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Increasing nutrient inputs risk a surge of nitrous oxide emission from global mangrove ecosystems.

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1 *One Earth*

2 **Increasing nutrient inputs risk a surge of nitrous oxide emissions**
3 **from global mangrove ecosystems**
4

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16

17

18 **SUMMARY**

19 We document a substantial increase in global N₂O emissions from mangroves. Based on our analysis
20 of two decades of mangrove N₂O emission studies, we estimate N₂O emission of 0.023 Tg N yr⁻¹ from
21 global mangrove ecosystems. N₂O fluxes from mangrove ecosystems are strongly increased by
22 sediment dissolved inorganic nitrogen (DIN) concentration transported from river catchments to
23 coastal waters. Continuing growth of nutrient inputs from anthropogenic sources, i.e. agricultural
24 intensification, excessive fertilizer use and waste water discharge, will appreciably increase DIN
25 loading and consequently global N₂O emission from mangroves. Based on the Millennium Ecosystem
26 Assessment scenarios of riverine DIN inputs into mangrove ecosystems coupled with our estimates
27 of DIN controlled emissions rates, we expect N₂O emission to increase by 20 to 51% by 2030 and 27
28 to 74% by 2050 compared to estimated emissions in the year 2000. These forecasts underline the
29 urgency of improvements in catchment-scale nitrogen management strategies.

