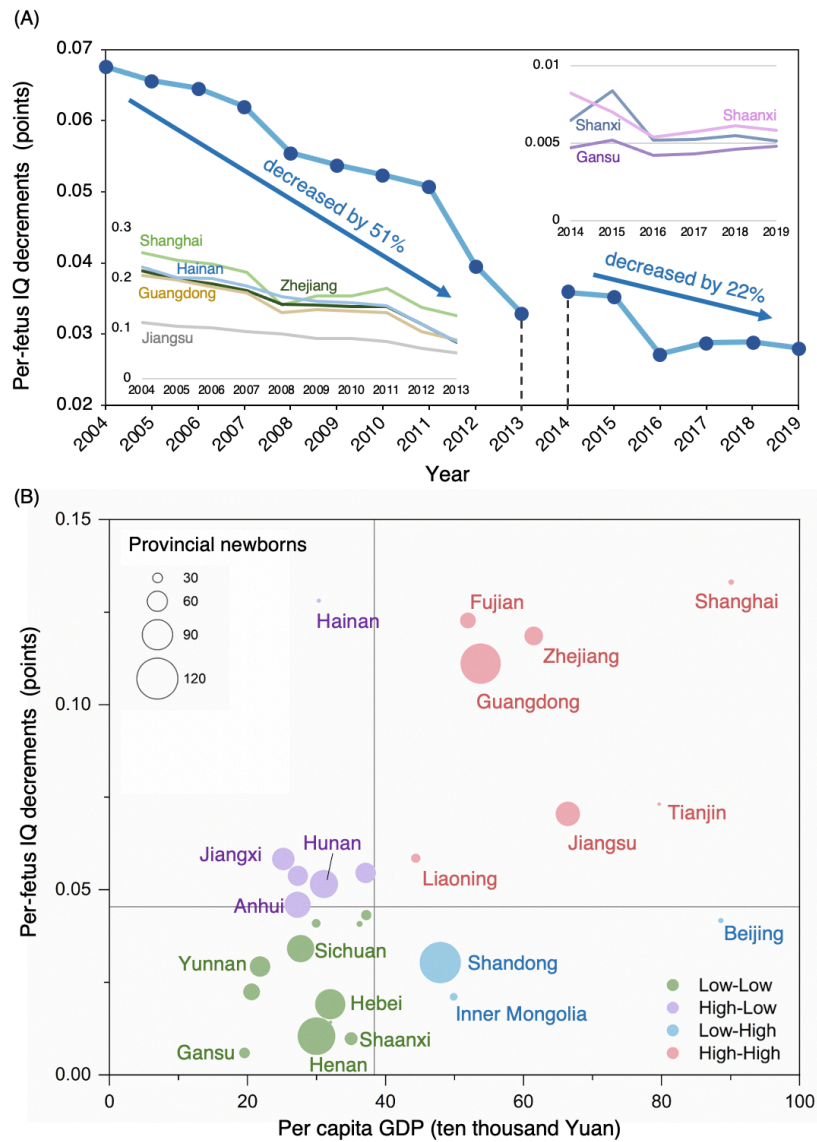
229 [Fig. S4](#) shows the relationship between measured data from existing studies and our
230 model estimates. It shows that the model estimates for provinces with larger newborn
231 populations are closer to the measured data. The national population-weighted mean of
232 blood Hg concentration of the measured data is 1.60 ± 0.57 $\mu\text{g/L}$. This is generally
233 consistent with the national mean of model estimate (1.50 ± 0.13 $\mu\text{g/L}$) in this study.

234 **Results**

235 **Changing per-fetus IQ decrements from dietary MeHg exposure.** China's national
236 per-fetus IQ decrements (national average using the neonatal population as the weight)
237 from dietary MeHg exposure have continuously decreased by nearly 60% during 2004-
238 2019, from 0.068 (0.011-0.132) points in 2004 to 0.028 (0.002-0.067) points in 2019.
239 Per-fetus IQ decrements in China have experienced a rapid decline during 2004-2013,
240 falling by 51% ([Fig. 1A](#)). The provinces with the greatest declines of per-fetus IQ
241 decrements during this period include Hainan, Zhejiang, Guangdong, Shanghai, and
242 Jiangsu. They are all located along the southern coast of China. However, during 2014-
243 2019, China's national per-fetus IQ decrements has decreased by only 22% at a
244 significantly slower rate. Moreover, the downward trend has flattened after 2016, with
245 3% increase in per-fetus IQ decrements during 2016-2019. This situation is mainly due
246 to the increased per-fetus IQ decrements in Gansu, Shaanxi, and Shanxi. Gansu has the
247 largest increase in per-fetus IQ decrements (by 14%) during 2016-2019. Given the
248 generally declining trend of MeHg concentrations in foods consumed ([Fig. S3](#)), the
249 slowdown and flattening of the declining trend is attributable to the changes in dietary
250 patterns of the populations. It is worth noting that, the sudden increase of per-fetus IQ

251 decrements during 2013-2014 is mainly due to the change in the scope of statistics
252 between these two years.

253 The relationship between per-fetus IQ decrements and economic development
254 (represented by per capita GDP in this study) has been further investigated (Fig. 1B).
255 Some provinces with low per capita GDP have high per-fetus IQ decrements (e.g.,
256 Jiangxi, Hunan, and Anhui). These regions are mainly located in the middle reaches of
257 the Yangtze River (Fig. S5). Provinces in coastal areas have high per capita GDP and
258 high per-fetus IQ decrements (e.g., Shanghai and Guangdong). Provinces with high per
259 capita GDP and relatively low per-fetus IQ decrements are mainly located in northern
260 China, including Beijing, Shandong, and Inner Mongolia. In addition, provinces with
261 low per capita GDP and low per-fetus IQ decrements are mainly located in the North
262 and Southwest China (see Fig. S6 for regional grouping). Overall, the per-fetus IQ
263 decrements of the North are generally lower than the South. Most of the southern
264 provinces are high-risk areas, which may be related to their dietary patterns. Notably,
265 provinces with relatively low per capita GDP are the main contributors to the flattening
266 of risk declines after 2016, including Gansu, Shaanxi, and Shanxi (Fig. S7). Thus, for
267 regions with relatively low per capita GDP, the changes in health risk caused by their
268 rapid economic development deserve our attention.



