

This is an Open Access document downloaded from ORCA, Cardiff University's institutional repository: <https://orca.cardiff.ac.uk/id/eprint/137150/>

This is the author's version of a work that was submitted to / accepted for publication.

Citation for final published version:

Wu, Hao-liang, Jin, Fei , Zhou, An-nan and Du, Yan-Jun 2021. The engineering properties and reaction mechanism of MgO-activated slag cement-clayey sand-bentonite (MSB) cutoff wall backfills. Construction and Building Materials 271 , 121890. 10.1016/j.conbuildmat.2020.121890

Publishers page: <http://dx.doi.org/10.1016/j.conbuildmat.2020.12189...>

Please note:

Changes made as a result of publishing processes such as copy-editing, formatting and page numbers may not be reflected in this version. For the definitive version of this publication, please refer to the published source. You are advised to consult the publisher's version if you wish to cite this paper.

This version is being made available in accordance with publisher policies. See <http://orca.cf.ac.uk/policies.html> for usage policies. Copyright and moral rights for publications made available in ORCA are retained by the copyright holders.



**The engineering properties and reaction mechanism of MgO-activated slag
cement-clayey sand-bentonite (MSB) cutoff wall backfills**

Hao-Liang Wu^{1,2}, Fei Jin^{3,4}, An-nan Zhou⁵, Yan-Jun Du¹

¹Jiangsu Key Laboratory of Urban Underground Engineering & Environmental Safety,
Institute of Geotechnical Engineering, Southeast University, Nanjing 210096, China.

²Department of Civil and Environmental Engineering, Hong Kong University of Science and
Technology, Hong Kong SAR, China.

³Formerly at School of Engineering, University of Glasgow, Glasgow G12 8QQ, UK.

⁴Currently at School of Engineering, Cardiff University, Cardiff, CF10 3AT, UK.

⁵Civil and Infrastructure Engineering, School of Engineering, Royal Melbourne Institute of
Technology (RMIT), Melbourne, Vic 3000, Australia.

*Corresponding authors, Email: jinf2@cardiff.ac.uk and duyanjun@seu.edu.cn

Abstract

An innovative cutoff wall backfill consisting of reactive MgO, ground granulated blast furnace slag (GGBS), bentonite and local clayey sand (MSB) was developed recently for land remediation applications. This paper investigates the engineering characteristics (e.g. strength and permeability) and reaction mechanisms of the MSB backfills with different MgO-activated GGBS (i.e., the binder) and bentonite contents. A series of analytical techniques are employed to identify the hydration products in this complex system. The engineering properties are tested via the unconfined compressive strength (UCS) test and flexible-wall permeation test in a triaxial cell. Results show that UCS and dry density decrease with the increasing bentonite content, while the opposite trends are observed when increasing the binder content. The UCS values increase with curing time and become plateaued after ~90 days. Meanwhile, the hydraulic conductivity (k) decreases distinctly with the increase of the binder content and bentonite content. All backfills reach UCS of >100 kPa UCS and $k < 10^{-8}$ m/s at 28 days while curing for 90 days leads to increase of UCS by >1.5 times and reduction of k by nearly one order of magnitude. The major hydration products of MSB backfills are identified as hydrotalcite-like phases (Ht), calcium silicate hydrates (C-S-H), monosulfate (AFm) and portlandite (CH). The hydration products, binding adjacent soil particles, filling pores, together with the swelling of bentonite, contribute to the mechanical performance and impermeability of the backfills.

Keywords: MgO-activated GGBS; cutoff wall; hydration products; engineering properties; reaction mechanisms

1. Introduction

Ground granulated blast furnace slag (GGBS), an industrial by-product of iron production, has been extensively adopted as a replacement of ordinary Portland cement (OPC) (Wang, and Scrivener, 1995). It is known that GGBS imparts significant improvement on the cement and concrete properties including lower heat generation, higher ultimate strength, notably lower hydraulic conductivity, reduced alkalinity, better durability with additional environmental and economic benefits. This has led to the ever-increasing popularity of GGBS in various geotechnical/geoenvironmental applications in recent years (Lam and Jefferis, 2017; Wu et al. 2018a and 2019a; Zheng, et al., 2019; Park et al., 2020). For cutoff walls, partial replacement of OPC with GGBS is employed in several remediation projects (Opdyke and Evans, 2005; Joshi et al., 2009; Li et al., 2020). Findings show that over 60 - 80% OPC could be replaced by GGBS while providing increased strength, reduced bleed and hydraulic conductivity, and higher resistance to sulfate attack to the backfills (Garvin and Hayles, 1999).

Research efforts have been made to understand the activation mechanism and hydration products of GGBS in the presence of various alkalis (Ben Haha et al., 2011; Schöler et al, 2015). The activation process of GGBS begins with the destruction of its chemical bonds (i.e., Si-O, Al-O, and Si-O-Al), followed by the formation of the hydration products such as calcium silicate hydrates (C-S-H) and hydrotalcite ($\text{Mg}_6\text{Al}_2(\text{OH})_{16}\text{CO}_3 \cdot 4\text{H}_2\text{O}$)-like phases (Ht) via dynamic chemical equilibria and diffusion of reactive species (Krizan 2002; Bernal et al., 2011; Yi et al., 2013 and 2016). Various alkaline earth metal oxides (CaO and MgO) and hydroxides (i.e., $\text{Ca}(\text{OH})_2$, $\text{Sr}(\text{OH})_2$, and $\text{Ba}(\text{OH})_2$) have been found to facilitate the activation of slags (Song, 2000). Particularly, reactive MgO is an emerged novel activator, offering many potential

68 geomechanical and geoenvironmental benefits over conventional activators (Jin et al. 2015;
69 Du et al. 2016; Wang et al. 2016; Yi et al., 2016). Research reveals that the main hydration
70 products of MgO activated GGBS are C-S-H gel and Ht, resulting in a denser microstructure,
71 improved mechanical performance, as well as excellent pH-buffering and adsorptive capacities,
72 compared to OPC-based binders in heavy metals immobilization (Jin and Al-Tabbaa, 2014a, b
73 and c; Du et al. 2016; Li et al., 2020; Jin and Al-Tabbaa, 2020). Therefore, it is envisaged that
74 MgO-activated GGBS as a binder in cutoff wall applications could provide a significant
75 improvement on the engineering properties with improved long-term durability in land
76 remediation projects.

77
78 In previous studies (Wu et al., 2019a and b), laboratory investigations focused on the
79 engineering properties and wetting-drying durability of the novel backfill consisting of MgO-
80 activated GGBS (i.e., the binder), bentonite, and local clayey sand (MSB). Results show that
81 MSB with different mix compositions at 90 days could achieve unconfined compressive
82 strength (UCS) values of 230 - 520 kPa and hydraulic conductivity (k) in clean tap water of
83 1.1×10^{-10} - 6.3×10^{-10} m/s. It is noted that the k values of MSB backfill using sodium sulfate (30
84 mmol/L) and lead-zinc (Pb = 0.1 mg/L, Zn = 5 mg/L) solutions as permeant liquids are 1-2
85 orders of magnitudes lower than those of OPC-based backfills (Wu et al. 2019a). Regarding
86 durability, MSB is marginally more sensitive to a fluctuating groundwater condition (i.e., wet-
87 dry cycles) than OPC-based backfills (Wu et al. 2019b). The UCS values of the MSB backfills
88 decline from 240-560 kPa to 180-420 kPa after 10 wet-dry cycles, though it is still above the
89 commonly adopted design limit (≥ 100 kPa). Despite these promising results, the detailed
90 reaction mechanisms and microstructural characteristics of the MSB mixtures and hence their

potential responses to the varying service conditions remain unaddressed. The reaction mechanism of MgO and GGBS blends has been investigated recently (Jin et al., 2015; Park et al., 2020), whereas the interactions between the MgO-activated GGBS, bentonite, and soil remain unresolved. It was reported that the backfill pH and saline cations affected the chemical stability of montmorillonite in bentonite (Jefferis 2012; Du et al. 2020). For example, a high concentration of Ca^{2+} could adversely impact the swelling properties of bentonite and then reduce the hydraulic conductivity (Wu et al. 2019a). Therefore, further studies to reveal the multiscale characteristics of MSB backfills and their interactions with the complex exposure environment in land remediation projects are warranted.