30

31 Keywords: mangroves, nitrous oxide, emissions, nutrients, river catchments

32

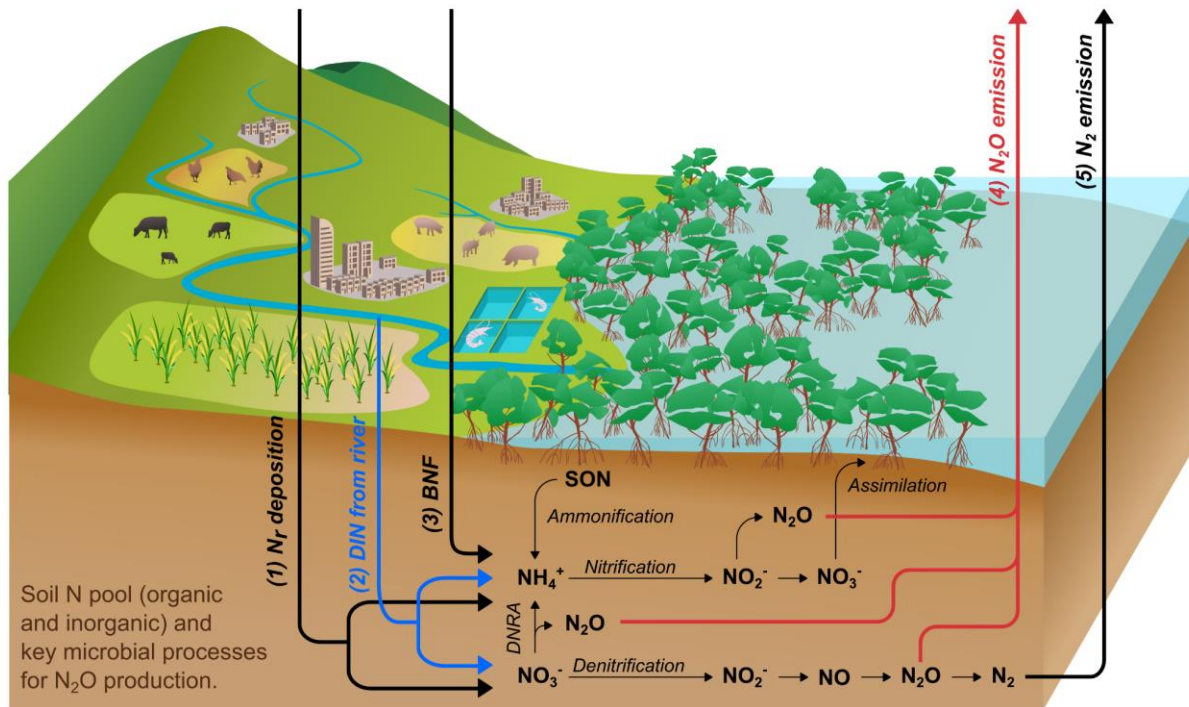
33 INTRODUCTION

34 Mangrove ecosystems are distributed in coastal waters in tropical and subtropical regions, delivering
35 a range of essential ecosystem services and playing an important role in climate change mitigation
36 ^{1,2}. They are regarded as overall sinks for atmospheric carbon (C) due to their waterlogged and
37 predominantly anoxic soil conditions and the inhibition of methanogenesis by the inflow and preferred
38 reduction of marine sulphate. Although both nitrification and denitrification can be sources of N₂O
39 emissions from mangroves, net emissions are often limited by the availability of mineral nitrogen (N)
40 as a substrate for chemolithotrophic microbes (nitrification) and an electron acceptor for microbial
41 respiration (denitrification) in most intact mangrove ecosystems ³. Under natural conditions, the
42 demand for available N to support C capture by mangrove trees is met mainly by mineralisation of
43 soil organic N, biological N fixation and through riverine input of dissolved inorganic N (DIN) into
44 mangroves (**Figure 1**). However, due to the growth of nutrient discharge from anthropogenic sources,
45 such as agricultural intensification coupled with excessive synthetic N fertilizer use worldwide, riverine
46 DIN loads have increased substantially ⁴, chronically elevating downstream receptor systems,
47 including mangroves (**Figure 1**).

48 Increased DIN inputs to mangroves in excess of plant metabolic demands pose a significant risk to
49 rising N₂O emission (**Figure 1**), enhancing global warming and stratospheric ozone destruction ⁵⁻⁸.
50 Given that N₂O is 298 times more potent in inducing global warming than CO₂ in the atmosphere on
51 a mass basis, there is a severe risk that potential benefits of nitrate removal in eutrophied coastal
52 ecosystems such as mangroves can be significantly offset by enhanced N₂O emission.

53 Unfortunately, global estimates of N₂O emission from mangroves are currently not explicitly included
54 in the annual national greenhouse gas emission budgets and scenario projections of the IPCC ⁹.
55 Instead, previous IPCC assessments considered mangroves only under the broader category of
56 global estuaries and riverine systems where emission factors have been developed only for former
57 mangrove ecosystems that have been converted to aquaculture ⁹. N₂O emission factors for
58 undisturbed, unconverted, or restored mangrove ecosystems that are similarly exposed to increasing

59 anthropogenic DIN loading have yet to be developed to correctly estimate the real extent of current
 60 N₂O emissions from mangroves. Such estimates are needed to accurately assess the implications for
 61 future N₂O emissions from mangroves under the scenarios of global increases of DIN inputs into
 62 coastal ecosystems¹⁰.



63

64 **Figure 1. The Nitrogen Cycle in the Mangrove Habitat**

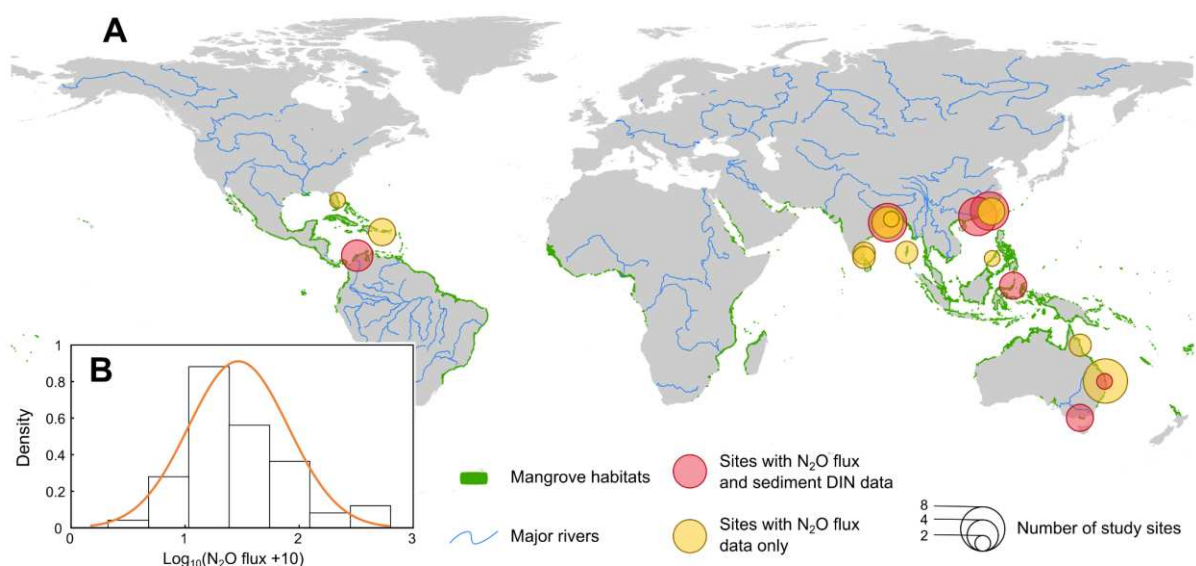
65 Nr: reactive nitrogen; SON: soil organic nitrogen; DIN: dissolved inorganic nitrogen; BNF: biological nitrogen fixation; DNRA:
 66 dissimilatory nitrate reduction to ammonium.

