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Supplementary Material

The chemical composition and toxicological effects of fine particulate matter (PM_{2.5}) emitted from different cooking styles

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30

31 **Text S1:** Organic carbon (OC) and elemental carbon (EC) analysis

32 OC and EC were analysed on a punch (0.526 cm²) from quartz filter by thermal optical
33 reflectance (TOR) technique following the IMPROVE_A protocol on a thermal/optical carbon
34 analyser (DRI Model 2001, Atmoslytic Inc., Calabasas, CA). The detection limit of EC and OC
35 were below 1.0 µg m⁻³. Details of the chemical analysis can be referred to Pathak et al. (2011).

36

37 **Text S2:** Analysis of inorganic elements and water soluble ions

38 Teflon-membrane filter samples were sent to the Institute of Earth Environment, Chinese
39 Academy Sciences (IEECAS, Xi'an, China) in a temperature controlled package (< 4 °C) for
40 elemental analysis by an Energy Dispersive X-Ray Fluorescence analyzer (ED-XRF, Epsilon 5,
41 PANalytical Company, Almelo, The Netherlands) (Watson et al.,1999; Chow and Watson,
42 2012). Twenty inorganic elements (Na, Mg, Al, Si, S, Cl, K, Ca, Ti, V, Cr, Mn, Fe, Co, Ni, Cu,
43 Zn, Sb, Ba and Pb) returned concentrations exceeding method detection limits (MDL) for >50%
44 of the samples, and these elements are included in data analysis. Field blanks were analyzed
45 following the same procedure. MDLs were within the range of 0.5 to 33 ng/m³. The
46 concentrations of ions were determined in the aqueous extracts of the filter samples. One-fourth
47 of each quartz filter were placed into a separate 15 mL vials containing 10 mL distilled–

48 deionized water (18.2 M Ω resistivity). The vials were placed in ultrasonic water bath for 60
49 min and then shaken by mechanical shaker for complete extraction. The extracts were then
50 filtered with a 0.45 μ m pore size microporous membrane, and the filtrates were stored at 4°C
51 in a clean tube until analysis. A Dionex-600 Ion Chromatograph (Dionex Inc., Sunnyvale, CA,
52 USA) was used for the determinations of cations and anions in the aqueous extracts.
53 IonPacCS12A column (20 mM methanesulfonic acid as the eluent) was used for cation analyses,
54 and IonPac AS14A column (8 mM Na₂CO₃/1 mM NaHCO₃ as the eluent) was used for anions
55 analysis. The method detection limits were: 4.6 mg L⁻¹ for Na⁺, 4.0 mg L⁻¹ for NH₄⁺, 0.5 mg
56 L⁻¹ for Cl⁻, 10.0 mg L⁻¹ for K⁺, Mg²⁺ and Ca²⁺ and NO₃⁻, and 20 mg L⁻¹ for SO₄²⁻. The blank
57 values were subtracted from sample concentrations. One sample in each group of ten samples
58 was analyzed twice for quality control.

59

60 **Text S3:** Determination of polycyclic aromatic compounds (PACs)

61 Each quartz filter (with the sampled PM_{2.5}) was cut into smaller pieces and transferred into a
62 33 mL accelerated solvent extractor (ASE) extraction cell and spiked with 40 μ L of a mixture
63 seven deuterated-PAHs (naphthalene-D8, acenaphthene-D10, phenanthrene-D10, pyrene-D10,
64 chrysene-D12, perylene-D12 and benzo[ghi]perylene-D12 each at a concentration of 10
65 μ g/mL), and 40 μ L of mixture of three deuterated compounds [benzophenone-D5 (20 μ g/mL),
66 9,10-anthraquinone-D8 (20 μ g/mL) and carbazole-D8 (5 μ g/mL)]. The seven deuterated PAHs,

67 the two oxygenated compounds (benzophenone-D5 & 9,10-anthraquinone-D8) and
68 carbazoleD8 served as the internal standards for the quantification of the PAHs, OPAHs and
69 AZAs, respectively. Then inert bulk sorbent (Isolute HMN, Biotage, Uppsala, Sweden) were
70 used to fill up the extra space of the extraction cells. Blanks samples (n =2) made of bulk sorbent
71 alone were also transferred into ASE extraction cells and spiked with same amount of
72 deuterated internal standards as with the samples. Each sample was and extracted twice by
73 pressurized liquid extraction (ASE 200, Dionex, Sunnyvale, CA, USA) with the same ASE
74 instrument parameters as previously outlined (Bandowe and Wilcke, 2010; Bandowe et al.,
75 2014; Bandowe et al., 2016). The organic solvent used in the first extraction of each sample
76 was dichloromethane. Each sample was extracted a second time with acetone:dichloromethane
77 mixture (2:1 v/v). Extracts from each sample were combined and transferred into a turbo-vap
78 extraction tube, 10 ml of hexane and 3 drops of toluene was then added. The extracts were then
79 concentrated on a Turbo Vap evaporating system (at 35°C) until a volume < 1 mL. Each sample
80 was then transferred into a 2 ml GC-vial. Floranthene-D10 (25 µL 22 µg/mL) was spiked into
81 some of the extracts in the GC-vials to serve as recovery standards. The target compounds (27
82 PAHs, 18 OPAHs and 4 azaarenes, Table S2) in each extract were measured with an Agilent
83 7890N gas chromatograph coupled to an Agilent 5975C mass selective detector (GC-MS)
84 operating in selected ion monitoring mode. The quantification of the target compounds were
85 performed with an internal standard methods with the deuterated compounds spiked into each

86 sample before extraction. Further information about the analysis method can be referred to in
87 previous articles (Bandowe and Wilcke, 2010; Bandowe et al. 2014a,b; Bandowe et al., 2016;
88 Lui et al., 2016).