269

270 **Figure 1. National and provincial per-fetus IQ decrements in China due to dietary**
 271 **MeHg exposure during 2004-2019.** Graphs showing the temporal change of
 272 population-weighted mean per-fetus IQ decrements in China (A) and the relationship
 273 between average provincial per-fetus IQ decrements and per capita GDP (B). The
 274 bottom left corner of panel A represents the top five provinces with the largest declines
 275 of IQ loss during 2004-2013, and the upper right corner represents the provinces with
 276 relatively less decline in per-fetus IQ decrements during 2014-2019. Based on the
 277 relationship between per-fetus IQ decrements and per capita GDP during 2004-2019,
 278 we divided China's 30 provinces into four groups: High-High (i.e., High risk High per
 279 capita GDP), Low-High (i.e., Low risk-High per capita GDP), Low-Low (i.e., Low
 280 risk-Low per capita GDP), and High-Low (i.e., High risk-Low per capita GDP).

281 **Food sources of per-fetus IQ decrements from dietary MeHg exposure.** Food
282 sources of per-fetus IQ decrements from dietary MeHg exposure in China during 2004-
283 2019 are shown in Fig. 2. The consumption of fishes and rice is considered as the most
284 significant source of dietary MeHg in existing studies (Giang and Selin, 2016; Qing et
285 al., 2022; Selin et al., 2010; Zhang et al., 2010). Moreover, MeHg is mainly detected in
286 aquatic products and rice, and the data on MeHg concentrations of other foods are
287 usually unavailable (Li et al., 2010). Thus, this study only considers per-fetus IQ
288 decrements caused by the consumption of marine fishes, freshwater fishes, and rice.

289 Fig. 2A shows **temporal changes of food sources** for the per-fetus IQ decrements
290 during 2004-2019. The per-fetus IQ decrements caused by marine fishes have increased
291 significantly, with the proportion rising by about 13 percentage points during this
292 period. Six provinces including Heilongjiang, Tianjin, Shanghai, Guangxi, Beijing, and
293 Hebei contributed about 50% of the increase in marine fish sources. In contrast, the per-
294 fetus IQ decrements induced by MeHg exposure from rice intake decreased by 14
295 percentage points. This decline is mainly contributed by several western (e.g., Sichuan,
296 Ningxia, Shaanxi, and Xinjiang) and central (e.g., Anhui and Jiangxi) provinces. The
297 per-fetus IQ decrements caused by freshwater fish intake remain relatively stable
298 during 2004-2019. These changes in food sources reflect the transition of people's
299 dietary patterns due to rising living standards.

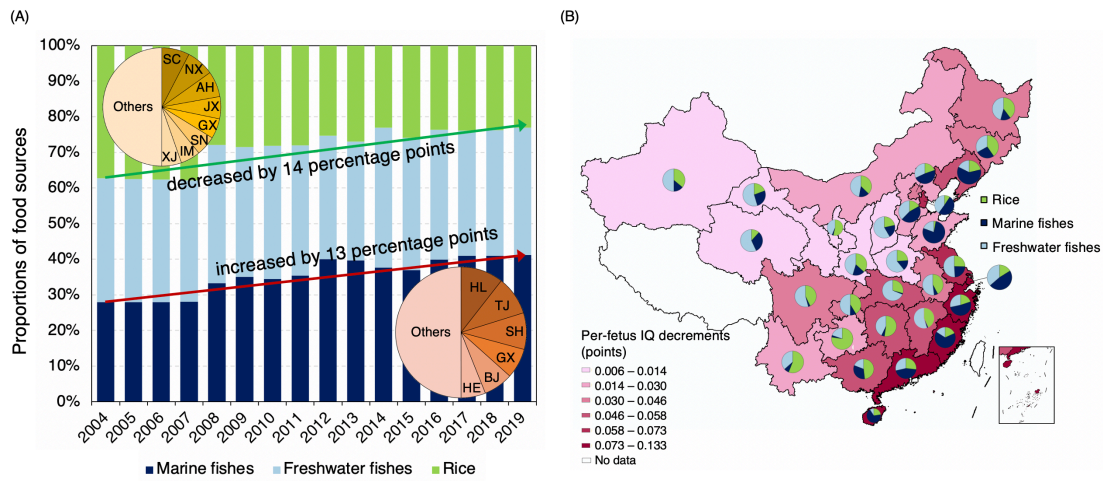
300 Fig. 2B shows **significant spatial heterogeneity of food sources** for the per-fetus
301 IQ decrements during 2004-2019. The per-fetus IQ decrements generally decreased
302 from the southeast coast to the northwest. Shanghai, Hainan, Fujian, Zhejiang, and

303 Guangdong have suffered fairly high per-fetus IQ decrements during 2004-2019,
304 followed by Tianjin, and Jiangsu. These provinces are located in coastal areas and their
305 per-fetus IQ decrements are mainly from the intake of **marine fishes** with high MeHg
306 concentrations, except for Jiangsu. Nearly 70% of the per-fetus IQ decrements in
307 Hainan and Fujian come from the intake of marine fishes. Other provinces with a high
308 proportion of per-fetus IQ decrements caused by marine fishes include Shandong and
309 Liaoning. These provinces are also located along the coast. Notably, about 78% of the
310 per-fetus IQ decrements in Shandong come from the intake of marine fishes. However,
311 the per-fetus IQ decrements in Shandong are much lower than those in southern coastal
312 provinces. This may result from the differences in intake levels and differences in
313 MeHg concentrations of marine fishes from different seas.

314 The provinces where per-fetus IQ decrements are dominated by **freshwater fish**
315 intake are mainly located in the Northwest (e.g., Xinjiang and Qinghai), Southwest (e.g.,
316 Sichuan and Chongqing), and middle reaches of the Yangtze River (e.g., Anhui and
317 Jiangxi). In particular, Hubei has the largest proportion of per-fetus IQ decrements from
318 freshwater fish intake, accounting for 69% among all food sources.

319 In addition to aquatic products, **rice**, as a staple food in southern China, also poses
320 significant health risk. For instance, per-fetus IQ decrements in Guizhou, Yunnan, and
321 Hunan mainly result from the intake of rice. The per-fetus IQ decrements from rice
322 account for 77%, 56%, and 52% of the total risk from all food sources in these three
323 provinces, respectively, during 2004-2019. Although MeHg concentrations of rice are

324 lower than those of fishes, high levels of rice intake have ultimately led to high per-
325 fetus IQ decrements in these southern inland areas.



326

327 **Figure 2. Food sources of per-fetus IQ decrements during 2004-2019.** Graphs
328 showing the temporal change in the proportion of food sources for per-fetus IQ
329 decrements (A) and spatial distribution of dominating food sources for per-fetus IQ
330 decrements (B). The pie charts in panel A show the major regions contributing to the
331 change in proportions of food sources during 2004-2019. Full names of the region
332 abbreviations are shown in Fig. S6. The colored background in panel B represents the
333 provincial population-weighted mean per-fetus IQ decrements during 2004-2019, and
334 pie charts represent the proportion of food sources for per-fetus IQ decrements in each
335 province.