This paper investigates the evolution of engineering properties, including UCS and hydraulic conductivity, with curing time and different mix compositions of MSB backfills. The geotechnical properties including dry density and void ratio are also reported. Reaction mechanisms and microstructural characteristics of the MSB backfills are revealed via X-ray diffraction (XRD), scanning electron microscopy (SEM), Fourier-transform infrared spectroscopy (FTIR), and Mercury Intrusion Porosimetry (MIP). The hydration products of MSB backfills that consist of a local clayey sand, various dosages of binder and sodium activated calcium-bentonite are probed. The relationship between the engineering properties, hydration products and microstructure of MSB backfills is then discussed.

2. Materials and testing program

2.1 Material

The raw materials used in this study include a local clayey sand, MgO, GGBS and sodium

activated calcium-bentonite. The local clayey sand is classified as per ASTM D2487 (2017a) and was collected from the floodplain in Nanjing, China. The basic physicochemical properties of the soil and the chemical compositions of the raw materials measured by X-ray fluorescence can be found in Wu et al., (2019a).

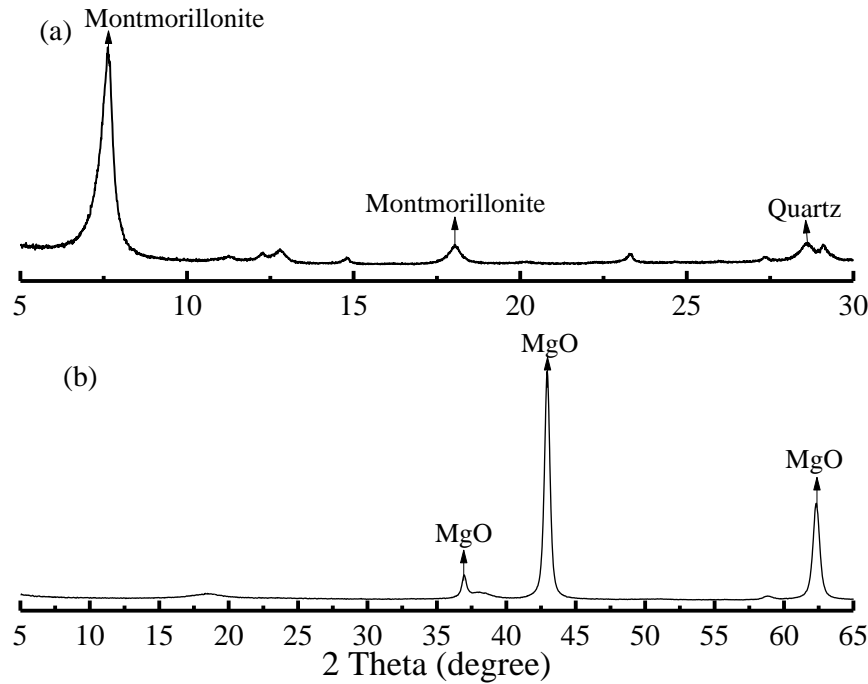


Fig. 1 X-ray diffractograms patterns for the sodium activated calcium-bentonite (a) and MgO (b)

The commercial powdered sodium activated calcium-bentonite was provided by MuFeng mineral processing plant in Zhenjiang, China. Its specific surface area (SSA) is 378.5 m²/g, as measured by the nitrogen adsorption method. The MgO (from Jiangsu-Onada Cement Corp.) blended with GGBS (from Nanjing Iron and Steel Group) was used as the binder. The SSA and chemical reactivity of the MgO is 28 m²/g and 102 s (using the acetic acid test according to Shand (2006)), respectively. It can be categorized as the medium reactivity MgO (Jin and Al-Tabbaa, 2015). The X-ray diffractograms for the sodium activated calcium-bentonite and MgO is shown in **Fig. 1**, which shows the major crystalline phases are montmorillonite and magnesium oxide, respectively. The XRD spectra were obtained using RigakuD/Max-2500

spectrometer using a Cu-K α source with a wavelength of 1.5405 Å. The instrument was operated at 40 kV and 20 mA. A step size of $2\theta = 0.02^\circ$ and a scanning speed of 5 s/step were used.

2.2 Specimens

The dosage of the binder and bentonite is 5% - 10% and 5% - 15% by weight of the soil, respectively. The weight ratio of GGBS: MgO in the binder is fixed at 9:1 in this study, which renders adequate strength and lowest hydraulic conductivity as indicated in previous studies (Jin et al. 2015; Wu et al. 2018a and. 2019a). Table 1 shows the mix proportions of the studied backfills, which are part of the mixes in (Wu et al., (2019a). The raw materials, including the soil, bentonite and MgO-activated GGBS were weighted first according to the mix design. A 2-L stainless steel mixer was used to blend the dry raw materials at 30 rpm for 5 min. A predetermined amount (to achieve the slump value of 150 ± 5 mm for each mixture according to a preliminary study) of tap water (pH = 6.8; $EC = 3.3 \mu\text{S}/\text{cm}$) was poured into the dry mixture and mixed at 60 rpm for 10 min to achieve homogeneous slurries. The fresh slurries were then cast into cylindrical plastic moulds ($\Phi 50 \times H100$ mm) to cure for 14-120 days at the temperature of $20 \pm 2^\circ\text{C}$ and relative humidity of $95 \pm 1\%$ in sealed plastic boxes until testing.

Table 1. Mix design of the backfills used in this study (wt%)

Category ID	Raw materials (g)			Slump ^a	Water	Density	Water	Dry	Void
				(mm)	content ^b , w	of soil	to	density ^c ,	ratio ^d ,
					(%)	particle,	binder	ρ_d	(e)
	Soil	Bentonite	Binder (MgO+GGBS)			ρ_s	ratio	(g/cm ³)	
						(g/cm ³)			
MS5B5	1000	50	50	152	23.9	2.64	5.26	1.42	0.85
MS5B10	1000	100	50	154	33.6	2.65	7.73	1.36	0.93
MS5B15	1000	150	50	152	34.7	2.66	8.33	1.36	0.95
MS10B10	1000	100	100	153	35.4	2.65	4.25	1.37	0.94

MS10B15	1000	150	100	155	36.5	2.66	4.56	1.36	0.94
---------	------	-----	-----	-----	------	------	------	------	------

^a The slump was measured as per [ASTM D6103](#)

^b The moisture was measured as per [ASTM D2216](#)

^c The dry density was measured as per [ASTM D7263](#)

^d Void ratio (e) was determined by $e = \rho_s / \rho_d - 1$, where ρ_s is the soil particle density of backfills and calculated from the raw materials, w is the water content, and ρ_d is the dry density of backfills.

2.3 Testing procedure and instruments

UCS was measured in triplicate with a vertical load applied at a constant strain rate of 1%/min according to ASTM D4219 ([ASTM 2008](#)) after 14, 28, 60, 90 and 120 days of curing. Dry density test was conducted in triplicate as per ASTM D7263 ([ASTM 2018b](#)) after each UCS test. The hydraulic conductivity was measured with a flexible-wall permeameter at 22 ± 2 °C as detailed in [Wu et al., \(2019a\)](#) for each mixture cured after 28 and 90 days. MS10B10 backfill specimens were prepared with freeze-drying methods (dried at -80 °C) for 24 hours for micro-analysis after curing for 90 days. Prior to XRD and FTIR tests, the dry specimens were ground and sieved to < 0.075 mm. FTIR was conducted with a mid-infrared thermo scientific spectrometer (Nicolet 6700) at 4 cm^{-1} scanning resolution. The surface morphology of the MS10B10 backfill was observed by SEM ((LEO1530VP) using 5 kV accelerating voltage. The MIP tests were conducted on freeze-drying samples of MS5B5, MS5B10, MS5B15, MS10B10, MS10B15. As per [ASTM D4404 \(2010\)](#), dry samples were carefully cut from the core of cured samples in approximately 1 cm^3 size and then tested with an Auto Pore IV 9510 mercury intrusion porosimeter at a pressure range from subambient to 413 MPa to determine the pore size distribution.

3. Macroscopic performance: strength and permeability

3.1 Unconfined compressive strength (UCS)

Fig. 2 shows the development of UCS, dry density, void ratio and pH of the MSB specimens cured from 14 to 120 days. For each mix composition, a clear trend that UCS increases with increasing curing time is observed. UCS decreases with the increasing bentonite content regardless of the binder content at the same slump value. This can be attributed to the higher water content associated with higher bentonite content to achieve the equivalent workability during the mixing process (**Table 1**).