89 High-purity solvents (pesticide residue grade) were used for all extractions, rinsings, and
90 preparation of standards. Glassware and metallic parts of ASE cells, and other labware were
91 machine-washed and baked at 250 °C. Prior to usage, glassware was further rinsed with high-
92 purity solvents. Target compounds were determined in blanks which were analyzed with the
93 analytical procedure as the samples. The average amount of the target compounds that were
94 found in the blanks were deducted from that of the same compound found in the sample to
95 correct for laboratory contamination and field contamination. Limit of detection was defined as
96 a mass of target compound three times greater than the baseline noise ($S/N = 3$). The recovery
97 of the deuterated internal standards spiked to the samples before extracted were determined as
98 indicator of the accuracy of the analytical measurement. The recoveries (mean \pm standard
99 deviation) of the deuterated internal standards were: naphthalene-D8 ($72\pm 8\%$), acenaphthene-
100 D10 ($73\pm 8\%$), phenanthrene-D10 ($73\pm 8\%$), pyrene-D10 ($76\pm 7\%$), chrysene-D12 ($78\pm 6\%$),
101 perylene-D12 ($81\pm 12\%$), benzo(ghi)perylene-D12 ($88\pm 14\%$), benzophenone-D5 ($42\pm 4\%$),
102 9,10-anthraquinone-D8 ($55\pm 5\%$) and carbazole-D8 ($46\pm 4\%$).

103 The accuracy and precision of the analytical method for target PAHs, alkyl PAHs, OPAHs and
104 and azaarenes were checked by a spike and recovery experiments ($n = 3$) and reported in

105 previous articles (Bandowe et al., 2014; Bandowe et al., 2016). Average recoveries of targeted
106 PAHs/alkyl PAHs and OPAHs were 102% (range: 67 to 154%) and 96% (64 to 152 %),
107 respectively (Bandowe et al., 2014). The relative standard deviation (RSD) for the PAHs were
108 5% (range: 0.7 to 10%) and 9% (range: 2 to 31%) for PAHs/alkyl-PAHs and OPAHs
109 respectively. (Bandowe et al., 2014). The recoveries (mean \pm standard deviation) of azaarenes
110 in spike and recovery experiment were $75\pm 5\%$, $87\pm 6\%$ and $54\pm 5\%$, for quinoline,
111 benzo[h]quinoline and acridine, respectively (Bandowe et al., 2016). The method we applied in
112 the current study is only a slight modification of the method applied in the previous works. The
113 solvent used for the extraction in the current study was DCM (first extraction) followed by
114 acetone: DCM (2:1 v/v). This extraction solvent was already applied to extract same target
115 PACs from PM_{2.5} on filters (Lui et al., 2016). Since the method applied in the current study is
116 very similar to the previous study (Bandowe et al., 2014; Bandowe et al., 2016), the accuracy
117 and precision can be gauged from the results of the previous spike and recovery study (Bandowe
118 et al., 2014; Bandowe et al., 2016).

119

120 **Text S4:** Determination of carbonyls

121 Each sample (PM_{2.5} on the filters) were transferred into separate in 50 mL Falcon tubes and
122 extracted with 20 mL ultrapure methanol (HPLC grade, Sigma-Aldrich Corporation, USA) on
123 an ultrasonic bath (Branson 5510E-DTH, 40 kHz) containing water at 25 °C for 20 minutes.

124 The extract was transferred to a round-bottom flask and evaporated by rotary evaporator (RV10
125 Basic Rotary Evaporators, IKA Works, VWR, USA) at 30 °C until 5 ml remained. The
126 concentrated extract from each sample was transferred to Eppendorf vials and purged with
127 nitrogen at room temperature until dryness. The dried extract was stored at -20 °C until the
128 analysis of carbonyls. In summary, the dried sample extracts (containing the carbonyl
129 compounds) were re-dissolved in in water to a concentration of 1 mg/L. An excess amount of
130 O-(2,3,4,5,6-pentafluorobenzyl)hydroxylamine hydrochloride (PFBHA) in aqueous solution
131 (e.g., 0.5 ml of 5 mg/ml solution) was added to 5 ml of the extracts. The target carbonyls in the
132 extracts react with the PFBHA to form derivatives. The solution containing the PFBHA-
133 derivatives was then acidified to pH 2 and allowed to stand for 24 hours at room temperature.
134 This solution was then extracted with 2 ml hexane, separated from the aqueous layer and dried
135 with 50 mg anhydrous Na₂SO₄. Finally, 1 µl of analyte dissolved in hexane was transferred into
136 a vial for gas chromatography–mass spectrometry (GC/MS) analysis. The analysed carbonyls
137 are shown in Table S2 (Supplementary Material). Detailed description of the method for the
138 derivatization of carbonyls with PFBHA and the subsequent GC/MS analytical procedure was
139 described previously (Yu et al., 1995). Further experimental details were described previously
140 (Dai et al., 2012).

141

142 **Text S5:** Cell culture and cell viability

143 PM sample were prepared using two-stage sonication of the Teflon filter in methanol and
144 followed by drying under nitrogen air (Totlandsdal et al., 2012). The PM samples were re-
145 suspended in dimethyl sulfoxide (DMSO) (<0.01 vol/vol %) and mixed with Roswell Park
146 Memorial Institute (RPMI) cell culture medium before being used for exposure of cells. Control
147 sample was prepared by a blank filter as the sample preparation described above.

148 A549 cells were obtained from the American Type Culture Collection (ATCC) and cultured
149 using RPMI cell culture medium (Thermo Fisher Scientific Inc., MA, USA) supplemented with
150 10% heat-inactivated fetal bovine serum (Biowest, MO, USA) and 1% antibiotics
151 penicillin/streptomycin (100 U mL⁻¹), in a humidified incubator supplied with 5% carbon
152 dioxide (CO₂) at 37 °C. A549 cells were seeded onto inserts in 24-well transwells at a density
153 of 1×10^5 cells ml⁻¹ and incubated for 24 h. The cell medium was removed and replaced with
154 300 µl of the prepared samples for the next 24 h. Each experiment was conducted in
155 quadruplicate.

156 Cell viability were determined by MTT (3-[4, 5-dimethylthiazol-2-yl]-2, 5 diphenyl tetrazolium
157 bromide) assay. After PM_{2.5} exposure for 24 hrs, the supernatant was collected. Then, 100 ml
158 of 10% MTT solution (Sigma Aldrich, St. Louis, MO, USA) were added at 37 °C for 4 h for
159 color development. Optical density was measured at 540 nm by absorbance microplate reader
160 (ELx800, BioTek, VT, USA). The results were presented as a percentage of the absorbance of
161 control.

162

163 **Text S6:** Reactive Oxygen Species (ROS) and TNF- α analysis

164 Cellular ROS was determined by the fluorogenic cell-based method using 2',7'-
165 dichlorodihydrofluorescein diacetate (DCFH-DA) as the indicator, which has been commonly
166 used for environmental toxicology (Eruslanov and Kusmartsev, 2010; Montesinos et al., 2015).