67

68 The potential oversight of explicit inclusion of current and future N₂O fluxes from mangrove
 69 ecosystems suggest that there is an urgent need for development of emission factors for N₂O fluxes
 70 under current and projected DIN loading into mangrove ecosystems. Previously estimated values of
 71 global mangrove N₂O emissions are highly uncertain^{3,6}, because they have been upscaled from
 72 single-region or median flux values without detailed evaluation of uncertainties. The wide range of
 73 estimates is likely caused by regional bias of previous studies and by technological challenges of
 74 adequately capturing the inherent spatial variability and temporal dynamics of N₂O emission from
 75 mangrove ecosystems. Existing estimates of global mangrove N₂O emissions do not support reliable

76 projections of future increases in anthropogenic riverine DIN inputs and consequent future emissions
77 of N₂O from mangrove ecosystems.

78 To assess the current extent and future trajectory of global N₂O emission from mangroves under
79 increasing anthropogenic DIN inputs, we performed a systematic global meta-analysis of N₂O
80 emission studies conducted during the past two decades. These studies include 58 site-based
81 mangrove N₂O fluxes of which 26 also contained data on sediment DIN concentration (**Table S1**;
82 **Figure 2A**). We used this information to develop a comparison between mangrove N₂O fluxes and
83 sediment DIN concentrations which were then used in scenarios analyses to estimate future N₂O
84 emission based on the projections of DIN discharge from catchments.



85
86 **Figure 2. The Distribution of Mangrove Habitats and Mangrove Field Study Sites**

87 **(A)** Global distribution of mangrove habitats based on Giri et al.,² and spatial distribution of mangrove field study sites with
88 exclusively N₂O flux data (N = 29) as well as sites with both N₂O flux and sediment DIN concentration data (N = 29). **(B)**
89 Distribution of experimentally observed mangrove N₂O fluxes obtained from 58 field sampling sites globally.