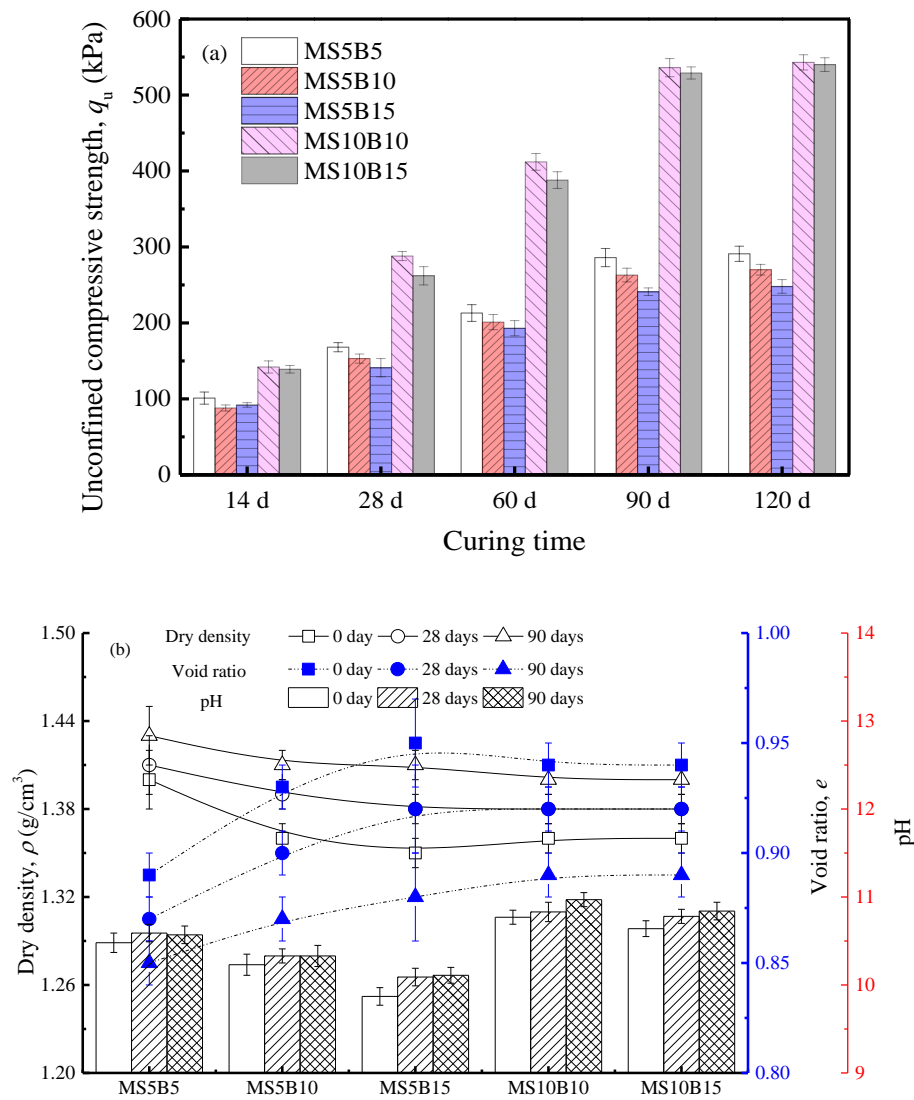


Fig. 2 Variation of unconfined compressive strength (a) and dry density, void ratio and pH (b) of MSB specimens with curing time (partial data reported in Wu et al., (2019a))

At 5% binder content, increasing the bentonite content from 5% to 15% (i.e., comparing

MS5B5 with MS5B10 and MS5B15, respectively) is associated with higher initial water content, leading to a slight decrease in dry density and pH value and increase in void ratio, which is consistent with the declined trend of UCS. The ultimate UCS values are approximately doubled when the binder content increased from 5% to 10% at the same bentonite content (i.e., MS5B10 vs. MS10B10 & MS5B15 vs. MS10B15). The enhanced UCS is mainly due to more binding phases formed when more binder is present, which is also manifested via the elevated pH value as can be see from **Fig. 2 (b)**. The higher pH value facilitates the activation of GGBS and hence producing more hydration products (i.e., C-S-H and Ht) in MSB specimens, leading to higher UCS. On the other hand, there are only marginal changes of the dry density and void ratio since all particles have similar densities and the initial water contents are also close as shown in Table 1.

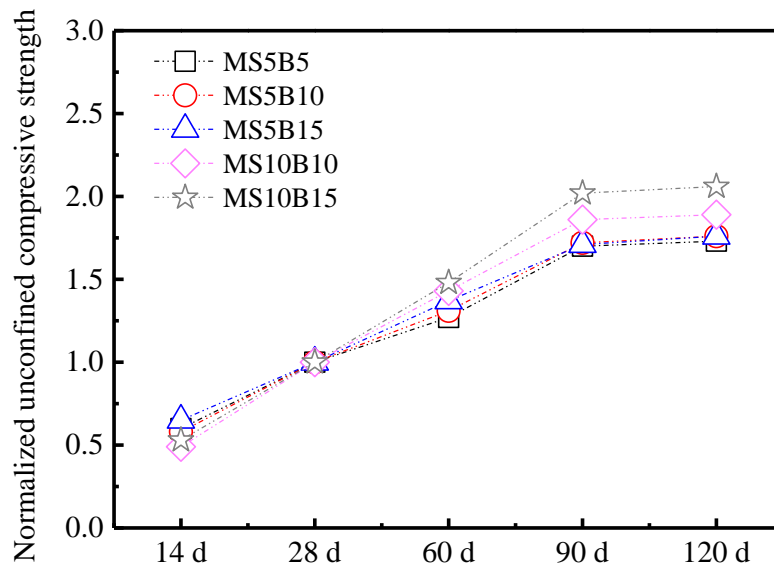


Fig. 3 Normalized unconfined compressive strength at 28-day value of MSB specimens

The UCS values are all normalized to the 28-day value for each mixture as shown in **Fig. 3**.

The UCS values of the MSB backfills keep increasing after 28 days and plateau after ~90

days. This finding is useful for designing and evaluating the site performance of novel backfill materials as more reasonable curing ages should be adopted for assessment. Particularly, the strength and impermeability development (see Section 4.2) should be modelled properly to account for their long-term performance in order to achieve a more economical design.

3.2 Hydraulic conductivity

Fig. 4 displays the variations of hydraulic conductivity (k) with curing time. The k values of MSB show a pronounced reduction by more than one order of magnitude as curing time increases from 28 to 90 days. The values of k decrease by 52% when the binder content increased from 5% to 10% at 90 days (2.3×10^{-10} m/s for MS5B15 vs. 1.1×10^{-10} m/s for MS10B15). On the other hand, increasing bentonite dosage from 5% to 15% in MSB backfill specimens at 90 days reduces k by ~63% (from 6.3×10^{-10} m/s to 2.3×10^{-10} m/s), which can be attributed to the expanded bentonite particles filling pores. The swelling potential of the bentonite particles are well preserved in the low-pH MgO-activated GGBS cement as indicated in [Wu et al., \(2019a\)](#). It should be noted that the k values of MSB are lower than 1.0×10^{-8} m/s at 28 days as suggested by Evans (1993) and lower than 1.0×10^{-9} m/s at 90 days as required by ICE (1999).

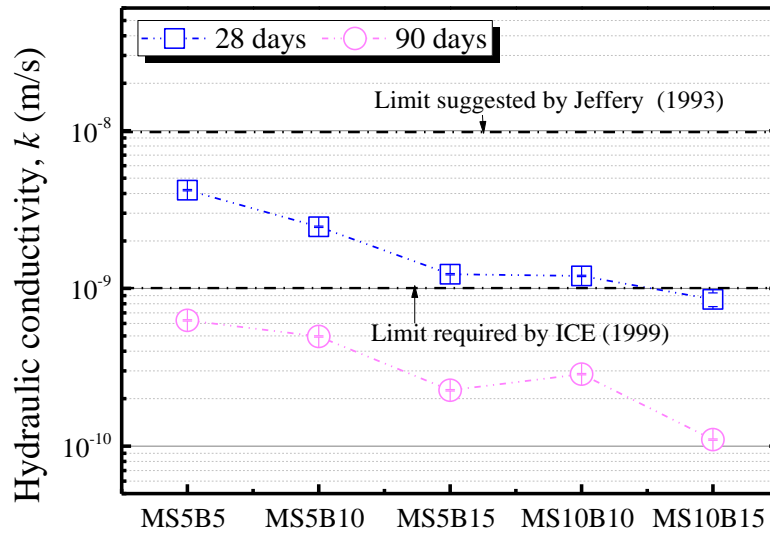


Fig. 4 Hydraulic conductivity measured in the flexible wall cells with tap water for each mixture after curing for 28 days and 90 days (partial data reported in Wu et al., (2019a))