167 After 24 hours exposure to PM_{2.5}, DCFH-DA was added to the A549 cells, and cultured for 30
168 min. Each well was washed with PBS to remove the DCFH-DA that did not combined with
169 cells. The fluorescence intensity (IF) was determined by a Light Luminescence Plate Reader
170 (VICTOR™ X; PerkinElmer, Waltham, USA) at an excitation wavelength of 485 nm and an
171 emission wavelength of 530 nm.

172 Enzyme-linked immunosorbent assay (ELISA) (R&D Systems, Inc., MN, USA) was used to
173 determine tumor necrosis factor- α (TNF- α) levels in supernatant according to the
174 manufacturer's instructions (Chuang et al., 2018).

175 For PSA, gels (0.6% Agarose; Biorline, UK) were prepared using Tris/Borate/EDTA (TBE)
176 buffer solution (Thermo Scientific, UK) diluted 10 times with agarose and the solution was
177 heated by microwave (EMS-820; Electron Microscopy Services, USA) to clarity and
178 transparency. The solidified gel was placed in an electrophoresis cell (DYCP-34A type;
179 NANBEI, China) containing 10 times diluted TBE buffer.

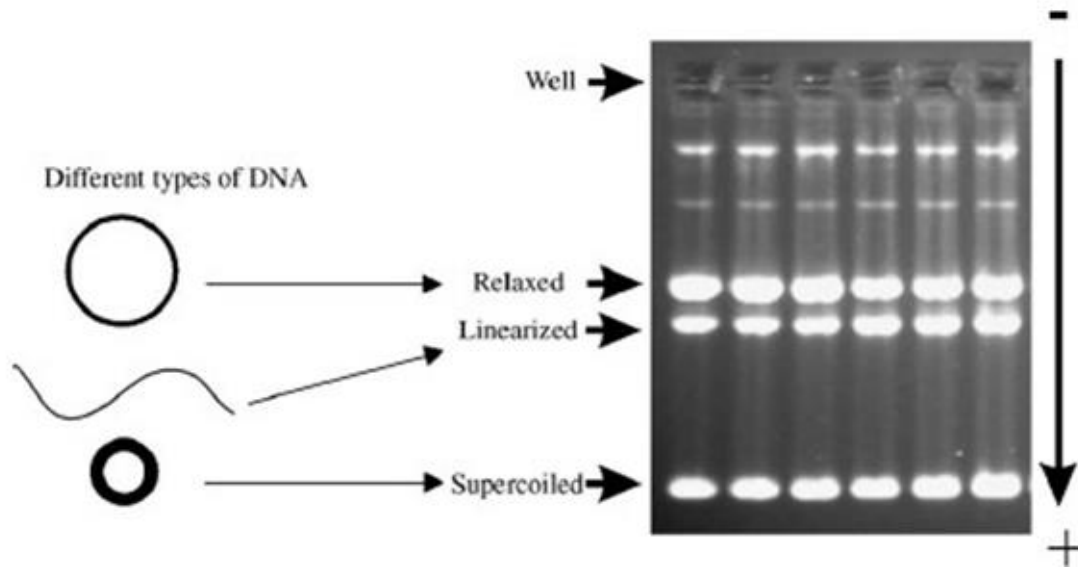
180 Bromophenol blue stain (14 μ L; Sigma-Aldrich, UK) was added to the DNA-PM_{2.5} samples

181 and placed on a rocking platform (Bio-Rad, UK) for 4 hours. Post-mixing, 20 μ L of the DNA-
182 PM_{2.5} mixtures were aliquoted into each gel well. Three parallel samples were made for each
183 sample. Ethidium bromide (EB; 20 μ L; Sigma-Aldrich, China) was added to both sides of the
184 electrophoresis tank (NANBEI, China). After the EB was fully dissolved in the buffer, the
185 laboratory electrophoresis power supply (DYY-6C; NANBEI, China) was turned on and
186 operated at 30 Volts for 16 hours.

187 Post-electrophoresis, the optical densities of three different DNA morphologies (i.e. super-
188 coiled, relaxed and linear) in the gel were captured using a gel documentation system
189 (ChemiDoc, Bio-Rad, UK) and the GeneTools (Version 4.3.10; Syngene, USA) image analysis
190 software program was utilized to calculate the toxic dose of PM_{2.5} causing DNA damage (%).