336 **Geographical sources of per-fetus IQ decrements from dietary MeHg exposure.**

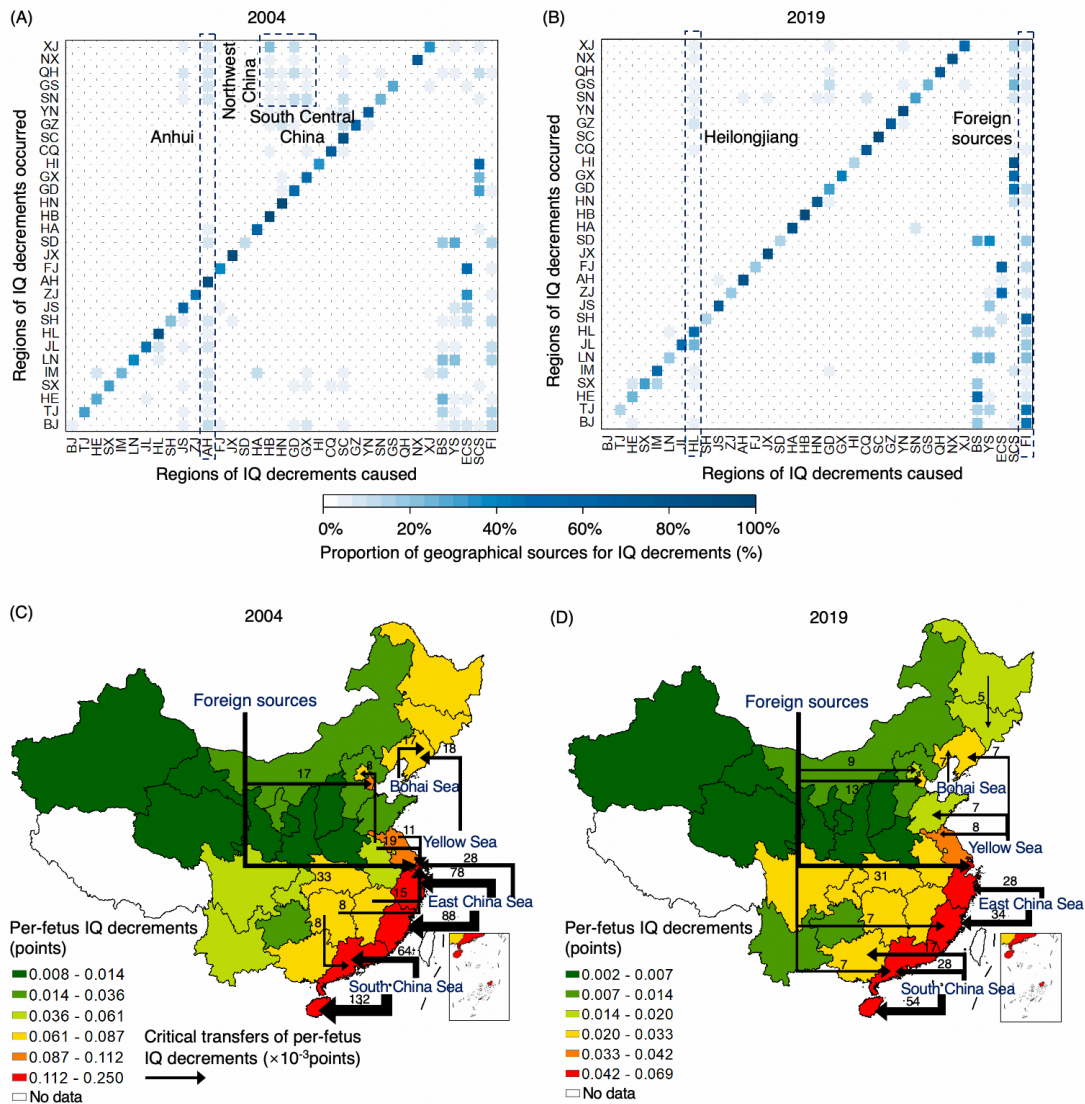
337 Fig. 3A-B show the geographical sources of per-fetus IQ decrements. Per-fetus IQ
338 decrements in many southern provinces (e.g., Jiangxi, Hubei, Hunan, and Anhui)
339 mainly result from the consumption of local foods, with local sources accounting for
340 over 80% during 2004-2019. Per-fetus IQ decrements occurred in these provinces are
341 dominated by the consumption of rice and freshwater fishes (Fig. 2A). In addition to
342 local food consumption, the adjacent sea areas and foreign regions are also major
343 sources of provincial per-fetus IQ decrements. For example, per-fetus IQ decrements

344 in Hainan, Guangxi, and Guangdong mainly come from the consumption of aquatic
345 products from the South China Sea. Per-fetus IQ decrements in Fujian and Zhejiang are
346 closely related to aquatic products from the East China Sea.

347 It is worth noting that in 2004, Anhui in East China is an important supplier of food
348 products (Fig. 3A). It poses per-fetus IQ decrements in most provinces of China,
349 especially in North (e.g., Beijing, Shanxi, and Inner Mongolia) and Northwest China
350 (e.g., Shaanxi, Gansu, Qinghai, Ningxia, and Xinjiang). In 2019, Heilongjiang in
351 Northeast China replaced Anhui to be an important geographical source of per-fetus IQ
352 decrements (Fig. 3B). The consumption of food products from Heilongjiang posed per-
353 fetus IQ decrements in Northwest China and Southwest China. In addition, foreign
354 sources have been becoming increasingly important in the contributions to China's per-
355 fetus IQ decrements during 2004-2019. For example, 58%, 45%, and 44% of the per-
356 fetus IQ decrements of Shanghai, Beijing, and Tianjin in 2019 are caused by the
357 consumption of foods from foreign regions.

358 **Interregional flows of per-fetus IQ decrements induced by food trade.** We further
359 identified major interregional flows of per-fetus IQ decrements caused by food trade in
360 2004 (Fig. 3C) and 2019 (Fig. 3D). During 2004-2019, the adjacent sea areas have been
361 major sources of per-fetus IQ decrements. For example, the largest flow of per-fetus IQ
362 decrements is the South China Sea–Hainan. The consumption of marine fishes from the
363 South China Sea has led to 0.132 and 0.054 points of per-fetus IQ decrements in Hainan
364 in 2004 and 2019, respectively.