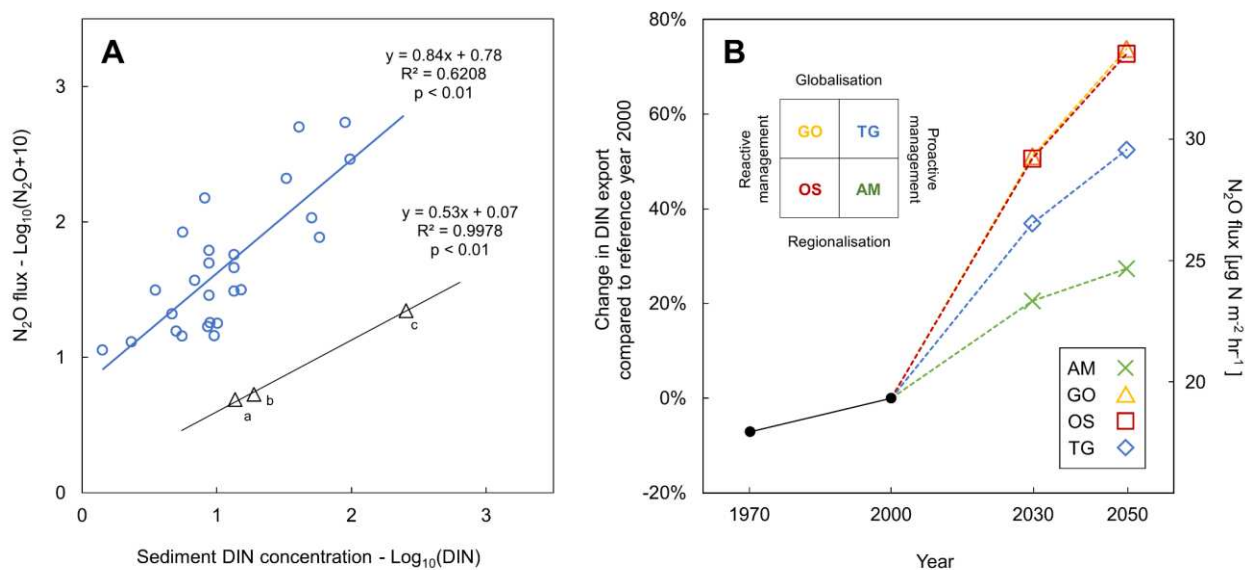
90

91 **RESULTS AND DISCUSSION**

92 **Global N₂O emission from mangroves**

93 We estimate a global mean mangrove N₂O flux of 19.31 $\mu\text{g N m}^{-2} \text{hr}^{-1}$ (95% CIs: [12.61, 28.00]) from
94 the 58 site observations that follow a normal distribution passing a Kolmogorov–Smirnov test (**Figure**

95 **2B)**. This average emission rate lies within the range of previously reported N₂O fluxes that vary from
 96 0.42 to 37.50 μg N m⁻² hr⁻¹ ³. According to a global survey of mangrove ecosystems ², in the year 2000
 97 mangroves covered a total surface area of 137,760 km² in coastal regions north and south of the
 98 equator. Based on this surface area, the estimated total global N₂O emission from mangroves for the
 99 year 2000 is 0.023 Tg N yr⁻¹ (95% CIs: [0.015, 0.034]), which is of similar magnitude as previously
 100 reported ranges by Murray et al. ⁶ with 0.01-0.09 Tg N yr⁻¹, Corredor et al. ⁵ with 0.009-0.356 Tg N yr⁻¹
 101 ¹, and Chauhan et al. ¹¹ with 0.07 Tg N yr⁻¹ estimated by upscaling single-region fluxes or median
 102 values of several reported ranges. The global warming potential of N₂O emission of 0.023 Tg as N yr⁻¹
 103 ¹ (0.037 Tg as N₂O yr⁻¹) is equal to an emission of 3.01 Tg as C yr⁻¹ given that N₂O is 298 times potent
 104 in inducing global warming relative to CO₂ on a mass basis for 100-year time horizon. This is even
 105 higher than the mangrove CH₄ emission (2.53 Tg C year⁻¹) over 100 years ¹².



106

107 **Figure 3. The Relationship between N₂O Flux and Sediment DIN Concentration and Their Predictions in Future**
 108 **Scenarios**

109 **(A)** Log-log plot highlighting the relationship between N₂O flux and sediment DIN concentration. A total of three linear
 110 regression models were performed. The first model identified three outliers from the 29 sites, and two additional models
 111 were consequently conducted based on the 26 sites and the 3 identified outliers respectively (see Data and Methods). The
 112 outcomes of the latter two regression models are presented: the blue solid line denotes the fitted line for 26 points where
 113 the 3 outliers (triangles) are excluded; and the black line is the fitted line for the 3 outliers. The three outliers are: a) Likupand,
 114 Sulawesi island, Indonesia ³⁴; b) Teremaal, Sulawesi island, Indonesia ³⁴; c) Chelmer, Queensland, Australia ⁴⁰. **(B)** Trends
 115 and future scenarios of the catchment-exported DIN in coastal mangroves and the N₂O flux. The four Millennium Ecosystem
 116 Assessment scenarios are: AM – Adapting Mosaic; GO – Global Orchestration; OS – Order from Strength; TG – Techno-
 117 garden.

118

119 **Relationship between N₂O flux and sediment DIN**

120 Our meta-analysis showed a log-log correlation ($R^2 = 0.6208$, $p < 0.01$) between mangrove N₂O flux
121 and sediment DIN concentration observed globally (**Figure 3A**). This indicates that allochthonous
122 nutrient sources from rivers can increase the DIN levels in coastal habitats and sediments^{13,14}, making
123 substantial contributions to mangrove nitrogen processing and N₂O emission. Using this correlation
124 allows us to estimate future global trajectories of N₂O emission from mangrove ecosystems based on
125 projected anthropogenic DIN fluxes from their respective river basins. Organic matter decomposition
126 in mangrove sediments is slow as compared to catchment imports as evidenced in global reviews of
127 nitrogen dynamics in mangroves e.g.³. Therefore, this correlation compellingly indicates that
128 mangroves are highly efficient users of DIN from riverine source waters, especially in the condition of
129 nitrogen enrichment caused by anthropogenic effluents. The availability of DIN is a key limiting factor
130 for nitrogen transformation rates in the sediment of fringe and interior mangroves, which are generally
131 DIN sinks, taking nitrogen from overlaying waters^{6,15,16}. In addition, the autochthonous N input through
132 biological N fixation in mangroves can be suppressed by the high loading of DIN given that BNF is an
133 energy expensive process^{17,18}. For example, 10 to 28 g of glucose is invested per gram of N fixed
134 and thus DIN loading will suppress fixation¹⁹. This suggests that N₂O emission in mangroves may
135 increase as a result of anthropogenic N enrichment³.