191 Additional information about the PSA procedure can be found in Chuang et al., (2011).
192

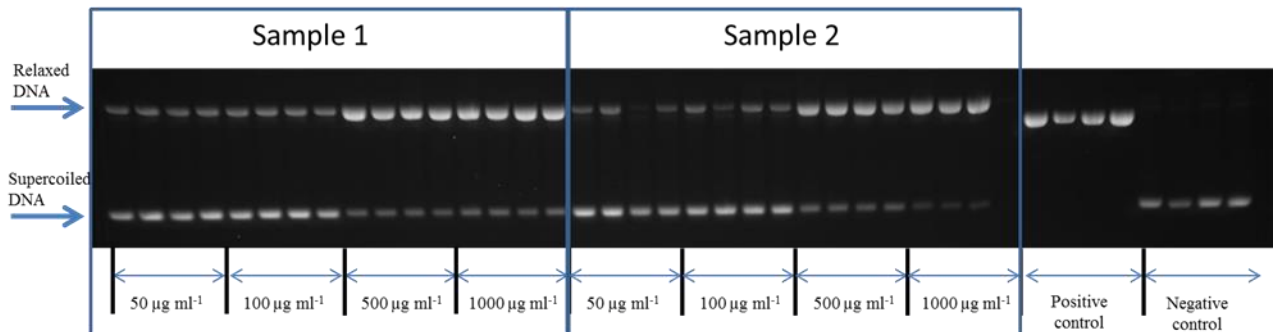
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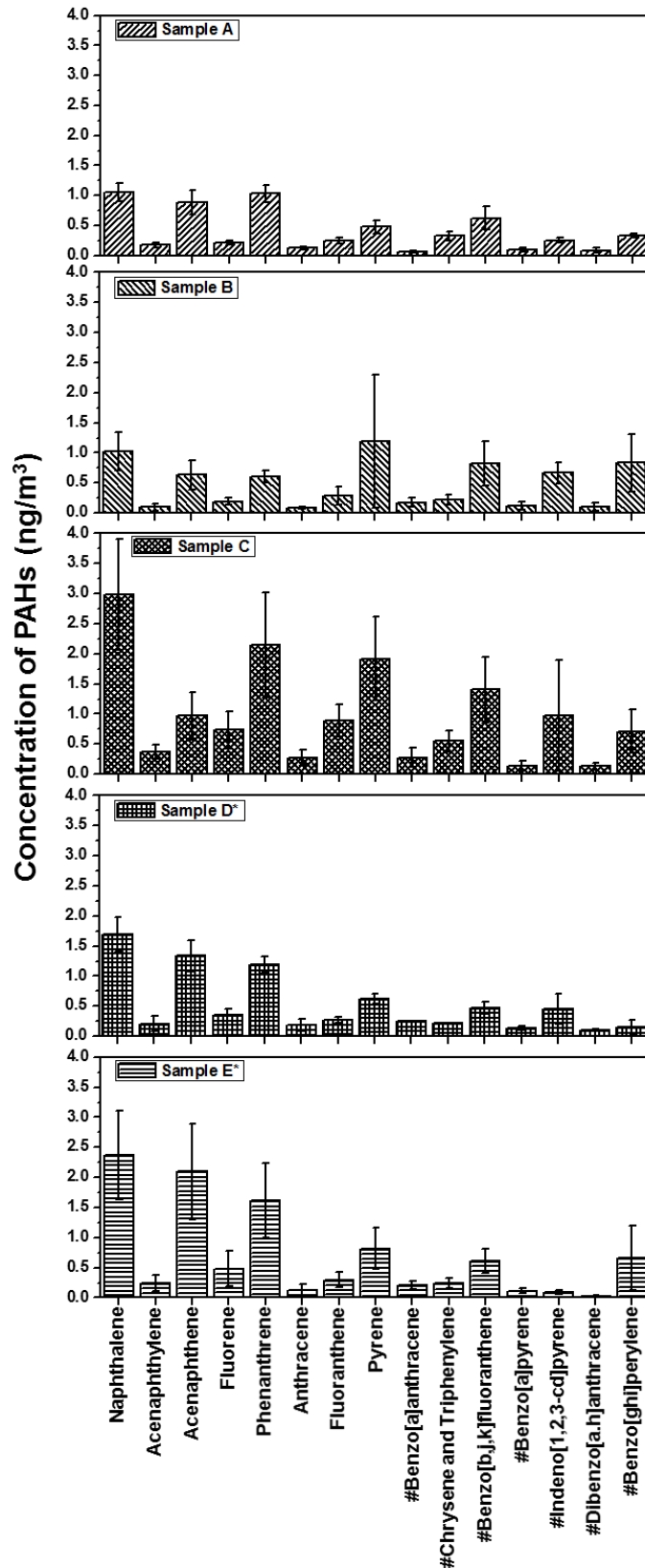
195 **Figure S1:** Identification of different types of plasmid Φ X174-RF DNA (Promega, London,
196 UK) in gel electrophoresis.

197



198

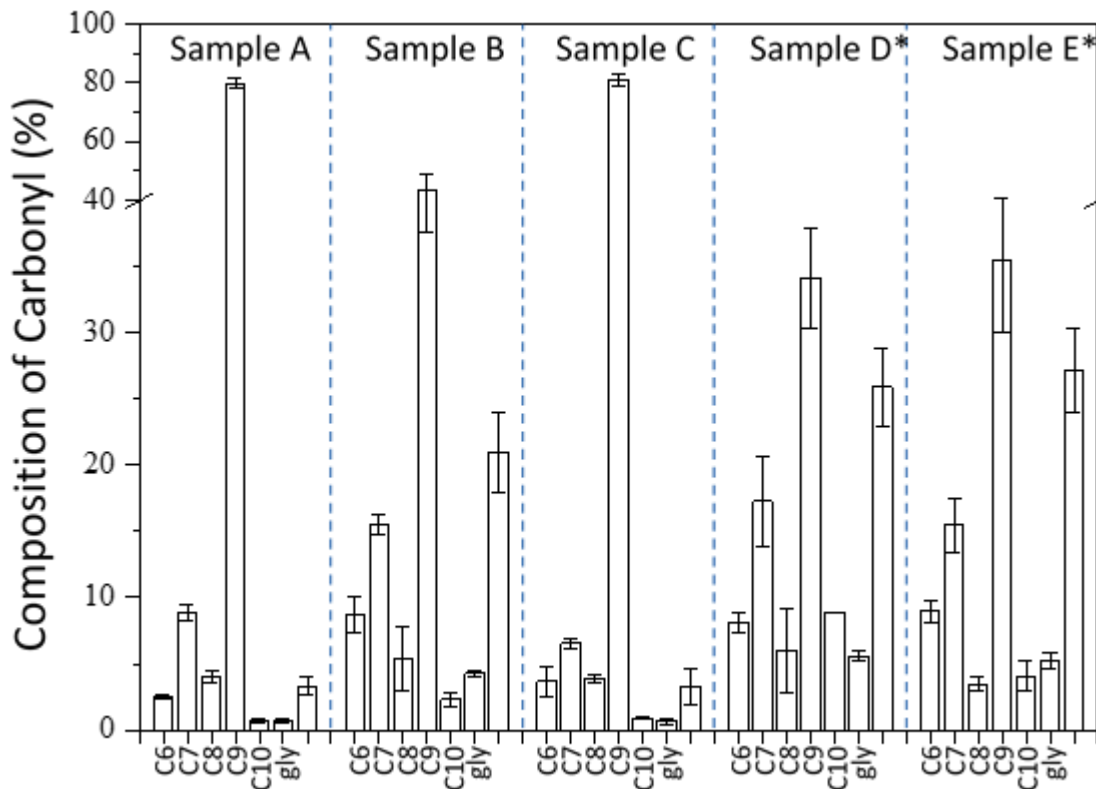
199 **Figure S2:** Gel images demonstrate oxidative damage to supercoiled DNA induced by PM_{2.5}
200 sample.



201

202 **Figure S3:** Descriptive analysis and relative abundances of U.S. EPA Priority PAHs. Hash (#)
 203 indicated by U.S. EPA as probable human carcinogen.