4 Microstructural analysis

4.1 X-ray diffraction

The XRD results in **Fig. 5** present the hydration products of the MS10B10 backfills cured for 90 days. The main hydration products are hydrotalcite-like phases (Ht), calcium silicate hydrates (C-S-H), and monosulfate (AFm). The peaks for Ht (identified using $\text{Mg}_6\text{Al}_2\text{CO}_3(\text{OH})_{16} \cdot 4\text{H}_2\text{O}$, PDF#00-0220700) at $2\theta \approx 12.5^\circ$ and 24.0° can be identified, and their intensities increase with the augment in the binder dosage (i.e., MS5B10 vs. MS10B10 and MS5B15 vs. MS10B15) (Jin et al., 2015). The weak Bragg reflections of C-S-H can be detected at $2\theta \approx 23.0^\circ$, 31.6° and 37.5° . The characteristic peak of quartz (SiO_2) in the MSB backfills is detected at $2\theta \approx 21.6^\circ$ from the raw local clayey sand and bentonite (Wu et al., 2019b). The similar magnitude of peaks corresponding to the monosulfate ($\text{Ca}_4\text{Al}_2(\text{OH})_{12}(\text{S}_{0.5}\text{O}_2(\text{OH}_2)_6)$, PDF# 01-0831289) at $2\theta \approx 9.9^\circ$ and 19.8° (Ramachandran et al., 2002) can be observed in all mixes ascribed to the sulfate in the GGBS (**Table 2**). Additionally, the characteristic peak of bentonite (PDF# 00-0030019) is detected at $2\theta \approx 7.48^\circ$, agreeing well

with the previous findings (Yang et al., 2018). The peak of calcite (CaCO_3 , PDF# 00-0030569) ($2\theta \approx 29.8^\circ$), possibly from the raw materials or the carbonation of the hydration products, overlaps with that of C-S-H.

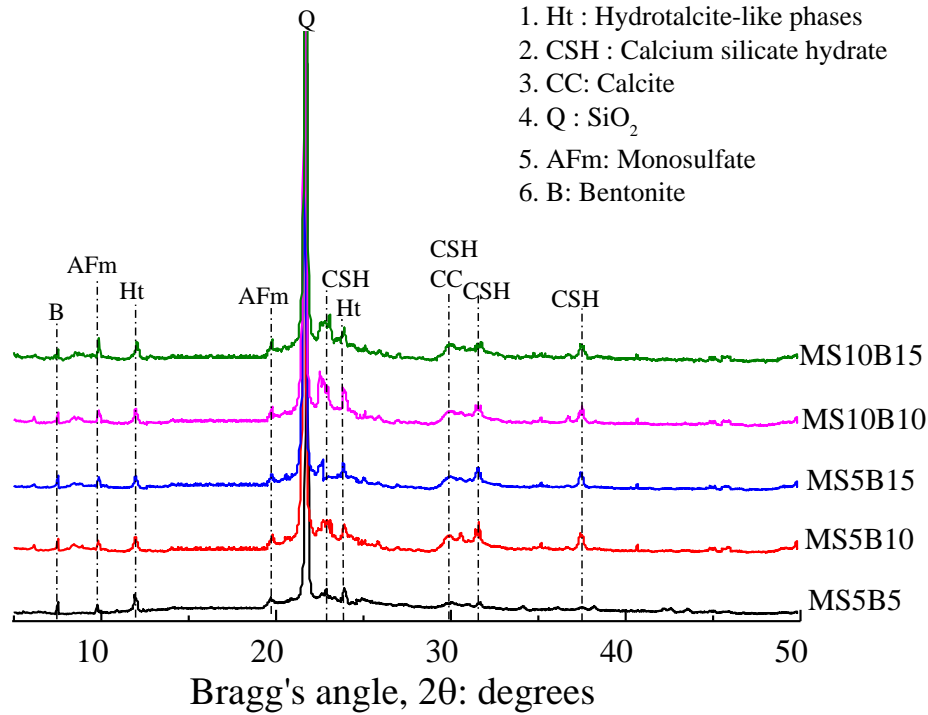


Fig. 5 X-ray diffractograms patterns for the 90-day curing of MS10B10 backfills

4.2 Scanning electron microscopy and energy dispersive X-ray spectroscopy

Fig. 6 shows the microstructure of MS10B10 backfills after 90 days' curing, in which various phases of different morphologies can be observed. It can be seen from **Fig. 6 (a)** that the thin sheet of bentonite particles is characterized by a curly and smooth feature. The large flocculated bentonite and hydration products of MS10B10 backfills fill the pore spaces of soil particles (**Fig. 6 (b)**) and then induced a low hydraulic conductivity (Yang et al., 2018). As shown in **Fig. 6 (c)**, the well-dispersed bentonite particles within the matrix also wrap the soil particles. The well-defined hexagonal platelets of monosulfate (AFm) can be observed in **Fig. 6 (d)**.

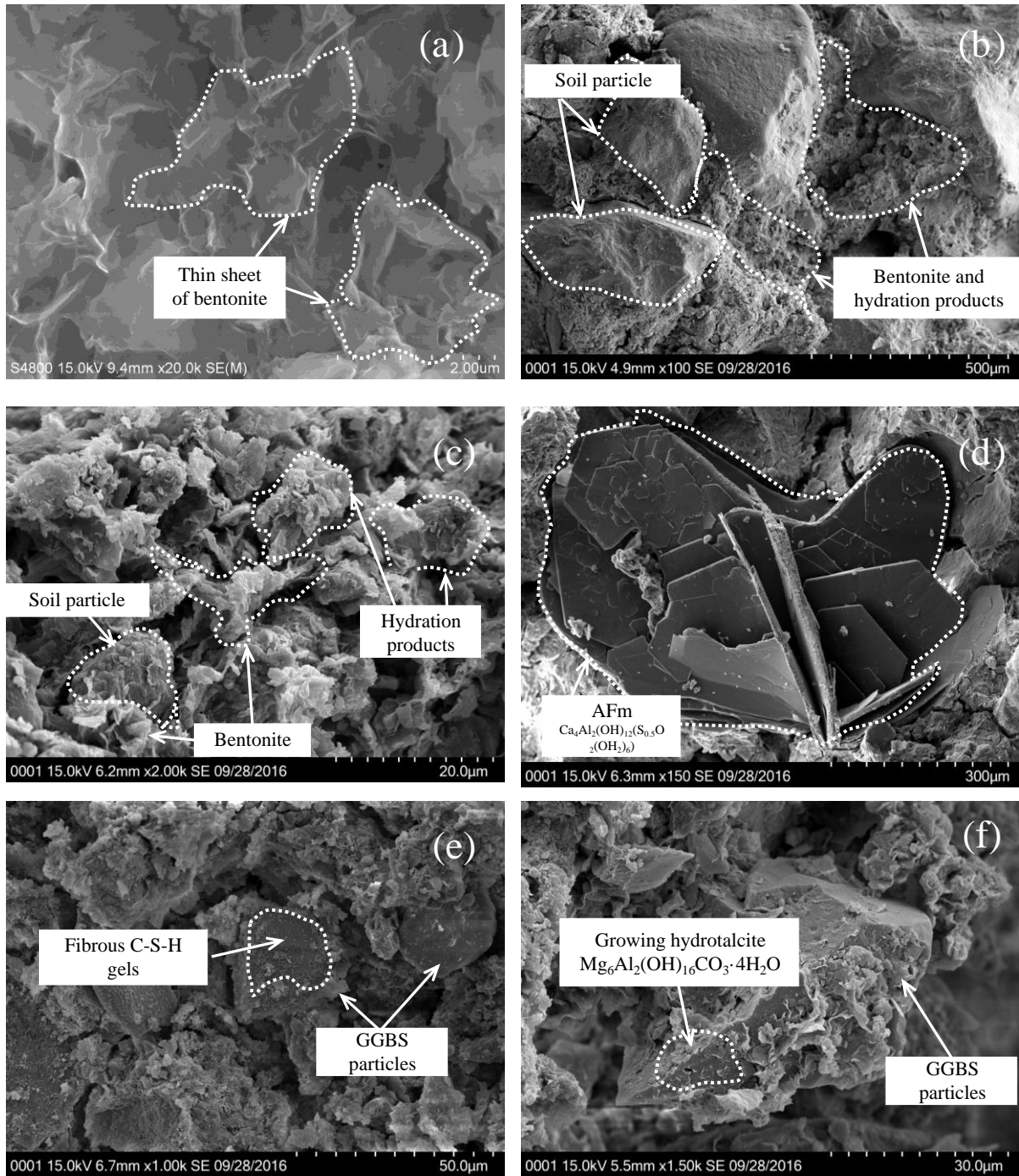


Fig. 6 SEM images for MS10B10 backfills after 90-day curing

The rough surface of slag particles is undergoing dissolution and then form the fibrous C-S-H (Fig. 6 (e)). The dissolution of irregular slag particles also facilitates the formation of Ht as shown in Fig. 6 (f), and the observed morphology is consistent with the previous research (Kovanda et al., 2005). These hydration products, swelling bentonite and unreacted slag

particles connect to form a dense microstructure, which leads to an increase in strength and decrease in hydraulic conductivity of the MSB backfills.