136 **Scenarios of future global N₂O emission**

137 We predict how global N₂O emissions from current mangrove ecosystems might rise in the future,
138 based on projections of riverine DIN exports in the four Millennium Ecosystem Assessment scenarios
139 (Adapting Mosaic - AM, Global Orchestration - GO, Order from Strength - OS, Techno-garden - TG)
140 (**Figure 3B**) that consider four different combinations of (i) globalisation vs regionalisation, and (ii)
141 reactive vs proactive approaches to environmental issues^{4,20}, see Table S3. As a result of such
142 projected increases of riverine DIN exports from their catchments, N₂O emission from mangrove
143 ecosystems will rise by 21-51% in 2030, and by 27-74% in 2050 compared to the emission estimates
144 of year 2000, depending on the respective scenarios (**Figure 3B**).

145 Thus, these results indicate consistent increases in global N₂O emission from mangrove ecosystems
146 under all of the four scenarios. They contrast with previous studies that show mixed trends of changes
147 in anthropogenic N₂O emissions (from agriculture, energy, transport, biomass burning, etc.) from 2005
148 to 2050 depending on strategies and scenarios (business-as-usual: 83% increase; moderate
149 mitigation: 26% increase; concerted mitigation: 21% decrease)²¹. Differences in mangrove N₂O
150 emissions among respective scenarios suggests the potential efficacy of various interventions aimed
151 at effectively managing N derived from river catchments. Predicted future N₂O fluxes in the GO and
152 OS scenarios that both follow reactive approaches are similar and among the highest, exceeding
153 predicted future fluxes of the scenarios assuming proactive approaches by 20-21 % (TG) or even 45-
154 46 % (AM) in 2050. The lowest increase in N₂O emissions was predicted for the AM scenario, which
155 has a regional watershed-scale focus and follows strongly proactive environmental management
156 strategies. The recycling of animal manure and human excreta in agriculture in the AM scenario
157 results in decreased DIN exports from catchments towards their coastal mangrove wetlands than the
158 other scenarios that are dependent on the use of synthetic N fertilization of agricultural land uses
159 **(Table S3)**.

160 These results suggest the global significance of a linear increase in N₂O emissions from mangrove
161 ecosystems being triggered by the predicted rising riverine DIN fluxes. The relative proportion of N₂O
162 emitted from mangroves to the total global N₂O emissions (natural and anthropogenic sources) on
163 area (land cover basis) exceeds their earth surface cover by 374.07% (see Supplementary
164 Information), with flux values (19.31 µg N m⁻² hr⁻¹) that match those from herbaceous wetlands (19.29
165 µg N m⁻² hr⁻¹), but greatly exceed that from freshwater forested wetlands (6.05 µg N m⁻² hr⁻¹), shrubs
166 (6.96 µg N m⁻² hr⁻¹), grasslands (10.73 µg N m⁻² hr⁻¹), and other types of forests (boreal: 5.37,
167 temperate: 12.22, tropical dry: 12.56 µg N m⁻² hr⁻¹)²². However, our results indicate that in contrast to
168 other greenhouse gas sinks that turn into sources (e.g. boreal wetlands and melting permafrost), the
169 increase in emissions from mangroves can be efficiently controlled if riverine DIN exports are reduced.
170 A reduction of accumulated N₂O emissions from mangroves as between scenarios GO and AM

171 (161.20 Tg CO₂ yr⁻¹) is comparable to the global N₂O emissions from the industry (137.91 Tg CO₂ yr⁻¹) and energy (175.86 Tg CO₂ yr⁻¹) sectors in 2010 ²³.