365 The critical flows of per-fetus IQ decrements have been spatially shifting during
366 2004-2019. For example, Shanghai is a main receptor of per-fetus IQ decrements in
367 2004 (Fig. 3C). The IQ decrements occurred in Shanghai came not only from foreign
368 regions and adjacent seas, but also from the consumption of foods in Anhui (0.019
369 points), Jiangxi (0.015 points), Jiangsu (0.011 points), and Hunan (0.008 points).
370 However, in 2019, per-fetus IQ decrements in Shanghai are mainly from the
371 consumption of foods from foreign regions. Foreign regions have been becoming a
372 more and more important geographical source during 2004-2019. For instance, the
373 consumption of imported foods from foreign regions became an important geographical
374 source for per-fetus IQ decrements in Beijing, Fujian, and Guangdong in 2019. We
375 observed similar situations for the Yellow Sea. For example, the intake of marine fishes
376 from the Yellow Sea has been becoming an important source for per-fetus IQ
377 decrements in more and more regions (e.g., Shandong and Jiangsu) during 2004-2019.
378 This may be related to the increased MeHg levels in marine fishes from the Yellow Sea
379 (Supplementary Data 9).



381

382 **Figure 3. Geographical flows of per-fetus IQ decrements during 2004-2019.** Graphs
 383 **A, B** showing the proportion of per-fetus IQ decrements from geographical sources to
 384 each province in 2004 (**A**) and 2019 (**B**). The vertical axis represents the provinces
 385 where IQ decrements occur, and the horizontal axis represents the geographical sources
 386 of IQ decrements. Each cell represents the proportion of per-fetus IQ decrements in the
 387 province from different regions. Graphs **C, D** showing critical interregional flows of
 388 per-fetus IQ decrements ($\times 10^{-3}$ points) in 2004 (**C**) and 2019 (**D**). Full names and
 389 geographical locations of the region abbreviations are in the Supplementary
 390 Information ([Fig. S6](#)).

391 **Midpoint factors affecting China's per-fetus IQ decrements.** The per-fetus IQ
392 decrements of MeHg intake are closely associated with MeHg concentrations of foods
393 and dietary factors including dietary intake levels, dietary structure, and geographical
394 source structure (i.e., food trade structure). The per-fetus IQ decrements have generally
395 shown a declining trend during 2004-2019 (Fig. 1). Such decline results from the joint
396 effects of dietary shifts and the decrease of MeHg concentrations in foods. The dietary
397 shifts have contributed 44% of the decreased per-fetus IQ decrements during 2004-
398 2019, while the reductions in MeHg concentrations of foods contributed 56%.
399 Moreover, the contributions of these factors are spatially heterogeneous. For example,
400 for coastal regions, changes in dietary factors are the leading cause of the decline,
401 contributing 61% of the decreased per-fetus IQ decrements. In contrast, for northern
402 China, the reductions of MeHg concentrations take the leading role.

403 However, the declining trend has slowed down after 2014 and even leveled off after
404 2016. Per-fetus IQ decrements have decreased by 51% during 2004-2013, but only 22%
405 during 2014-2019. The average annual contribution of dietary pattern changes to
406 changes in per-fetus IQ decrements during 2014-2019 has decreased by 53% than that
407 during 2004-2013. In contrast, that for the decline in MeHg concentrations of foods has
408 decreased by only 12%. Thus, such slowdown and flattening are mainly caused by
409 changes in dietary patterns. This indicates that dietary pattern changes play an
410 important role in reducing per-fetus IQ decrements. Thus, optimizing dietary patterns
411 is critical to further reduce the per-fetus IQ decrements from dietary MeHg exposure.

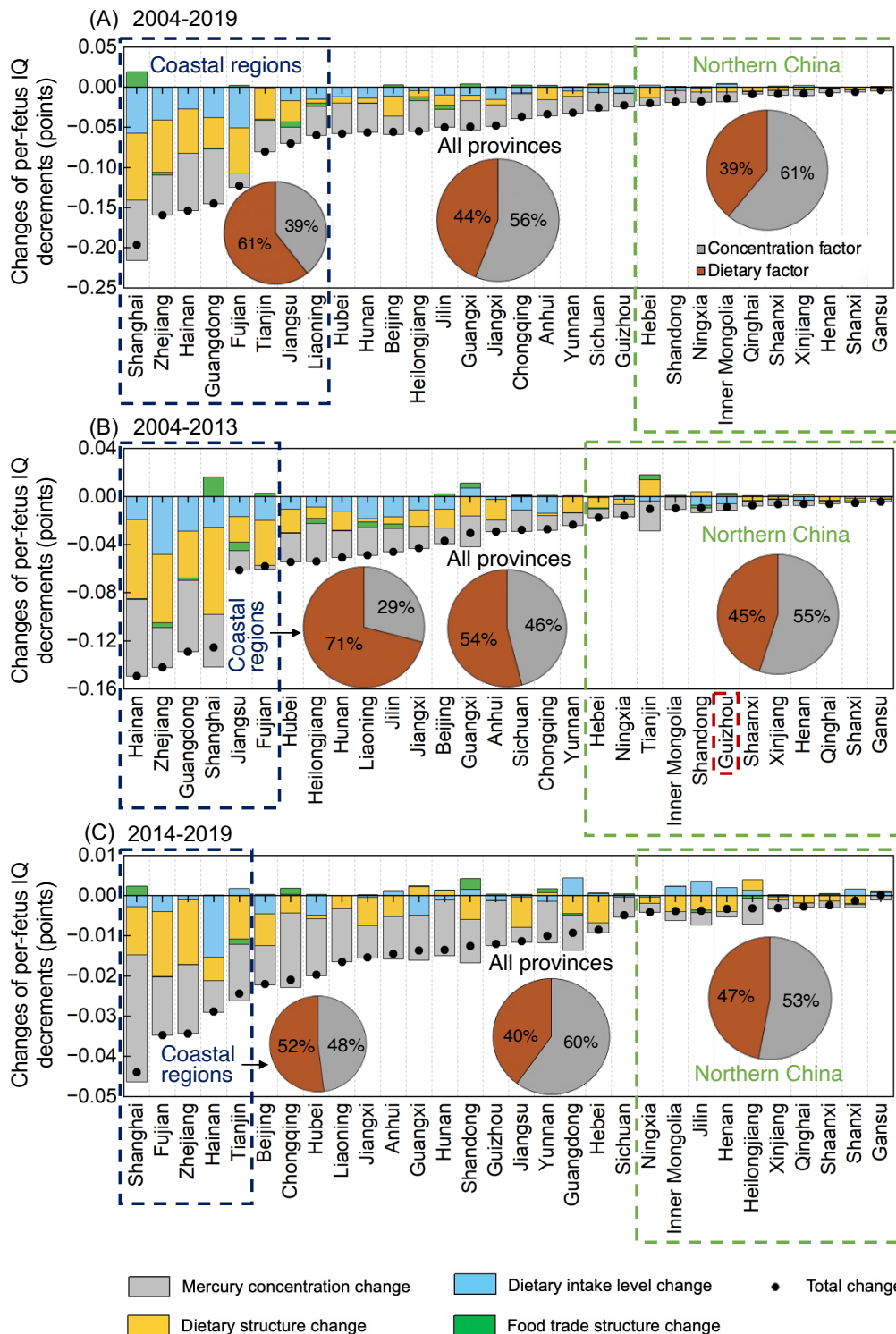
412 The following paragraphs specially focus on the time periods of 2004-2013 and 2014-
413 2019 (Fig. 4).