4.3 Fourier-transform infrared spectroscopy

The FTIR spectra of various MSB backfills and the summary of transmittance bands for hydration products of slag-based materials are presented in **Fig. 7** and **Table 2**, respectively.

The characteristic peaks of H-O, C-O, Si-O, Al-O and Si-O-X (Mg/Al) bonds can be detected in this study. The characteristic bands of O-H bond and H₂O molecules appear at around 3480 cm⁻¹ and 1628 cm⁻¹ (Fernández-Jiménez and Palomo, 2005), which indicate the presence of water as shown in **Fig. 7**.

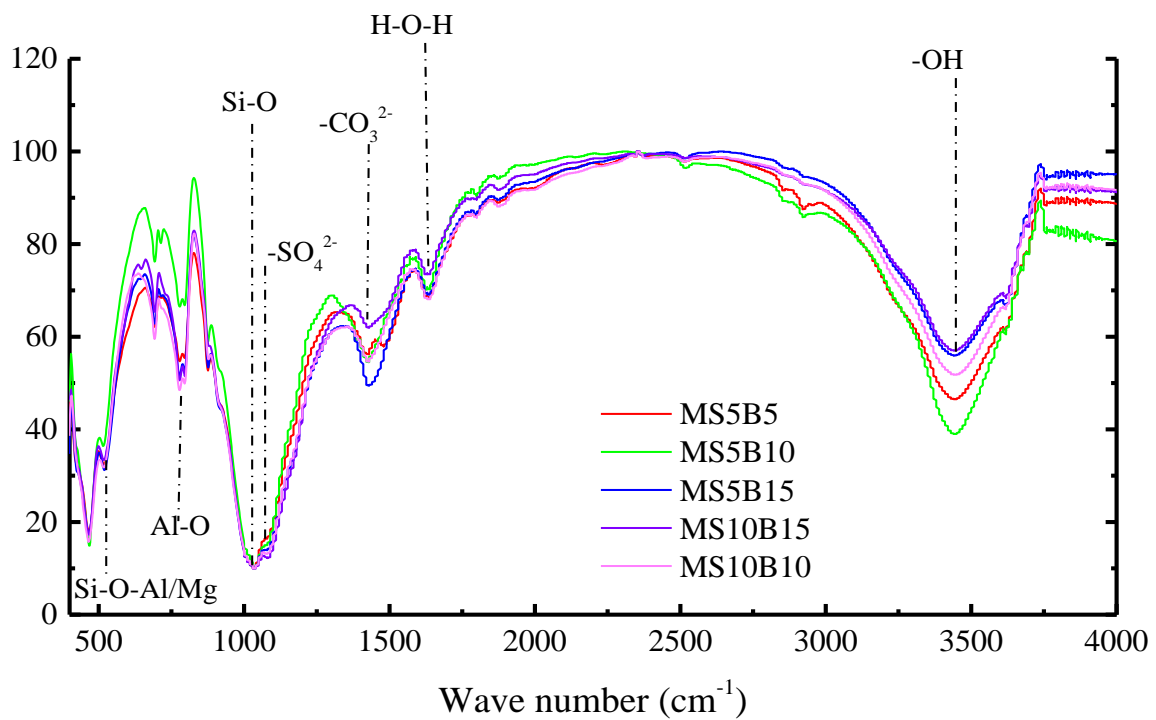


Fig. 7 Fourier-transform infrared spectra conducted on the 90-day MS10B10 backfill

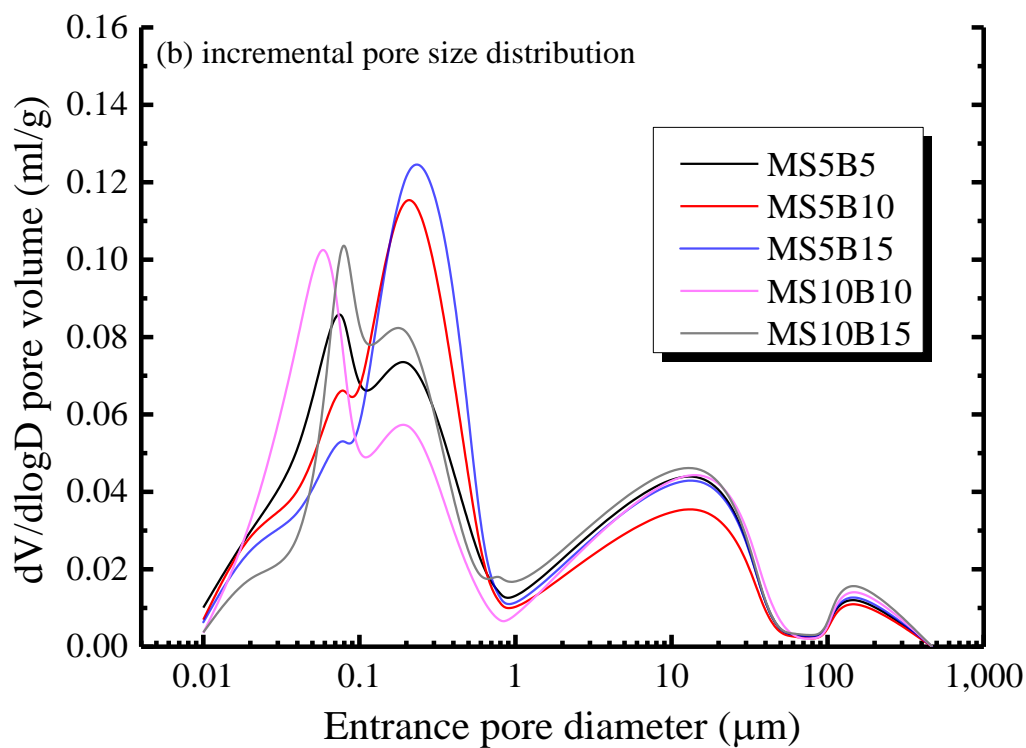
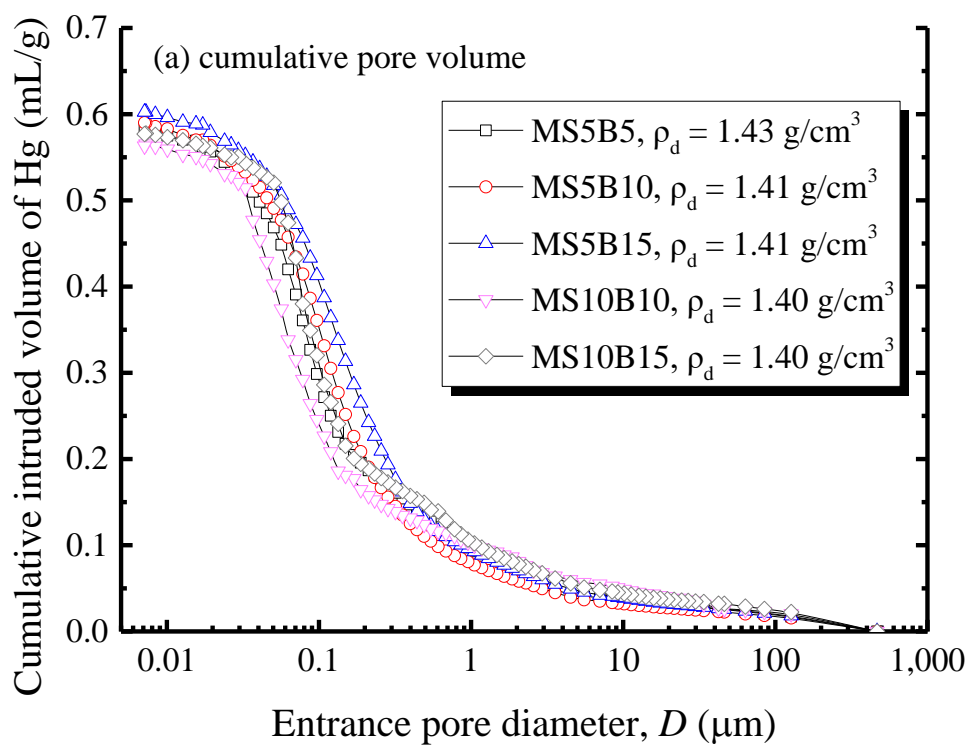
The band of -CO₃²⁻ at ~1440 cm⁻¹ (Mozgawa and Deja, 2009) confirms the presence of calcite as shown in **Fig. 5**. The slight characteristic -SO₄²⁻ stretching band at ~1057 cm⁻¹ (Della et al. 2013) indicates the formation of AFm.