173 **Uncertainties in estimation**

174 We contend that our estimates of N₂O fluxes are conservative, being primarily based on surface water
175 flux data using static chambers (**Table S1**). These estimates therefore do not include N₂O transfer
176 from soils through mangrove tree stems and shoots into air under wet and/or inundated conditions,
177 as trees in wetland environments can be significant conduits for transporting greenhouse gases ^{24–26}.
178 In addition, the majority of the assessed N₂O water-air fluxes are originating from mangrove rivers
179 instead of inundated mangrove habitats (Table S2). Although some pristine rivers are found to act as
180 net sinks of N₂O ^{e.g. 27,28}, mangrove rivers are in general sources of N₂O, especially when they are
181 receiving an increasing load of anthropogenic DIN ³. However, these water bodies as N₂O sources
182 (see Figure S3) are not necessarily counted in the total estimated area of mangrove ecosystems ²,
183 which may consequently underestimate the total N₂O flux. Lastly, the underestimation may also be
184 caused by the under-represented mangrove habitats in equatorial regions (10°S - 10°N) (see Figure
185 S1), which may have larger N₂O fluxes because of their higher sediment temperature and increased
186 bacterial productivity potentials ⁶.

187 Based on the revealed relationship between N₂O fluxes and sediment DIN, the N₂O emissions in
188 future scenarios were predicted using projections of riverine DIN exports. Although external DIN also
189 enters coastal areas via fresh groundwater discharges, their contribution is relatively a small
190 proportion of riverine sources (from 2.4% to 7.5% depending on studies ^{29–31}) and there is a lack of
191 systematic global projections of DIN input via coastal groundwater discharges to fully demarcate its
192 net contribution to the DIN loading of the mangrove forest sediments. However, given intensified
193 agriculture and population growth posing a risk of increasing DIN discharge via fresh groundwater ²⁹,
194 our projected N₂O emissions are likely to be rather conservative as they do not include the impact of
195 less well constrained groundwater DIN contributions as a predictor. However, global datasets of
196 coastal groundwater discharge have been appearing in recent years ³², making it possible to establish
197 future research if this causes even higher N₂O emissions.

198 It has been a challenge to obtain a global estimate of the current and future N₂O emission by upscaling
199 local measurements, because of the high spatial and temporal variation of N₂O fluxes and sparse
200 data availability^{6,33}. This paper performs a systematic meta-analysis, attempting to overcome the
201 limitations of common extrapolation exercises by upscaling from single-region or median values. The
202 meta-analysis reveals a useful relation between mangrove N₂O flux and sediment DIN concentration,
203 and hence predicts a significant N₂O emission growth in the foreseeable future. However, the
204 estimation of global N₂O flux and emission carries uncertainty (**Figure 3A**). Parts of the uncertainty
205 were caused by the limited data coverage of the high heterogeneity of mangroves and their N₂O
206 emission (e.g. types, locations, seasons, or times of the day) in the existing studies, which potentially
207 obscures the impact of micro-scale hydro-biochemical processes on the relationship. For example,
208 the three outlier sites have much less N₂O fluxes compared to their sediment DIN concentration levels
209 (see **Figure 3A**), which might be caused by their micro-scale ecological attributes such as high salinity
210 that leads to decreased denitrification³⁴. Therefore, we suggest that further refinements through
211 experimental work and globally coordinated monitoring with broader coverage are needed to improve
212 the accuracy and robustness of the estimates so that reliable adaptive N and GHG management
213 strategies can be identified.

214 **CONCLUSIONS**

215 Based on a systematic meta-analysis, we estimate the global mean N₂O flux from mangrove
216 ecosystems to be 19.31 $\mu\text{g N m}^{-2} \text{ hr}^{-1}$. This number exceeds emissions from freshwater forested
217 wetlands, shrubs, grasslands, as well as boreal, temperate, and tropical dry forests. Considering the
218 global surface area covered by mangroves, N₂O emissions from world-wide mangrove ecosystems
219 are estimated to be 0.023 Tg N yr⁻¹.

220 Our analyses reveal a significant correlation between sediment DIN and N₂O emission, and a high
221 risk of increased mangrove N₂O emissions by up to 74% in 2050 due to enhanced nutrient inputs
222 from their upstream catchments. Because of their key impact on the global climate, we suggest
223 including global estimates of N₂O emission from mangroves in the annual national greenhouse gas

224 emission budgets and scenario projections of the IPCC. This result also implies that the benefits of
225 mangrove restoration efforts can be optimised if riverine DIN exports are managed efficiently to
226 reduce DIN loading in mangroves, and suggests that this significant biogeochemical connection
227 between rivers and coastal mangrove habitats is needed to be better reflected in a more integrated
228 river basin management approach.