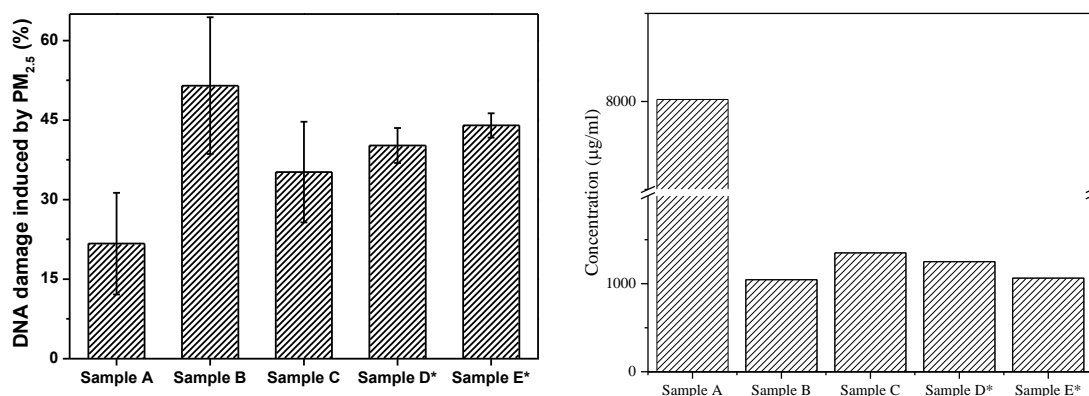
204



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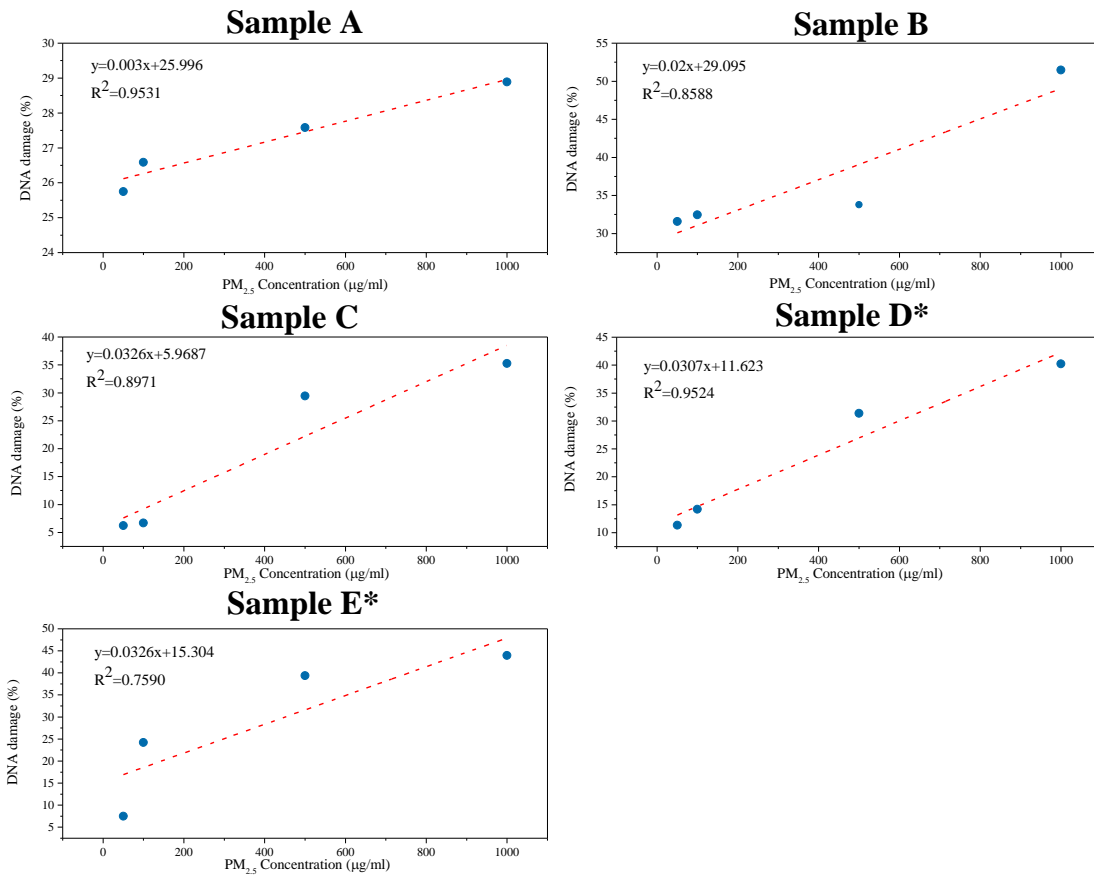
206 **Figure S4:** Mean contributions of individual carbonyl compounds to the total concentration of
 207 carbonyls in samples from different sampling locations. The y-axis was broken at 40% to
 208 enlarge the scale before the break. Error bars indicate standard deviations for each sample.

209



210

211 **Figure S5:** The average DNA damage induced by extracts of PM_{2.5} (1000 µg/ml) and LD₅₀ of
 212 PM_{2.5} collected from five sampling locations (n = 4 for Sample A, B, C and E* and n = 3 for
 213 sample D*).



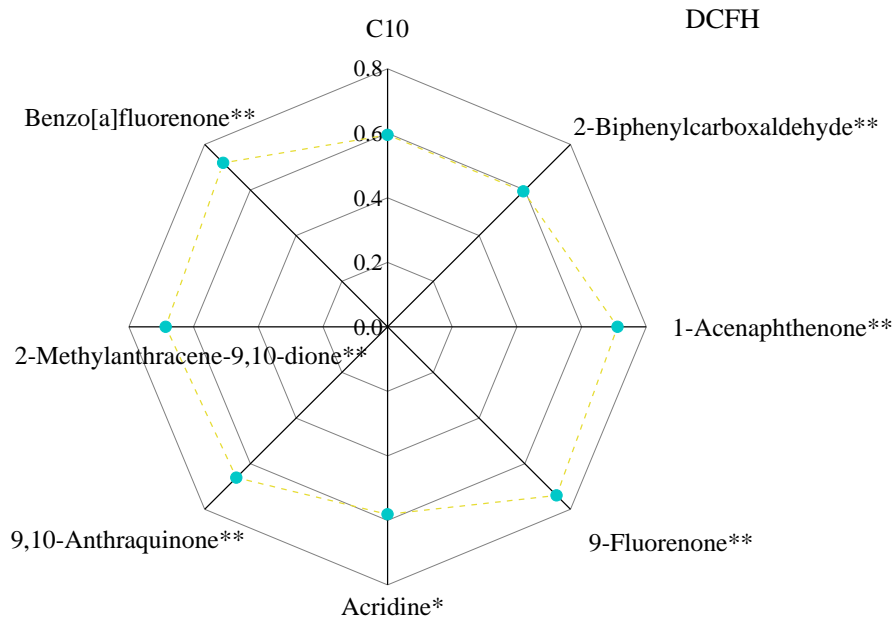
215

216 **Figure S6:** Dose response analysis between DNA damage and PM_{2.5} concentration using
 217 plasmid scission assay (PSA).