Table 3. Summary of transmittance bands for hydration products of slag-based materials

Characteristic Bond	Sharp Infrared bands	Reference
Si–O–Al/Mg	670 cm ⁻¹	Mozgawa and Deja, 2009
	540 cm ⁻¹	Zhang et al., 2007
Al–O	798 cm ⁻¹	Zhang et al., 2007
–SO ₄ ²⁻	1118 cm ⁻¹	Fernández-Jiménez and Palomo, 2005
–CO ₃ ²⁻	1415 cm ⁻¹	Mozgawa and Deja, 2009
	712 cm ⁻¹	Mozgawa and Deja, 2009
	1450 cm ⁻¹	Allahverdi and Najafi Kani, 2009
Si–O	467 cm ⁻¹	Zhang et al., 2007
	450 cm ⁻¹	Allahverdi and Najafi Kani, 2009
H ₂ O molecules	1645 cm ⁻¹	Fernández-Jiménez and Palomo, 2005
OH ⁻ groups	3450 cm ⁻¹	Fernández-Jiménez and Palomo, 2005

4.4 Mercury intrusion porosimetry

Fig. 8 shows the cumulative pore volumes (**Fig. 8 (a)**), incremental pore size distribution (**Fig. 8 (b)**) and proportions of different pore size ranges (**Fig. 8 (c)**) for the MSB backfills at 90 days. **Fig. 8 (a)** reveals that increasing bentonite or decreasing binder dosage increases cumulative pore volumes. The cumulative pore volume increases up to 7.9% as the bentonite dosage increases from 5% to 15% at 5% binder content in MSB backfills, i.e., MS5B5 versus MS5B15. It is also observed that the cumulative pore volume of MS10B10 is 12% lower than that of MS5B10 in MSB backfills. It should be noted that increasing bentonite dosage leads to increased initial water content and then increases the pore volume. With the binder content increasing from 5% to 10%, the decreased cumulative pore volume is attributed to the more hydration products (i.e., C-S-H and Ht) generated in the MSB backfills, which is consistent with the UCS trend. Meanwhile, the peak of incremental pore size distribution at 0.2-0.4 μm significantly increases as the bentonite dosage increases from 5% to 15% at 5% binder content (**Fig. 8 (b)**). The increased binder content from 5% to 10% (i.e., MS5B10 vs. MS10B10) shifts the peak from 0.2-0.4 μm to 0.04-0.05 μm.



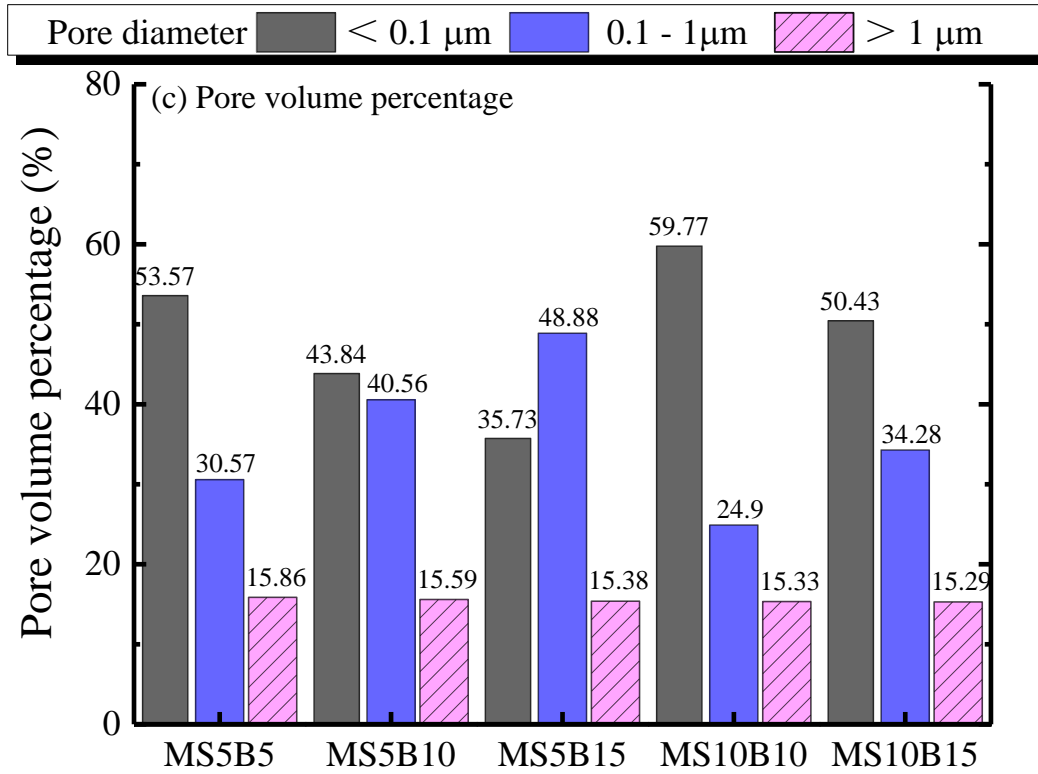


Fig. 8 Variation of cumulative pore volume (a), incremental pore size distribution and pore volume percentage (c) of 90-days MSB backfills

The three different pore size ranges for the MSB backfills can be classified: $>1 \mu\text{m}$, $0.1 - 1 \mu\text{m}$ and $< 0.1 \mu\text{m}$, respectively, as shown in **Fig. 8 (c)**. At 5% binder dosage, increasing bentonite dosage (MS5B5 vs. MS5B10 vs. MS5B15) results in larger volumes of pores with a diameter of $\sim 0.1 - 1 \mu\text{m}$, and smaller volumes of $< 0.1 \mu\text{m}$ pores. No significant changes are observed at the pores volume $> 1 \mu\text{m}$. Moreover, increasing the binder dosage at the same bentonite content (MS5B10 vs. MS10B10 and MS5B15 and MS10B15) significantly decreases the volume of pores with a diameter of $0.1 - 1 \mu\text{m}$ but increases the volumes of $< 0.1 \mu\text{m}$ pores.

5. Discussion

The reaction products and the microstructural features of the MSB backfills can be summarized as follows. Firstly, MgO particles dissolve in contact with water. The dissolution process of MgO provides an alkaline environment followed by the initial destruction of the covalent bonds

(i.e., Si-O, Al-O and Si-O-Al) in GGBS particles (Jin et al., 2015). Meanwhile, bentonite particles once encounter water, start to swell and fill the pore space between the individual soil particles and results in a significant decrease of hydraulic conductivity. In addition, the Ca^{2+} ions leached out from the GGBS and Mg^{2+} replace monovalent cations (i.e., Na^+), which are held on the exchangeable sites of bentonite particles and led to their flocculation. Additionally, the hydroxyl ions could react with the montmorillonite from the bentonite.