229 Our study uncovers critical knowledge gaps related to the often site-specific mechanisms that control
230 mangrove N₂O emissions, highlighted by the existence of almost 10% of outliers in the otherwise
231 strong DIN/N₂O relationship found, which suggests that emission rates may be controlled by local or
232 regional factors. We therefore call for globally coordinated mangrove GHG emission monitoring to
233 improve accuracy and robustness of N₂O emission estimates and support mangrove management in
234 the future.

235 **EXPERIMENTAL PROCEDURES**

236 **Resource Availability**

237 ***Lead Contact***

238 Further information and requests for resources should be directed to and will be fulfilled by the Lead
239 Contact, Feng Mao (MaoF1@cardiff.ac.uk).

240 ***Materials Availability***

241 This study did not generate new unique materials.

242 ***Data and Code Availability***

243 The published article is a systematic meta-analysis and did not generate new datasets. The article
244 includes all datasets analysed during this study.

245 **Mangrove N₂O emission and sediment DIN metadata**

246 Mangrove N₂O flux and sediment DIN concentration data were collected from published literature,
247 based on Web of Science searches using a combination of two topic keywords 'mangrove' and
248 'nitrous oxide'. We went through all the returned 77 papers and found 2 more papers using the

249 'snowballing' techniques – these 2 papers were not in the returned searching results but were
250 mentioned or cited by them. 30 out of the 79 papers contain N₂O flux data in mangrove habitats.
251 Within the 30 papers, 26 have flux information at the sediment-air interface, and 6 provide data about
252 N₂O fluxes from surface water. Several of those 26 papers with flux information also provide nitrogen
253 concentration data in different forms (e.g. NH₄-N, NO₃-N and total nitrogen) and locations (e.g. in
254 sediment, porewater and surrounding surface waters). Among these, the sediment dissolved
255 inorganic nitrogen (DIN) information can be obtained directly or by combining NH₄-N, NO₃-N and NO₂-
256 N concentration in 16 of all papers, making DIN the most common and comparable nitrogen
257 concentration variable. The DIN concentration data from 29 sites of 12 scientific publications are
258 provided in 'mass per mass' units (e.g. µg N g⁻¹) while the data from 4 sites in 4 papers are provided
259 in 'mass per volume' units (e.g. µM). These two unit-systems are not directly comparable, so that the
260 first 29 sites were used for analysing the relationship between N₂O flux and sediment DIN
261 concentration.

262 Within each paper, the N₂O flux for a sampling site may be measured in different replicates, locations
263 (e.g. distance to the coast), seasons, or times of the day. These measurements were combined to
264 estimate a mean N₂O flux for each sampling site in a standardised unit (µg N m⁻² hr⁻¹).

265 Not all published flux and concentration data were provided explicitly in text or tables. For the those
266 presented in bar-charts, the length of bars was measured and converted into standard data formats
267 using ImageJ (version 1.51j8).

268 Two sets of data were consequently formed (**Figure 2A**): the first set has 58 sites from 26 scientific
269 publications, each with data of N₂O flux at the sediment-air interface; the second set is comprised of
270 29 sites from 12 publications with both the data of N₂O flux at the sediment-air interface and the data
271 of DIN concentration in sediment (see **Table S1**).

272 **Global mangrove distribution**

273 According to the map produced by Giri et al. ², mangroves had a global total area of 137,760 km² in
274 the year 2000. The World Atlas of Mangroves ³⁵ also has an estimation of the total mangrove area of

275 152,361 km². Spalding's study uses a range of different image-processing techniques but has
276 considerable expert validation while Giri's work provides a high-resolution dataset using consistent
277 and repeatable methods although with less expert review ³⁶. We used Giri's map in this research and
278 the data was downloaded from the UNEP-WCMC's ocean data viewer ([http://data.unep-
279 wcmc.org/datasets/4](http://data.unep-wcmc.org/datasets/4)).

280 **Estimation of mangrove N₂O flux**

281 The global mean mangrove N₂O flux was estimated based on the 58 sites from 26 papers that provide
282 flux data (**Table S1**).