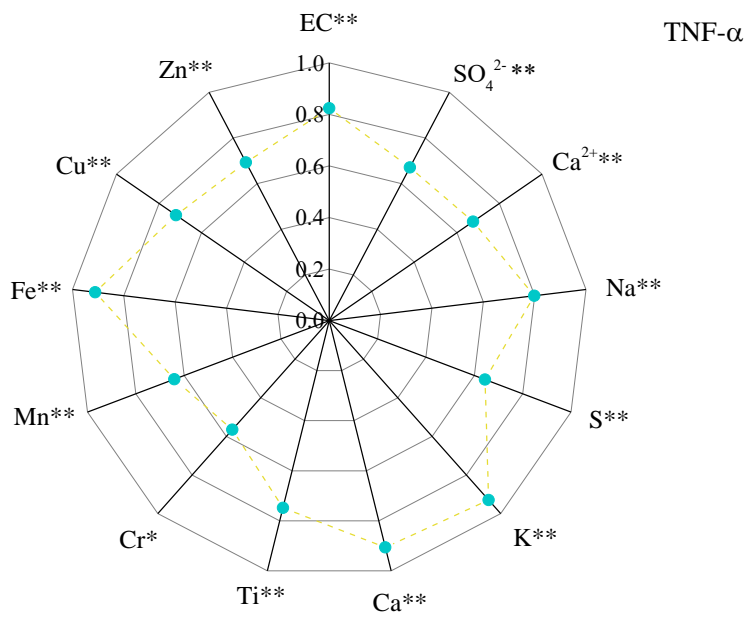
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220



221



222

223 **Figure S7:** Correlations between the concentrations of chemical species extracted from PM_{2.5}
 224 and biological effects [ROS generation (DCFH) and TNF-α].
 225

226

227 **Table S1:** List of chemical species analyzed in samples their abbreviations

Group				
Carbonaceous Species	Organic carbon (OC)	Elemental carbon (EC)		
Ions	Sodium ion (Na ⁺)	Calcium ion (Ca ²⁺)	Ammonium (NH ₄ ⁺)	Chloride (Cl ⁻)
&	Nitrate (NO ₃ ⁻)	Sulfate (SO ₄ ²⁻)		
Inorganic Elements	Sodium (Na)	Potassium (K)	Calcium (Ca)	Chlorine (Cl)
	Sulfur (S)	Chromium (Cr)	Titanium (Ti)	Iron (Fe)
	Zinc (Zn)	Nickel (Ni)	Lead (Pb)	Magnesium (Mg)
	Vanadium (V)	Stibium (Sb)	Cobalt (Co)	Rubidium (Rb)
	Aluminum (Al)	Silicon (Si)	Manganese (Mn)	Copper (Cu)
	Barium (Ba)	Calcium (Ca)		
	1,2,3,4-Tetrahydronaphthalene (THNAPH)	Naphthalene (NAPH)	2-Methylnaphthalene (2-MNAPH)	1-Methylnaphthalene (1-MNAPH)
	Biphenyl (BP)	1,3-Dimethylnaphthalene (1,3-DMNAPH)	Acenaphthylene (ACENY)	Acenaphthene (ACEN)
Polycyclic Aromatic Hydrocarbons (PAHs) and Alkyl-PAHs	Fluorene (FLUO)	Phenanthrene (PHEN)	Anthracene (ANTH)	4H-Cyclopenta(d,e,f)phenanthrene (CPHEN)
	1-Methylphenanthrene (1-MPHEN)	3,6-Dimethylphenanthrene (3,6-DMPHEN)	Fluoranthene (FLUA)	Pyrene (PYR)
	Retene (RET)	Benz[a]anthracene (B(A)A)	Chrysene and Triphenylene (CHRY)	Benzo[b+j+k]fluoranthene (B(BJK))
	Benzo[e]pyrene (B(E)P)	Benzo[a]pyrene (B(A)P)	Perylene (PERY)	Indeno[1,2,3-cd]pyrene (IND)
	Dibenz[a,h]anthracene (DIBE)	Benzo[ghi]perylene (B(GHI))	Coronene (COR)	
Oxygenated-Polycyclic Aromatic Hydrocarbons (OPAHs)	1-Indanone (1-INDA)	1,4-Naphthoquinone (1,4-NQ)	1-Naphthaldehyde (1-NLD)	2-Biphenylcarboxaldehyde (2-BIP)
	1-Acenaphthenone (1-ACENONE)	9-Fluorenone (9-FLO)	9,10-Anthraquinone (9,10-ANQ)	1,8-Naphthalic anhydride (1,8-ANA)
	1,4-Anthraquinone (1,4-ANQ)	4H-Cyclopenta[d,e,f]phenanthrene-4-one (CPHENONE)	2-Methylanthracene-9,10-dione (2-METH)	Benzo[a]fluorenone (B(A)FLUONE)
	7H-Benz[de]anthracene-7-one (BANTONE)	Benz[a]anthracene-7,12-dione (7,12-B(A)A)	1,4-Chrysenequinone (1,4-CHQ)	Naphthacene-5,12-dione (NAP-5,12)
	6H-benzo[c,d]pyrene-6-one (BPYRONE)			
Azaarenes	Quinoline (QUI)	Benzo[h]quinolone (BQI)	Acridine (ACR)	Carbazole (CBZ)
Carbonyls	Hexaldehyde (C6)	Heptaldehyde (C7)	Octaldehyde (C8)	Nonaldehyde (C9)
	Decaldehyde (C10)	Glyoxal (gly)	Methylglyoxal (mgly)	