414 During 2004-2013, the decrease of China's per-fetus IQ decrements are mainly
415 dominated by coastal regions, especially Hainan, Zhejiang, Guangdong, and Shanghai
416 (Fig. 4B). This decline is mainly due to dietary pattern changes, with dietary factors
417 contributing 71% of the decreased risk in coastal regions. The northern regions,
418 especially Northwest (e.g., Gansu, Xinjiang, Qinghai, and Ningxia) and North China
419 (e.g., Shanxi, Tianjin, Inner Mongolia, Shandong, and Hebei), have relatively small
420 reductions of per-fetus IQ decrements. In general, the reduction of per-fetus IQ
421 decrements is mainly caused by changes in dietary intake levels and dietary structure.
422 The proportion of MeHg intake through marine fishes has increased in most provinces
423 due to relatively high MeHg concentrations in marine fishes (Fig. S8E). However, the
424 increased per-fetus IQ decrements are offset by a reduction in the proportion of MeHg
425 intake through rice and freshwater fishes. For example, Hainan has the largest decline
426 of per-fetus IQ decrements during 2004-2013. It is mainly due to the decreased
427 proportion of MeHg intake through rice and freshwater fishes.

428 During 2014-2019, China's per-fetus IQ decrements have decreased by 22% (Fig. 1).
429 The per-fetus IQ decrements have declined in most provinces, especially the coastal
430 regions such as Shanghai and Fujian (Fig. 4C). The decreased level and proportion of
431 MeHg intake through fishes are important factors for such decline (Fig. S8). Notably,
432 even though MeHg concentrations of rice are much lower than fishes, changes in the
433 intake levels of rice can still pose large impacts on per-fetus IQ decrements. For

434 example, changes in dietary intake levels of Hebei have contributed 74% of the
435 reduction in per-fetus IQ decrements. This is mainly due to the reduced proportion of
436 MeHg intake from rice (Fig. S8F).

437 Notably, changes in food trade structure of certain provinces have resulted in the
438 increase of per-fetus IQ decrements during 2014-2019. This increase was offset by the
439 decrease in per-fetus IQ decrements caused by changes in dietary intake levels and
440 dietary structure. Especially, changes in food trade structure of Shanghai have induced
441 the increase of per-fetus IQ decrements by 0.002 points. During 2014-2019, Shanghai
442 has significantly increased the consumption of marine fishes from foreign regions,
443 while reduced the consumption of marine fishes from South China Sea, Bohai Sea, and
444 Yellow Sea (Fig. S9A). This may explain the increased per-fetus IQ decrements caused
445 by the changes in food trade structure of Shanghai, because the MeHg concentrations
446 of marine fishes from foreign regions are much higher than those in China
447 (Supplementary Data 9). Moreover, changes in food trade structure of Chongqing have
448 contributed to the increase of per-fetus IQ decrements. During 2014-2019, Chongqing
449 has increased the consumption of marine fishes from foreign regions (Fig. S9B). In
450 addition, Chongqing has reduced the consumption of rice from the local, while
451 increased that from Heilongjiang. The rice from Heilongjiang has relatively higher
452 MeHg concentrations than that of Chongqing during 2014-2019 (Supplementary Data
453 9).



454

455 **Figure 4. Contributions of midpoint factors to changes in per-fetus IQ decrements**
 456 **in China.** Graphs **A, B, C** showing the relative contributions of MeHg concentrations
 457 and dietary factors to changes in per-fetus IQ decrements during 2004-2019 (**A**), 2004-
 458 2013 (**B**), and 2014-2019 (**C**). The pie charts in panels **A, B, C** represent the proportions
 459 for the contributions of MeHg concentrations and dietary factors to the reductions of
 460 per-fetus IQ decrements.

461 **Endpoint factors affecting China's per-fetus IQ decrements.** Dietary factors within
462 the food system are termed as midpoint factors in this study. They are further affected
463 by endpoint factors (i.e., affluence, education levels, and traffic accessibility in this
464 study) (da Costa et al., 2022; Galvan-Portillo et al., 2018; Mullie et al., 2010). We
465 further identified endpoint factors which are closely associated with China's per-fetus
466 IQ decrements. This could help to formulate socioeconomic policies that complement
467 diet-related control measures. We observe significant correlations between per-fetus IQ
468 decrements and endpoint factors including affluence level, education level, and traffic
469 accessibility (Table 1).

470 The affluence (represented by resident incomes) and education levels have
471 significantly negative associations with per-fetus IQ decrements. This is because people
472 with high levels of affluence and education have strong risk perception ability (Zobrist
473 et al., 2009). These groups pay more attention to heavy metal contamination in diets.
474 For example, they may be more likely to choose foods with low MeHg concentrations
475 or foods from regions with less Hg pollution. Meanwhile, people with higher levels of
476 affluence and education tend to consume healthier foods (e.g., vegetables, fruits, and
477 other plant-based foods) that induce less MeHg-related health risk (Lenthe et al., 2015).

478 The traffic accessibility (represented by road areas) has significantly negative
479 associations with per-fetus IQ decrements. It is probably because that good transport
480 conditions are conducive to interregional food trade (Ge et al., 2021). Foods with
481 relatively low MeHg concentrations can be more easily transported into regions with
482 higher traffic accessibility. Thus, people in these regions have more opportunities to

483 obtain foods with relatively low MeHg concentrations, and consequently suffer smaller
484 per-fetus IQ decrements.

485 **Table 1. The empirical relationships between China's per-fetus IQ decrements and**
486 **endpoint factors based on panel regression model.**

| Variables | Coefficients | Standard Errors | Probability |
|--|---------------------|------------------------|--------------------|
| <i>LnIncome</i> | -1.369*** | 0.304 | 0.000 |
| <i>LnEdu</i> | -0.319** | 0.126 | 0.017 |
| <i>LnRoad</i> | -0.547*** | 0.142 | 0.001 |
| R ² | 0.8 | | |
| Observations | 480 | | |
| Provincial level fixed effects | Yes | | |
| Hausman Test | 0.000 | | |

487 **Notes:** ** indicates the 5% statistical significance level and *** indicates the 1%
488 statistical significance level. The *LnIncome* represents the natural logarithm of
489 disposable income of residents; The *LnEdu* represents the natural logarithm of
490 education levels (i.e., the share of the population with educated college degree or above
491 in the total population above 6 years); and The *LnRoad* represents the natural logarithm
492 of road areas. The Hausman test is used to determine whether to take a random effects
493 regression or a fixed effects regression, with the original hypothesis that a random
494 effects model is fully efficient.