Literature shows that bentonite is relatively stable under a pH value of 12.6 (Ramírez et al., 2002; Sánchez et al., 2006; Omar et al., 2016; Du et al. 2020), which is much higher than the MSB system at pH of 10.2 - 10.8 (Wu et al., 2019a). It is also noted that the swelling performance of sodium-bentonite reaches its peak when bentonite was exposed to alum solution of pH values at 9.8 - 10.6 (Omar et al., 2016). Therefore, excellent impermeability of MSB backfills can be observed in this study due to the preserved swelling potential of bentonite particles as compared to those in the OPC-based backfills. Meanwhile, free ions (Ca^{2+} and Mg^{2+}) released from GGBS and MgO could react with the broken Si-O or Al-O bonds of GGBS to form C-S-H and Ht (Figs. 5-6) (with the possible inclusion of Al and sulfate in the structures). The increased binder dosage leads to lower cumulative pore volumes according to MIP and void ratio results (Figs. 2 and 8), which are correlated with the increased UCS. Further, increasing bentonite dosage leads to increased initial water content and then increases the pore volume (measured at the dry state by MIP). Here a distinction must be made regarding the pore volume at wet and dry states. At the wet state, increasing bentonite dosage can increase swelling volume and may reduce the pore volume and hydraulic conductivity. The monosulfate phase (Figs. 5) is produced from the relatively small amount of sulfate content in GGBS (and

possibly from the local soil), which also could contribute to filling the large-sized pores and converting them to relatively small-sized pores (**Fig. 8**). In summary, the improved UCS and impermeability of the MSB system can be attributed to the binding and pore filling effect of the various hydration products as well as the swelling bentonite particles (**Figs. 4 & 8**).

Previous studies extensively investigated the reaction mechanisms of the MgO-GGBS binder, showing its unique features of low-pH, high strength and low permeability in the long term (> 90 days) and superior durability. The present work focuses on the engineering properties and reaction mechanisms of MgO-GGBS-bentonite-soil mixtures targeting cutoff wall applications. The results apparently show that this novel backfill material satisfies both strength and permeability requirements at 28 days and much enhanced performance is demonstrated at 90 days. The preserved bentonite observed at 90 days shows the advantages of using low-pH MgO-GGBS in vertical cutoff walls and provides insights into developing more durable cutoff wall backfills. Moreover, the use of site soil proves to be successful in this study which would contribute to more economical cutoff wall backfills in the real applications. In order to promote the design/adoption of more sustainable and durable cutoff wall materials, more work is desired regarding contaminant transport modeling, contaminated site soil utilization and full-scale field trials. Besides, more fundamental research is warranted such as determining the reaction kinetics of the MgO-GGBS-bentonite system under different underground conditions and modification of bentonite to enhance its stability and sorption capacity.

6. Conclusions

The multiscale engineering characteristics, hydration products and microstructural

characteristics of MSB backfills with different dosages of MgO-activated GGBS (i.e., the binder) and bentonite are evaluated in this study. The following conclusions can be drawn:

(1) The strength and hydraulic conductivity of MSB backfills can achieve ~520 kPa and 1.1×10^{-10} m/s after curing 90 days, respectively. The strength and dry density increase with the increase of the binder content and decrease of bentonite dosage. Meanwhile, the hydraulic conductivity can be lowered by 52% and 63% as the binder increased from 5% to 10% and bentonite dosage increased from 5% to 15%, respectively.

(2) The engineering properties of MSB backfills improved significantly from 28 to 90 days. For geotechnical practices especially the cutoff walls, it is important to evaluate the performance of MSB backfills properly taking into account of the evolution of the strength and impermeability over the long term to enable more efficient and economical design.

(3) The hydration products of MSB backfills include hydrotalcite-like phases, calcium silicate hydrates, monosulfate, and portlandite. Furthermore, the increased binder dosage leads to lower cumulative pore volume, which agrees well with the trends observed in strength and hydraulic conductivity development.

(4) The bentonite particles are relatively stable in the MSB system. The hydration products, binding the adjacent soil particles and filling the pores in between, together with the swelling of bentonite, contribute to the enhancement of the mechanical performance and impermeability of the MSB backfills.

Acknowledgments

Financial support for this project was provided by National Key R & D Program of China

375 (Grant No. 2018YFC1803100), National Natural Science Foundation of China (Grant No.
376 41877248), Primary Research & Development Plan of Jiangsu Province (Grant No.
377 BE2017715).

378

379 **References**

380 [1] Allahverdi, A., & Najafi Kani, E. 2009. Construction wastes as raw materials for
381 geopolymer binders. *International Journal of Civil Engineering*, 7(3), 154-160.

382 [2] ASTM. 2008. Standard test method for unconfined compressive strength index of
383 chemical-grouted soils. ASTM D4219. West Conshohocken, PA, USA: ASTM
384 International.

385 [3] ASTM. 2010a. Standard Test Methods for Laboratory Determination of Water (Moisture)
386 Content of Soil and Rock by Mass. ASTM D2216. West Conshohocken, PA, USA: ASTM
387 International.

388 [4] ASTM. 2010b. Standard Test Methods for Liquid Limit, Plastic Limit, and Plasticity Index
389 of Soils. ASTM D4318, West Conshohocken, PA, USA: ASTM International.

390 [5] ASTM, 2010c. Standard Test Method for Measuring the Exchange Complex and Cation
391 Exchange Capacity of Inorganic Fine-Grained Soils. ASTM D7053, West Conshohocken,
392 PA, USA: ASTM International.

393 [6] ASTM, 2010d. Standard test method for determination of pore volume and pore volume
394 distribution of soil and rock by mercury intrusion porosimetry. ASTM D4404. West
395 Conshohocken, PA, USA: ASTM International.

396 [7] ASTM. 2016. Standard test methods for measurement of hydraulic conductivity of

397 saturated porous materials using a flexible wall permeameter. ASTM D5084. West
398 Conshohocken, PA, USA: ASTM International.

399 [8] ASTM. 2017a. Standard Practice for Classification of Soils for Engineering Purposes
400 (Unified Soil Classification System). ASTM D2487. West Conshohocken, PA, USA:
401 ASTM International.

402 [9] ASTM. 2017b. Standard Test Method for Flow Consistency of Controlled Low Strength
403 Material (CLSM). ASTM D6103. West Conshohocken, PA, USA: ASTM International.

404 [10] ASTM. 2018a. Standard Test Method for pH of Soils. ASTM D4972. West Conshohocken,
405 PA, USA: ASTM International.

406 [11] ASTM. 2018b. Standard test methods for laboratory determination of density (unit weight)
407 of soil specimens. ASTM D7263. West Conshohocken, PA, USA: ASTM International.

408 [12] Ben Haha, M., Lothenbach, B., Le Saout, G., and Winnefeld, F. 2011. Influence of slag
409 chemistry on the hydration of alkali-activated blast-furnace slag - Part I: Effect of MgO.
410 Cement and Concrete Research, 41(9): 955–963.

411 [13] Bernal, S. A., Provis, J. L., Rose, V., & De Gutierrez, R. M. 2011. Evolution of binder
412 structure in sodium silicate-activated slag-metakaolin blends. Cement and Concrete
413 Composites, 33(1), 46-54.

414 [14] Cerato, A. B., and Lutenecker, A. J. 2002. Determination of surface area of fine-grained
415 soils by the ethylene glycol monoethyl ether (EGME) method. Geotechnical Testing
416 Journal, 25(3), 315-321.

417 [15] Della Ventura, G., Ventruti, G., Bellatreccia, F., Scordari, F., & Guidi, M. C. 2013. FTIR
418 transmission spectroscopy of sideronatrite, a sodium-iron hydrous sulfate. Mineralogical

Magazine, 77(4), 499-507.

[16] Du, J., Zhou, A., Lin, X., Bu, Y. and Kodikara, J., 2020. Revealing Expansion Mechanism of Cement Stabilized Expansive Soil with Different Interlayer Cations through Molecular Dynamics Simulation. *The Journal of Physical Chemistry C*, 124(27), 14672–14684. <https://doi.org/10.1021/acs.jpcc.0c03376>

[17] Du, Y. J., Bo, Y. L., Jin, F., and Liu, C. Y. 2016. Durability of reactive magnesia-activated slag-stabilized low plasticity clay subjected to drying–wetting cycle. *European Journal of Environmental and Civil Engineering*, 20(2), 215-230.

[18] Evans, J.C. 1993. Vertical cutoff walls. In *Geotechnical Practice for Waste Disposal*; Daniel, D.E., Ed.; Chapman & Hall: London, UK, pp. 430–454.

[19] Fernández-Jiménez, A., & Palomo, A. 2005. Composition and microstructure of alkali activated fly ash binder: Effect of the activator. *Cement and concrete research*, 35(10), 1984-1992.

[20] Garvin, S. L., & Hayles, C. S. 1999. The chemical compatibility of cement–bentonite cut-off wall material. *Construction and Building Materials*, 13(6), 329-341.