228

229 **Table S2:** DNA damage induced by the PM_{2.5} emitted from five sampling locations

	Concentration (µg/ml)	Average DNA damage (%)	Concentration (µg/ml)	Average DNA damage (%)	Concentration (µg/ml)	Average DNA damage (%)	Concentration (µg/ml)	Average DNA damage (%)
Sample A (sub-sample number 1-4)	1		2		3		4	
	50	25.3	50	21.3	50	20.1	50	36.3
	100	35.7	100	35.7	100	16.9	100	18.1
	500	24.0	500	26.8	500	24.6	500	34.9
	1000	15.5	1000	37.8	1000	13.7	1000	19.9
Sample B (sub-sample number 5-8)	5		6		7		8	
	50	29.5	50	36.0	50	25.7	50	35.1
	100	30.3	100	29.9	100	37.4	100	32.2
	500	25.9	500	43.8	500	30.7	500	34.8
	1000	33.4	1000	61.1	1000	65.8	1000	45.6
Sample C (sub-sample number 9-12)	9		10		11		12	
	50	7.1	50	11.4	50	2.3	50	4.0
	100	10.7	100	6.9	100	4.9	100	4.2
	500	44.9	500	33.6	500	18.0	500	21.3
	1000	40.6	1000	48.2	1000	26.8	1000	25.5
Sample D** (sub-sample number 13-15)	13		14		15			
	50	6.4	50	7.7	50	19.8		
	100	7.8	100	16.2	100	18.6		
	500	29.8	500	30.9	500	33.5		
	1000	40.6	1000	44.1	1000	36.0		
Sample E* (sub-sample number 16-19)	16		17		18		19	
	50	4.4	50	7.7	50	4.7	50	13.2
	100	21.3	100	19.7	100	28.7	100	27.1
	500	36.6	500	35.7	500	41.9	500	43.6
	1000	44.3	1000	46.8	1000	40.4	1000	44.4

230 ** represents samples collected from stainless steel environmental chamber that mimic residential kitchen hood condition.

231 **Table S3:** Selected mean concentration ratios (\pm standard deviations) of PAHs, OPAHs, AZAs in different samples.

	Sample A	Sample B	Sample C	Sample D*	Sample E*
2-MNAPH/NAPH	0.20 \pm 0.05	0.19 \pm 0.01	0.18 \pm 0.02	0.20 \pm 0.01	0.17 \pm 0.02
1-MNAPH/NAPH	0.25 \pm 0.07	0.22 \pm 0.02	0.20 \pm 0.01	0.25 \pm 0.03	0.38 \pm 0.08
1,3-DMNAPH/NAPH	0.50 \pm 0.10	0.28 \pm 0.04	0.30 \pm 0.04	0.29 \pm 0.03	0.24 \pm 0.02
Σ Alkyl-NAPH/NAPH	0.45 \pm 0.12	0.41 \pm 0.03	0.38 \pm 0.02	0.45 \pm 0.04	0.55 \pm 0.07
2-MPHEN/PHEN	0.18 \pm 0.05	0.16 \pm 0.03	0.22 \pm 0.02	0.16 \pm 0.02	0.20 \pm 0.03
3,6-DMPHEN/PHEN	0.09 \pm 0.02	0.10 \pm 0.03	0.12 \pm 0.02	0.12 \pm 0.01	0.16 \pm 0.01
Σ Alkyl-PHEN/PHEN	0.27 \pm 0.07	0.27 \pm 0.06	0.34 \pm 0.03	0.28 \pm 0.02	0.36 \pm 0.03
PHEN/(PHEN+ANTH)	0.89 \pm 0.03	0.88 \pm 0.01	0.89 \pm 0.01	0.87 \pm 0.05	0.93 \pm 0.02
B(A)A/(B(A)A+CHR)	0.17 \pm 0.03	0.45 \pm 0.02	0.31 \pm 0.08	0.54 \pm 0.02	0.47 \pm 0.02
FLUA/(FLUA+PYR)	0.35 \pm 0.02	0.24 \pm 0.10	0.32 \pm 0.04	0.31 \pm 0.01	0.27 \pm 0.02
IND/(IND+B(GHI))	0.43 \pm 0.03	0.48 \pm 0.16	0.50 \pm 0.26	0.73 \pm 0.28	0.14 \pm 0.05
Σ LMW-PAHs/ Σ HMW-PAHs	1.41 \pm 0.16	0.65 \pm 0.18	1.08 \pm 0.18	1.86 \pm 0.25	2.30 \pm 0.14
Σ 6alkPAHs/ Σ 21PAHs	0.29 \pm 0.04	0.17 \pm 0.03	0.25 \pm 0.03	0.35 \pm 0.02	0.44 \pm 0.05
Σ 6AlkylPAHs/ Σ 29PAHs	0.20 \pm 0.02	0.13 \pm 0.02	0.17 \pm 0.01	0.23 \pm 0.01	0.27 \pm 0.03
Σ 17OPAHs/ Σ 29PAHs	0.06 \pm 0.02	0.05 \pm 0.01	0.15 \pm 0.05	0.10 \pm 0.01	0.16 \pm 0.04
Σ 17OPAHs/ Σ 21PAHs	0.09 \pm 0.03	0.06 \pm 0.02	0.22 \pm 0.09	0.15 \pm 0.01	0.27 \pm 0.07
B(E)P/B(A)P	2.60 \pm 1.11	3.03 \pm 0.24	5.76 \pm 2.83	1.02 \pm 0.10	1.56 \pm 0.24
B(GHI)/B(A)P	3.76 \pm 0.85	6.63 \pm 2.54	7.38 \pm 5.99	0.99 \pm 0.66	5.18 \pm 2.74
1,4-NQ/NAPH	0.02 \pm 0.00	0.03 \pm 0.01	0.09 \pm 0.08	0.07 \pm 0.02	0.02 \pm 0.01
1-NLD/1-MNAPH	0.08 \pm 0.03	0.08 \pm 0.03	0.14 \pm 0.08	0.12 \pm 0.04	0.09 \pm 0.04
9-FLO/FLUO	0.25 \pm 0.05	0.19 \pm 0.08	0.67 \pm 0.23	0.34 \pm 0.10	0.81 \pm 0.28
9,10-ANQ/ANTH	0.34 \pm 0.14	0.45 \pm 0.19	1.51 \pm 0.18	0.58 \pm 0.14	1.88 \pm 0.64
7,12-B(A)A/B(A)A	0.23 \pm 0.05	0.06 \pm 0.02	0.43 \pm 0.09	0.05 \pm 0.01	0.12 \pm 0.03
1-ACEQ/ACEN	0.03 \pm 0.00	0.02 \pm 0.00	0.16 \pm 0.05	0.03 \pm 0.00	0.04 \pm 0.02
1-ACEQ/ACENY	0.14 \pm 0.06	0.16 \pm 0.05	0.43 \pm 0.17	0.22 \pm 0.10	0.35 \pm 0.07
1,4-CHRQ/CHR	0.49 \pm 0.46	0.36 \pm 0.33	0.50 \pm 0.27	0.55 \pm 0.17	0.52 \pm 0.47
QUI/NAPH	0.01 \pm 0.00	0.01 \pm 0.00	0.03 \pm 0.02	0.03 \pm 0.00	0.03 \pm 0.02
BQI/PHEN	0.01 \pm 0.00	0.01 \pm 0.00	0.02 \pm 0.01	0.03 \pm 0.02	0.02 \pm 0.01
ACR/ANTH	0.03 \pm 0.01	0.05 \pm 0.03	0.12 \pm 0.05	0.10 \pm 0.03	0.26 \pm 0.08
CBZ/FLUO	0.12 \pm 0.07	0.03 \pm 0.02	0.10 \pm 0.05	0.10 \pm 0.06	0.08 \pm 0.05