495 **DISCUSSION**

496 This study investigates the temporal change of China's per-fetus IQ decrements from
497 dietary MeHg exposure during 2004-2019 and reveals socioeconomic driving factors.
498 Our findings could provide the following implications for reducing MeHg-related

499 health risk in China, as well as new inspirations for other nations and regions around
500 the world.

501 **First, in addition to controlling Hg emissions, measures on health risk are**
502 **essential for the successful implementation of the Minamata Convention on**
503 **Mercury.** Many studies have analyzed Hg pollution control from the perspective of Hg
504 emission reduction. However, the ultimate goal of the Minamata Convention on
505 Mercury is to protect human beings from health threats caused by Hg pollution. Our
506 results show that some regions have low levels of Hg emissions but high per-fetus IQ
507 decrements. For example, the previous study shows that Hg emissions in Hainan are
508 far lower than Beijing (Liu et al., 2018a). However, this study finds that Hainan has
509 suffered per-fetus IQ decrements more than three times as much as Beijing during
510 2004-2019. That is, populations in low-emission regions may also suffer high health
511 risk from dietary MeHg exposure. This may be related to the differences in dietary
512 patterns between these two regions. According to the results of this study, the intake
513 level and ratio of marine fishes (with higher MeHg concentrations) in Hainan is much
514 higher than that in Beijing. Therefore, it is insufficient to implement Hg control
515 measures simply from the emission side. Investigating MeHg-related health risk and
516 underlying socioeconomic factors from the macro perspective is conducive to the better
517 implementation of the Minamata Convention on Mercury.

518 **Second, the increasing health risk from dietary MeHg exposure in**
519 **underdeveloped regions deserves more attention.** This study found that most
520 provinces with increased per-fetus IQ decrements after 2016 had relatively low per

521 capita GDP, including Gansu, Shaanxi, and Anhui. Gansu and Shaanxi have increased
522 their intake of marine fishes, while Anhui has increased the intake of freshwater fishes.,
523 With the rapid economic development in the future, these regions will continue to face
524 diet transitions. Relatively high MeHg concentrations of fishes may increase MeHg-
525 related health risk. Minor changes in fish intake structures can lead to significant
526 changes in health risk. It is important to establish strict food safety standards for fishes
527 and incorporate them into dietary guidelines. Moreover, guiding consumers to choose
528 marine fishes in areas with relatively low MeHg concentrations (e.g., Bohai Sea and
529 Yellow Sea) can help control MeHg-related health risk.

530 **Third, the role of rice intake in reducing MeHg-related health risk must be**
531 **concerned.** Existing studies find that the intake of rice and fishes is the primary
532 pathway of dietary MeHg exposure (Sunderland et al., 2018; Zhang et al., 2010). Rice
533 is a staple food in many regions of China and is consumed in large quantities. Therefore,
534 although the MeHg concentrations of rice are much lower than fishes, rice intake plays
535 an important role in the changes of health risk from dietary MeHg intake. For example,
536 the large decline of per-fetus IQ decrements in Hainan, Guangdong, and Zhejiang
537 during 2004-2013 was due to the reduction in the proportion of MeHg intake through
538 rice consumption. Therefore, regions dominated by rice (e.g., Guizhou, Hunan, and
539 Yunnan) can be encouraged to partly choose alternative staple foods with low MeHg
540 concentrations (e.g., wheat, corn, and potatoes), which could also help reduce
541 environmental impacts (Liu et al., 2021). Moreover, promoting the transformation of
542 dietary patterns to plant-based diets (e.g., vegetables, fruits, and nuts) can help offset

543 the increasing trend of MeHg-related health risk. This effort is also in line with the
544 planetary health diet proposed by the ETA Lancet Commission (Willett et al., 2019).

545 **Fourth, strengthening the guidance on consumption patterns is an effective way**
546 **to reduce health risk from dietary MeHg exposure.** The affluence and education
547 levels have significantly negative correlation with per-fetus IQ decrements from dietary
548 MeHg exposure. This is because that rich and well-educated people have higher ability
549 of risk perception. They are more likely to choose foods with low MeHg concentrations.
550 This further indicates the important role of consumption patterns. There is a need to
551 strengthen the guidance of rational consumption patterns. For example, in rich and
552 highly educated areas, foods could be labelled with MeHg concentrations. In this way,
553 consumers are encouraged to choose foods with relatively low MeHg concentrations.
554 Furthermore, guiding the advocacy for food safety in underdeveloped regions and
555 regions with low educational levels is conducive to controlling national MeHg-related
556 health risk.

557 **Finally, other nations around the world can refer to the findings of this study in**
558 **developing socioeconomic regulation policies.** Developing nations with severe Hg
559 pollution issues (e.g., India and Colombia) can refer to our findings. This study finds
560 that the most important factor influencing per-fetus IQ decrements in coastal regions is
561 the consumption of marine fishes. Thus, it is important for coastal regions of the world
562 to monitor the MeHg concentrations of seafoods and choose reliable sources of marine
563 fishes. Furthermore, international and intranational food trades can be strengthened.

564 This action could help reduce the dependence on foods from certain regions with
565 relatively high MeHg concentrations.

566 **Conclusions**

567 China plays an important and representative role in the implementation of the
568 Minamata Convention on Mercury which aims to protect human beings from adverse
569 impacts of mercury and its compounds. Chinese population has suffered severe health
570 risk from dietary MeHg exposure. However, nothing is known about the historical trend
571 of China's health risk from dietary MeHg exposure and relevant socioeconomic driving
572 factors. This study fulfills these knowledge gaps and investigates the historical trends
573 of China's MeHg-related health risk during 2004-2019. Moreover, this study identifies
574 critical socioeconomic factors influencing changes in health risk.

575 We find that China's per-fetus IQ decrements have decreased by 60% during this
576 period. This decline results from the joint effects of dietary shifts and the decrease of
577 MeHg concentrations in foods. However, the declining trend has slowed down after
578 2014 and even leveled off after 2016. Such slowdown and flattening are mainly caused
579 by dietary pattern changes. For example, the increased intake level and proportion of
580 fishes in underdeveloped provinces of China have dominated the slowdown of
581 declining trend after 2016. Moreover, the affluence and education levels have
582 significantly negative associations with per-fetus IQ decrements. Rich and well-
583 educated people have higher ability of risk perception, which indicates the importance
584 of rational consumption patterns.

585 The findings can help develop socioeconomic regulatory policies on reducing per-
586 fetus IQ decrements from dietary MeHg exposure in China. Relevant policy
587 implications include focusing on health risk alleviation in addition to Hg emission
588 control, guiding consumers to choose marine fishes in areas with relatively low MeHg
589 concentrations, establishing strict food safety standards for fishes, encouraging regions
590 dominated by rice to partly choose alternative staple foods with low MeHg
591 concentrations, as well as strengthening controls on international food trade.