283 The flux was logarithmised as:

$$284 \quad E_{log} = \log_{10}(E + 10)$$

285 where E denotes the original mangrove N₂O flux value ($\mu\text{g N m}^{-2} \text{ hr}^{-1}$), and E_{log} denotes the
286 logarithmised value. A 10 was added onto E since some of the original values were below zero.

287 The normality of the logarithmised sample distribution was tested using the Kolmogorov-Smirnov
288 normality test in R ³⁷. The mean and standard deviation of the normally distributed sample was
289 calculated by the maximum likelihood fitting using the 'fitdistrplus' package in R ³⁸.

290 **Relationship between N₂O flux and sediment DIN**

291 The sediment DIN was logarithmised as:

$$292 \quad DIN_{log} = \log_{10}(DIN)$$

293 where DIN denotes the original sediment DIN value ($\mu\text{g N g}^{-1}$), and DIN_{log} denotes the logarithmised
294 value.

295 The linear regression model was used to find the relationship between the N₂O flux and the sediment
296 DIN concentration.

$$297 \quad E_{log} = \alpha + \beta DIN_{log} + \epsilon$$

298 Linear regression was performed three times using slightly different datasets.

299 The first model was built based on 29 sites from 12 papers that provide both flux and sediment DIN
300 information (**Table S1**). Based on the result of the first model, 3 outlier sites were removed from the
301 29 sites because they are isolated from the general trend of the data pattern, deviate significantly
302 from the $y = x$ line of the diagnostic normal Q-Q plot with extreme standardised residuals (**Figure S2**).
303 The rest 26 sites from 10 papers were imported to build the second model, and the 3 outlier sites
304 were used to build the third model (**Figure 3**).

305 **DIN discharge from catchments**

306 The Nutrient Export from WaterSheds 2 (NEWS2) is an integrated model of nutrient exports from
307 more than 6000 rivers worldwide³⁹. Using this model, Seitzinger et al.⁴ estimate and analyse the past
308 trends and future trajectories of river DIN export in four Millennium Ecosystem Assessment scenarios,
309 namely Adapting Mosaic (AM), Global Orchestration (GO), Order from Strength (OS) and
310 Technogarden (TG) (**Table S3**). The model output that is the data of river DIN export from each river
311 were requested from the authors, covering the following years and scenarios: 1970, 2000, 2030-AM,
312 2030-GO, 2030-OS, 2030-TG, 2050-AM, 2050-GO, 2050-OS, 2050-TG. The changes in DIN export
313 are provided in **Figure S4** and **Figure S5**.

314

315 **Trend of mangrove N₂O emission**

316 The past and future trajectories of N₂O emission were estimated using two sets of information. First,
317 the calculated relationship between N₂O flux and sediment DIN concentration described above.
318 Second, the historical and predicted future DIN export from river basins that have river mouths
319 neighbouring coastal mangroves. For each mangrove habitat, the closest river basin was identified
320 according to the distance between the habitat and the river mouths using ArcMap 10.4.1. The DIN
321 export from the closest basin was regarded as the main influencer of the sediment DIN in the adjacent
322 mangrove habitat. The uncertainty of the prediction is presented in the form of scenario analysis, and
323 the 95% CIs of estimated N₂O flux and emission for each predicted year-scenario are provided in

324 **Table S4.** The total area of mangrove habitat (137,760 km²) is assumed to remain the same as in all
325 years and scenarios ², avoiding introducing additional uncertainties in predicting the distribution
326 change of mangrove under the joint influence of both forest degradation and restoration efforts.

327

328 **Comparison between mangroves' proportional global N₂O emissions and** 329 **proportional earth surface coverage**

330 According to Ciais et al. ¹⁰, the total global N₂O emission is 17.9 Tg N yr⁻¹, including 11.0 Tg N yr⁻¹
331 from natural sources and 6.9 Tg N yr⁻¹ from anthropogenic sources. Our estimated mangrove N₂O
332 emission in this paper is 0.023 Tg N yr⁻¹. The mangrove's proportion on global N₂O emission is
333 $0.023/17.9 = 0.128\%$. The total area of mangroves is 137,760 km² while the total earth surface area
334 is 510,072,000 km². The mangrove's proportion on earth surface is $137,760/510,072,000 = 0.027\%$.
335 Therefore, the mangrove's proportion on global N₂O emission exceeds their proportion on earth
336 surface cover by: $(0.128-0.027)/ 0.027 = 374.07\%$.

337

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353 **DECLARATION OF INTERESTS**

354 The authors declare no competing interests.

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