[21] Heikal, M., Aiad, I., Shoaib, M. M., & El-Didamony, H. 2001. Physico-chemical characteristics of some polymer cement composites. *Materials chemistry and physics*, 71(1), 76-83.

[22] ICE (Institution of Civil Engineers). 1999. Specification for the construction of slurry trench cut-off walls as barriers to pollution migration. London: Thomas Telford.

[23] Jin, F., Gu, K., and Al-Tabbaa, A. 2014a. Strength and drying shrinkage of reactive MgO modified alkali-activated slag paste. *Construction and Building Materials*, 51, 395–404.

- 441 [24]Jin, F., and Al-Tabbaa, A. 2014b. Characterisation of different commercial reactive
442 magnesia. *Advances in Cement Research*, 26(2), 101-113..
- 443 [25]Jin, F., and Al-Tabbaa, A. 2014c. Evaluation of novel reactive MgO activated slag binder
444 for the immobilization of lead and zinc. *Chemosphere*, 117: 285-294.
- 445 [26]Jin, F., Gu, K. and Al-Tabbaa, A. 2015. Strength and hydration properties of reactive MgO-
446 activated ground granulated blastfurnace slag paste. *Cement and Concrete Composites*, 57,
447 8-16.
- 448 [27]Jin, F., and Al-Tabbaa, A. 2020. Magnesia in alkali activated cements. *Magnesia Cements*,
449 Elsevier, Oxford, United Kingdom, pp. 213-241.
- 450 [28]Joshi, K., Kechavarzi, C., Sutherland, K., Ng, M. Y. A., Soga, K., and Tedd, P. 2009.
451 Laboratory and in situ tests for long-term hydraulic conductivity of a cement-bentonite
452 cutoff wall. *Journal of Geotechnical and Geoenvironmental Engineering*, 136(4), 562-572.
- 453 [29]Kovanda, F., Koloušek, D., Cílová, Z., & Hulínský, V. 2005. Crystallization of synthetic
454 hydrotalcite under hydrothermal conditions. *Applied clay science*, 28(1-4), 101-109.
- 455 [30]Krizan, D., & Zivanovic, B. (2002). Effects of dosage and modulus of water glass on early
456 hydration of alkali–slag cements. *Cement and Concrete Research*, 32(8), 1181-1188.
- 457 [31]Lam, C., and S. A. Jefferis. 2017. *Polymer Support Fluids in Civil Engineering*, ICE
458 Publishing, One Great George Street, Thomas Telford Limited, London.
- 459 [32]Li, W., Yi, Y., & Puppala, A. J. 2020. Suppressing Ettringite-Induced Swelling of Gypseous
460 Soil by Using Magnesia-Activated Ground Granulated Blast-Furnace Slag. *Journal of*
461 *Geotechnical and Geoenvironmental Engineering*, 146(7), 06020008.
- 462 [33]Lodeiro, I. G., Fernández-Jimenez, A., Palomo, A., & Macphee, D. E. 2010. Effect on fresh

463 CSH gels of the simultaneous addition of alkali and aluminium. *Cement and Concrete*
464 *Research*, 40(1), 27-32.

465 [34]Mozgawa, W., & Deja, J. 2009. Spectroscopic studies of alkaline activated slag
466 geopolymers. *Journal of Molecular Structure*, 924, 434-441.

467 [35]Omar, M., Shanableh, A., & Al Zaylaie, M. (2016). Modification of the swelling
468 characteristics and phosphorus retention of bentonite clay using alum. *Soils and*
469 *Foundations*, 56(5), 861-868.

470 [36]Opdyke, S. M., and Evans, J. C. 2005. Slag-cement-bentonite slurry walls. *Journal of*
471 *Geotechnical and Geoenvironmental Engineering*, 131(6), 673-681.

472 [37]Park, S., Park, H. M., Yoon, H. N., Seo, J., Yang, C. M., Provis, J. L., & Yang, B. 2020.
473 Hydration kinetics and products of MgO-activated blast furnace slag. *Construction and*
474 *Building Materials*, 249, 118700.

475 [38]Ramírez, S., Cuevas, J., Vigil, R., & Leguey, S. 2002. Hydrothermal alteration of “La
476 Serrata” bentonite (Almeria, Spain) by alkaline solutions. *Applied Clay Science*, 21(5-6),
477 257-269.

478 [39]Sánchez, L., Cuevas, J., Ramírez, S., De León, D. R., Fernández, R., Villa, R. V. D., &
479 Leguey, S. 2006. Reaction kinetics of FEBEX bentonite in hyperalkaline conditions
480 resembling the cement–bentonite interface. *Applied Clay Science*, 33(2), 125-141.

481 [40]Schöler, A., Lothenbach, B., Winnefeld, F., & Zajac, M. 2015. Hydration of quaternary
482 Portland cement blends containing blast-furnace slag, siliceous fly ash and limestone
483 powder. *Cement and Concrete Composites*, 55, 374-382.

484 [41]Shand, M. A. 2006. *The chemistry and technology of magnesia*, John Wiley & Sons.

485 [42] Song, S., Sohn, D., Jennings, H. M., & Mason, T. O. 2000. Hydration of alkali-activated
486 ground granulated blast furnace slag. *Journal of materials science*, 35(1), 249-257.

487 [43] Wang, F., Jin, F., Shen, Z., & Al-Tabbaa, A. 2016. Three-year performance of in-situ mass
488 stabilised contaminated site soils using MgO-bearing binders. *Journal of Hazardous*
489 *Materials*, 318, 302-307.

490 [44] Wang, S.D., and Scrivener, K.L. 1995. Hydration products of alkali activated slag cement.
491 *Cement and Concrete Research*, 25(3): 561–571.

492 [45] Wu, H. L., Jin, F., Bo, Y. L., Du, Y. J., and Zheng, J. X. 2018a. Leaching and
493 microstructural properties of lead contaminated kaolin stabilized by GGBS-MgO in semi-
494 dynamic leaching tests. *Construction and Building Materials*, 172, 626-634.

495 [46] Wu, H. L., Zhang, D., Ellis, B. R., and Li, V. C. 2018b. Development of reactive MgO-
496 based Engineered Cementitious Composite (ECC) through accelerated carbonation curing.
497 *Construction and Building Materials*, 191, 23-31.

498 [47] Wu, H. L., Jin, F., Ni, J. and Du, Y. J. 2019a. Engineering properties of vertical cutoff
499 walls consisting of reactive magnesia-activated slag and bentonite: workability, strength
500 and hydraulic conductivity. *Journal of Materials in Civil Engineering*.

501 [48] Wu, H. L., Jin, F., & Du, Y. J. 2019b. Influence of wet-dry cycles on vertical cutoff walls
502 made of reactive magnesia-slag-bentonite-soil mixtures. *Journal of Zhejiang University-*
503 *SCIENCE A*, 20(12), 948-960.

504 [49] Wu, H. L., Du, Y. J., Yu, J., Yang, Y. L., & Li, V. C. 2020. Hydraulic conductivity and
505 self-healing performance of Engineered Cementitious Composites exposed to Acid Mine
506 Drainage. *Science of The Total Environment*, 716, 137095.

- 507 [50] Yang, Y. L., Reddy, K. R., Du, Y. J., & Fan, R. D. 2018. Sodium hexametaphosphate
508 (SHMP)-amended calcium bentonite for slurry trench cutoff walls: workability and
509 microstructure characteristics. *Canadian Geotechnical Journal*, 55(4), 528-537.
- 510 [51] Yi, Y., Liska, M., Al-Tabbaa, A. 2013. Properties of two model soils stabilized with
511 different blends and contents of GGBS, MgO, lime, and PC. *Journal of Materials in Civil
512 Engineering*, 26, 267-274
- 513 [52] Yi, Y., Liska, M., Jin, F., & Al-Tabbaa, A. 2016. Mechanism of reactive magnesia-ground
514 granulated blastfurnace slag (GGBS) soil stabilization. *Canadian Geotechnical Journal*,
515 53(5), 773-782.
- 516 [53] Zhang Y.S., Sun W., Chen Q.L., C., & Chen L. 2007. Synthesis and heavy metal
517 immobilization behaviors of slag based geopolymer. *Journal of hazardous materials*,
518 143(1-2), 206-213.
- 519 [54] Zheng, J., Sun, X., Guo, L., Zhang, S., & Chen, J. 2019. Strength and hydration products
520 of cemented paste backfill from sulphide-rich tailings using reactive MgO-activated slag
521 as a binder. *Construction and Building Materials*, 203, 111-119.