592 **Author contributions**

593 S.L. designed this study. Y.L., H.Z., and X.W. collected the data. Y.L. and Q.Z.
594 conducted the calculations and interpretations of the results. S.L., Y.L., H.Z., X.W., P.H.,
595 Q.Z., and L.C. wrote and revised the paper. S.L. supervised this study.

596 **Declaration of interests**

597 The authors declare no competing interests.

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602 **References**

603 Axelrad, D.A., Bellinger, D.C., Ryan, L.M., Woodruff, T.J., 2007. Dose-response
604 relationship of prenatal mercury exposure and IQ: an integrative analysis of
605 epidemiologic data. *Environ. Health Perspect.* 115, 609-615.
606 Bose-O'Reilly, S., Schierl, R., Nowak, D., Siebert, U., William, J.F., Owi, F.T., Ir, Y.I.,
607 2016. A preliminary study on health effects in villagers exposed to mercury in a

608 small-scale artisanal gold mining area in Indonesia. *Environ. Res.* 149, 274-281.
609 Chen, L., Liang, S., Liu, M., Yi, Y., Mi, Z., Zhang, Y., Li, Y., Qi, J., Meng, J., Tang, X.,
610 Zhang, H., Tong, Y., Zhang, W., Wang, X., Shu, J., Yang, Z., 2019. Trans-
611 provincial health impacts of atmospheric mercury emissions in China. *Nat.*
612 *Commun.* 10, 1484.

613 Cheng, C., Ren, X.H., Dong, K.Y., Dong, X.C., Wang, Z., 2021. How does
614 technological innovation mitigate CO₂ emissions in OECD countries?
615 Heterogeneous analysis using panel quantile regression. *J. Environ. Manage.* 280.
616 da Costa, G.G., Nepomuceno, G.D., Pereira, A.D., Simoes, B.F.T., 2022. Worldwide
617 dietary patterns and their association with socioeconomic data: an ecological
618 exploratory study. *Globalization and Health* 18.

619 Deng, C.X., Zhang, G.J., Li, Z.W., Li, K., 2020. Interprovincial food trade and water
620 resources conservation in China. *Science of the Total Environment* 737.

621 Dietzenbacher, E., Los, B., 1998. Structural Decomposition Techniques: Sense and
622 Sensitivity. *Econ. Syst. Res.* 10, 307-324.

623 Driscoll, C.T., Mason, R.P., Chan, H.M., Jacob, D.J., Pirrone, N., 2013. Mercury as a
624 global pollutant: sources, pathways, and effects. *Environ. Sci. Technol.* 47, 4967-
625 4983.

626 Galvan-Portillo, M., Sánchez, E., Cárdenas-Cárdenas, L.M., Karam, R., Claudio, L.,
627 Cruz, M., Burguete-García, A.I., 2018. Dietary patterns in Mexican children and
628 adolescents: Characterization and relation with socioeconomic and home
629 environment factors. *Appetite* 121, 275-284.

630 Gao, Y.-X., Zhang, H., Yu, X., He, J.-l., Shang, X., Li, X., Zhao, Y., Wu, Y., 2014. Risk
631 and Benefit Assessment of Potential Neurodevelopmental Effect Resulting from
632 Consumption of Marine Fish from a Coastal Archipelago in China. *Journal of*
633 *Agricultural and Food Chemistry* 62, 5207-5213.

634 Ge, F.J., Chen, W.X., Zeng, Y.Y., Li, J.F., 2021. The Nexus between Urbanization and
635 Traffic Accessibility in the Middle Reaches of the Yangtze River Urban
636 Agglomerations, China. *International journal of environmental research and*
637 *public health* 18.

638 Giang, A., Selin, N.E., 2016. Benefits of mercury controls for the United States. *Proc.*
639 *Natl. Acad. Sci. U. S. A.* 113, 286-291.

640 Grandjean, P., Pichery, C., Bellanger, M., Budtzjorgensen, E., 2012. Calculation of
641 mercury's effects on neurodevelopment. *Environ. Health Perspect.* 120, 452.

642 Guan, D., Hubacek, K., Weber, C.L., Peters, G.P., Reiner, D., 2008. The drivers of
643 Chinese CO₂ emissions from 1980 to 2030. *Global Environ Chang* 18, 626-634.

644 He, P., Baiocchi, G., Feng, K.S., Hubacek, K., Yu, Y., 2019. Environmental impacts of
645 dietary quality improvement in China. *Journal of Environmental Management* 240,
646 518-526.

647 Henriques, M.C., Loureiro, S., Fardilha, M., Herdeiro, M.T., 2019. Exposure to
648 mercury and human reproductive health: A systematic review. *Reprod. Toxicol.* 85,
649 93-103.

650 Hoekstra, R., Bergh, J.C.J.M.v.d., 2002. Structural Decomposition Analysis of Physical
651 Flows in the Economy. *Environ. Resour. Econ.* 23, 357-378.

652 Hong, C., Yu, X., Liu, J., Cheng, Y., Rothenberg, S.E., 2016. Low-level methylmercury
653 exposure through rice ingestion in a cohort of pregnant mothers in rural China.
654 *Environ Res* 150, 519-527.

655 Hu, X.F., Laird, B.D., Chan, H.M., 2017. Mercury diminishes the cardiovascular
656 protective effect of omega-3 polyunsaturated fatty acids in the modern diet of Inuit
657 in Canada. *Environ. Res.* 152, 470-477.

658 Lenthe, F.V., Jansen, T., Kamphuis, C.B.M., 2015. Understanding socio-economic
659 inequalities in food choice behaviour: can Maslow's pyramid help? *Br. J. Nutr.* 113,
660 1139-1147.

661 Lenzen, M., Moran, D., Kanemoto, K., Foran, B., Lobefaro, L., Geschke, A., 2012.
662 International trade drives biodiversity threats in developing nations. *Nature* 486,
663 109-112.

664 Li, J., Zhou, S., Wei, W., Qi, J., Li, Y., Chen, B., Zhang, N., Guan, D., Qian, H., Wu, X.,
665 Miao, J., Chen, L., Feng, K., Liang, S., 2020a. China's retrofitting measures in
666 coal-fired power plants bring significant mercury-related health benefits. *One*
667 *Earth* 3, 777-787.

668 Li, P., Feng, X., Qiu, G., 2010. Methylmercury exposure and health effects from rice
669 and fish consumption: a review. *Int. J. Environ. Res. Public Health* 7, 2666-2691.

670 Li, P., Feng, X., Yuan, X., Chan, H.M., Qiu, G., Sun, G., Zhu, Y., 2012. Rice
671 consumption contributes to low level methylmercury exposure in southern China.
672 *Environ. Int.* 49, 18-23.