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233 **Table S4:** The average concentrations \pm standard deviation (ng/m³) of carbonyls in five sampling locations

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Compound ^a	Sample A	Sample B	Sample C	Sample D*	Sample E*
Hexaldehyde	23.8 \pm 5.5	13.4 \pm 2.6	199.8 \pm 88.3	21.6 \pm 7.1	37.4 \pm 19.1
Heptaldehyde	83.2 \pm 15.1	24.9 \pm 8.5	355.7 \pm 143.7	44.0 \pm 3.7	62.2 \pm 25.0
Octaldehyde	38.2 \pm 8.1	8.3 \pm 4.7	203.4 \pm 64.8	16.9 \pm 12.9	13.7 \pm 4.6
Nonaldehyde	757.1 \pm 168.2	69.8 \pm 25.0	4333.9 \pm 1681.7	88.3 \pm 14.1	153.2 \pm 98.7
Decaldehyde	6.6 \pm 2.4	3.3 \pm 1.5	48.1 \pm 16.9	29.5 \pm 2.1	16.8 \pm 9.7
Glyoxal	6.3 \pm 1.0	6.7 \pm 2.1	32.9 \pm 8.6	14.4 \pm 2.5	21.1 \pm 7.7
Methylglyoxal	30.8 \pm 5.3	32.8 \pm 10.5	153.8 \pm 26.6	66.6 \pm 7.2	107.6 \pm 39.0
Σ Carbonyls	946.0 \pm 170.4	158.4 \pm 42.2	5327.6 \pm 1709.6	261.7 \pm 49.3	412.1 \pm 164.2

235 ^aName of the individual compound can be referred to Table S1 (Supplementary Material).

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242 **Table S5:** Correlation coefficients between the concentrations of PACs and carbon fractions, ion, metal and carbonyl in all samples.

	$\Sigma 30\text{PAHs}$	$\Sigma 16\text{PAHs}$	$\Sigma 21\text{PAHs}$	$\Sigma \text{LMW-PAHs}$	$\Sigma \text{HMW-PAHs}$	$\Sigma \text{Alkyl-NAP}$	$\Sigma \text{Alkyl-PHEN}$	$\Sigma 6\text{alkPAHs}$	$\Sigma 4\text{AZAs}$	$\Sigma 17\text{OPAHs}$
PM _{2.5}	0.92**	0.89**	0.85**	0.95**	0.60**	0.91**	0.93**	0.93**	0.92**	0.94**
TC	0.90**	0.87**	0.83**	0.95**	0.56*	0.93**	0.92**	0.95**	0.92**	0.95**
OC	0.90**	0.87**	0.82**	0.95**	0.55*	0.93**	0.91**	0.95**	0.92**	0.95**
EC	0.55*	0.55*	0.60**	0.34	0.70**	0.26	0.44	0.26	0.31	0.38
Cl ⁻	0.80**	0.77**	0.75**	0.77**	0.57*	0.82**	0.77**	0.76**	0.70**	0.83**
NO ₃ ⁻	0.32	0.33	0.39	0.10	0.57*	-0.003	0.20	-0.003	0.06	0.13
SO ₄ ²⁻	0.14	0.16	0.21	-0.04	0.41	-0.17	0.02	-0.15	0.004	-0.02
Na ⁺	0.55*	0.57*	0.60*	0.38	0.64**	0.26	0.35	0.28	0.44	0.38
NH ₄ ⁺	0.42	0.42	0.46	0.27	0.57*	0.22	0.37	0.19	0.19	0.30
Ca ²⁺	0.13	0.12	0.16	0.02	0.27	-0.04	0.13	0.03	-0.03	0.03
Al	0.70**	0.70**	0.71**	0.64*	0.60*	0.54*	0.61*	0.61*	0.62*	0.59*
Si	0.71**	0.66**	0.64**	0.77**	0.37	0.79**	0.70**	0.79**	0.69**	0.76**
Cl	0.79**	0.76**	0.73**	0.79**	0.51*	0.87**	0.78**	0.81**	0.73**	0.84**
Cr	0.84**	0.83**	0.81**	0.79**	0.65**	0.73**	0.73**	0.75**	0.80**	0.82**
Ni	0.63**	0.61**	0.61**	0.56*	0.52*	0.50*	0.52*	0.48	0.64**	0.60*
Sb	0.82**	0.81**	0.78**	0.80**	0.57*	0.83**	0.69**	0.79**	0.78**	0.82**
Pb	0.63**	0.58*	0.56*	0.71**	0.29	0.66**	0.64**	0.67**	0.61**	0.69**
$\Sigma \text{Carbonyls}$	0.67**	0.63**	0.64**	0.62**	0.55*	0.49*	0.71**	0.55*	0.55*	0.62**

243 ** indicate significant correlations at the p = 0.01 level (2-tailed).

244 * indicate significant correlations at the p = 0.05 level (2-tailed).

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