673 Li, Y., Chen, L., Liang, S., Qi, J., Zhou, H., Feng, C., Yang, X., Wu, X., Mi, Z., Yang,
674 Z., 2020b. Spatially Explicit Global Hotspots Driving China's Mercury Related
675 Health Impacts. *Environ. Sci. Technol.* 54, 14547-14557.

676 Liang, S., Xu, M., Liu, Z., Suh, S., Zhang, T., 2013. Socioeconomic drivers of mercury
677 emissions in China from 1992 to 2007. *Environ. Sci. Technol.* 47, 3234-3240.

678 Liang, S., Zhang, C., Wang, Y., Xu, M., Liu, W., 2014. Virtual atmospheric mercury
679 emission network in China. *Environ. Sci. Technol.* 48, 2807-2815.

680 Liu, B.B., Gu, W.Y., Yang, Y., Lu, B.F., Wang, F., Zhang, B., Bi, J., 2021. Promoting
681 potato as staple food can reduce the carbon-land-water impacts of crops in China.
682 *Nat. Food* 2, 570-577.

683 Liu, K., Wang, S., Wu, Q., Wang, L., Ma, Q., Zhang, L., Li, G., Tian, H., Duan, L., Hao,
684 J., 2018a. A highly resolved mercury emission inventory of Chinese coal-fired
685 power plants. *Environ. Sci. Technol.* 52, 2400-2408.

686 Liu, M., Chen, L., He, Y., Baumann, Z., Mason, R.P., Shen, H., Yu, C., Zhang, W.,
687 Zhang, Q., Wang, X., 2018b. Impacts of farmed fish consumption and food trade
688 on methylmercury exposure in China. *Environ. Int.* 120, 333-344.

689 Liu, M., Cheng, M., Zhang, Q., Hansen, G., He, Y., Yu, C., Lin, H., Zhang, H., Wang,
690 X., 2020. Significant elevation of human methylmercury exposure induced by the
691 food trade in Beijing, a developing megacity. *Environ. Int.* 135, 105392.

692 Mahaffey, K.R., Clickner, R.P., Bodurow, C.C., 2004. Blood organic mercury and
693 dietary mercury intake: National Health and Nutrition Examination Survey, 1999
694 and 2000. *Environ. Health Perspect.* 112, 562-570.

695 Mullie, P., Clarys, P., Hulens, M., Vansant, G., 2010. Dietary patterns and

696 socioeconomic position. *European Journal of Clinical Nutrition* 64, 231-238.

697 Qing, Y., Li, Y.Z., Yang, J.Q., Li, S.C., Gu, K.X., Bao, Y.X., Zhan, Y.H., He, K., Wang,
698 X.Y., Li, Y.F., 2022. Risk assessment of mercury through dietary exposure in
699 China*. *Environmental Pollution* 312.

700 Rice, G.E., Hammitt, J.K., Evans, J.S., 2010. A Probabilistic Characterization of the
701 Health Benefits of Reducing Methyl Mercury Intake in the United States. *Environ.*
702 *Sci. Technol.* 44, 5216-5224.

703 Roman, H.A., Walsh, T.L., Coull, B.A., Dewailly, É., Guallar, E., Hattis, D., Mariën,
704 K., Schwartz, J., Stern, A.H., Virtanen, J.K., Rice, G., 2011. Evaluation of the
705 cardiovascular effects of methylmercury exposures: current evidence supports
706 development of a dose-response function for regulatory benefits analysis. *Environ.*
707 *Health Perspect.* 119, 607-614.

708 Rørnøse, P., Olsen, T., 2005. Structural Decomposition Analysis of Air Emissions in
709 Denmark 1980–2002., 15th International Conference on Input-Output Techniques,
710 Beijing, China.

711 Selin, N.E., Sunderland, E.M., Knightes, C.D., Mason, R.P., 2010. Sources of mercury
712 exposure for U.S. seafood consumers: implications for policy. *Environ. Health*
713 *Perspect.* 118, 137-143.

714 Stuart, W.B., Grace, L.A., Grala, R.K., 2010. Returns to scale in the Eastern United
715 States logging industry. *Forest Policy Econ.* 12, 451-456.

716 Sunderland, E.M., Li, M., Bullard, K., 2018. Decadal Changes in the Edible Supply of
717 Seafood and Methylmercury Exposure in the United States. *Environmental Health*
718 *Perspectives* 126, 017006.

719 Sundseth, K., Pacyna, J.M., Pacyna, E.G., Pirrone, N., Thorne, R.J., 2017. Global
720 Sources and Pathways of Mercury in the Context of Human Health. *Int. J Environ.*
721 *Res. Public Health* 14.

722 United Nations Environment Programme, 2013. Minamata Convention on Mercury.
723 Minamata, Japan: United Nations Environment Programme (2013).

724 United Nations Environment Programme, 2019. Global Mercury Assessment 2018. UN
725 Environment Programme, Chemicals and Health Branch Geneva, Switzerland.

726 Willett, W., Rockström, J., Loken, B., Springmann, M., Lang, T., Vermeulen, S., Garnett,
727 T., Tilman, D., DeClerck, F., Wood, A., Jonell, M., Clark, M., Gordon, L.J., Fanzo,
728 J., Hawkes, C., Zurayk, R., Rivera, J.A., De Vries, W., Majele Sibanda, L., Afshin,
729 A., Chaudhary, A., Herrero, M., Agustina, R., Branca, F., Lartey, A., Fan, S., Crona,
730 B., Fox, E., Bignet, V., Troell, M., Lindahl, T., Singh, S., Cornell, S.E., Srinath
731 Reddy, K., Narain, S., Nishtar, S., Murray, C.J.L., 2019. Food in the Anthropocene:
732 the EAT–Lancet Commission on healthy diets from sustainable food systems. *The*
733 *Lancet* 393, 447-492.

734 World Health Organization, 2017. Fact sheets: Mercury and health.
735 <https://www.who.int/en/news-room/fact-sheets/detail/mercury-and-health>.

736 Zhang, H., Feng, X., Larssen, T., Qiu, G., Vogt, R.D., 2010. In inland China, rice, rather
737 than fish, is the major pathway for methylmercury exposure. *Environ. Health*
738 *Perspect.* 118, 1183-1188.

739 Zobrist, J., Sima, M., Dogaru, D., Senila, M., Yang, H., Popescu, C., Roman, C., Bela,

740 A., Frei, L., Dold, B., Balteanu, D., 2009. Environmental and socioeconomic
741 assessment of impacts by mining activities-a case study in the Certej River
742 catchment, Western Carpathians, Romania. Environ. Sci. Pollut. Res. 16, 14-26.
743